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1	Effects of flow regime on benthic algae and macroinvertebrates - a comparison between
2	regulated and unregulated rivers
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15	Effects of river flow on benthic algae and macroinvertebrates
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23 Abstract

Natural fluctuations in flow are important for maintaining the ecological integrity of riverine 24 ecosystems. However, the flow regime of many rivers has been modified. We assessed the 25 26 impact of water chemistry, habitat and streamflow characteristics on macroinvertebrates and benthic algae, comparing 20 regulated with 20 unregulated sites. Flow regime, calculated 27 from daily averaged discharge over the five years preceding sampling, was generally more 28 stable at regulated sites, with higher relative discharges in winter, lower relative discharges in 29 spring and smaller differences between upper and lower percentiles. However, no consistent 30 differences in benthic algal or macroinvertebrate structural and functional traits occurred 31 32 between regulated and unregulated sites. When regulated and unregulated sites were pooled, overall flow regime, calculated as principal components of discharge characteristics over the 33 34 five years preceding sampling, affected macroinvertebrate species assemblages, but not 35 indices used for ecosystem status assessment or functional feeding groups. This indicates that, while species identity shifted with changing flow regime, the exchanged taxa had similar 36 feeding habits. In contrast to macroinvertebrates, overall flow regime did not affect benthic 37 algae. Our results indicate that overall flow regime affected the species pool of 38 macroinvertebrates from which recolonization after extreme events may occur, but not of 39 benthic algae. When individual components of flow regime were analyzed separately, high 40 June (i.e. three months before sampling) flow maxima were associated with low benthic algal 41 42 taxon richness, presumably due to scouring. Macroinvertebrate taxon richness decreased with lower relative minimum discharges, presumably due to temporary drying of parts of the 43 riverbed. However, recolonization after such extreme events presumably is fast. Generally, 44 45 macroinvertebrate and benthic algal assemblages were more closely related to water physicochemical than to hydrological variables. Our results suggest that macroinvertebrate and 46 47 benthic algal indices commonly used for ecological status assessment are applicable also in regulated rivers. 48

50 **1. Introduction**

Environmental gradients shape river ecosystems along with disturbances such as floods and 51 droughts, and the flow regime is often regarded to be a key driver of river ecosystems (Poff et 52 53 al., 1997; Bunn and Arthington, 2002). Substantial variability exists in natural river flow characteristics, which are related to climate, geology and topography, and natural fluctuations 54 in river flow are fundamentally important for the long-term sustainability and productivity of 55 riverine ecosystems, i.e. for the maintenance of their ecological integrity (Poff et al. 1997; 56 Naiman et al., 2008). However, the flow regime of many rivers has been modified, e.g. by 57 dampening or eliminating natural floods and droughts in order to meet human needs such as 58 59 transport, water supply, flood control or hydropower (Dynesius and Nilsson, 1994; Gleick, 2003). This may negatively affect river ecosystems, and indeed hydraulic engineering is, next 60 61 to pollution from agriculture, regarded as the main factor inhibiting the achievement of good 62 ecological status of European river basins (Menendez et al., 2006). Hydropower is an important global source of electricity (Gracey and Verones, 2016). In 63 Norway, almost all electricity is generated from hydropower plants (Linnerud and Holden, 64 2015), causing about 70% of river catchments to be affected by regulation (www.nve.no). 65 Apart from mandatory minimum flow releases, release of water from hydropower reservoirs 66 67 depends on short- and long-term electricity demand, such that river flow may undergo fluctuations that differ from the natural flow regime (Kern et al., 2012). 68 The flow regime of rivers and streams can be identified by several streamflow characteristics 69 which are deemed ecologically important; seasonal flow pattern, timing and magnitude of 70 extreme flows, frequency and duration of flow extremes and rate of change (Olden and Poff, 71 2003). Alterations to these streamflow characteristics may affect the structure and function of 72

rivers and contribute to the loss of biodiversity (Bunn and Arthington 2002). The

consequences of natural variation and anthropogenic modifications in flow to riverine

recosystems have been relatively well studied (Rolls et al., 2012). For example, streamflow

variability affects fish assemblages and traits (Poff and Allan, 1995; Murchie et al. 2008).

⁷⁷ Likewise, macroinvertebrate assemblages and traits are affected by droughts (Monk et al.,

⁷⁸ 2008; Bonada et al. 2007), but also by summer flow characteristics and by short-term

⁷⁹ hydrological events (Extence et al., 1999). Mass developments of submerged macrophytes in

80 regulated rivers have been related to enhanced winter discharges (which cause less freezing

damage; Johansen et al., 2000). However, conflicting results have also been reported. For

82 benthic algal assemblages, increases as well as decreases in biomass after large floods have

been observed (Power et al., 2008; Schneider, 2015), macrophyte mass developments occur in 83 some but not other rivers having enhanced winter discharges (Johansen et al., 2000), and wide 84 variation is displayed in the severity and direction of responses of fishes to river regulation 85 (Murchie et al., 2008). The varying response of biota after extreme events may partly be 86 explained by recolonization. For example, even if short term spates can decrease the 87 abundance and diversity of macroinvertebrates (Scrimgeour et al., 1988), recovery is often 88 rapid, presumably due to colonization from flow refuges, or from aerial ovipositing adults 89 90 (Müller, 1982; Palmer et al. 1992). Also adaptations, for example in life history, behavior, or morphology (Lytle, 2002; Lytle and Poff, 2004), may contribute to explaining varying 91 responses of the biota after extreme events. In addition, covariation of flow regime with other, 92 potentially influential parameters such as water chemistry may lead to unexplained variation 93 in the biological response. 94

95 However, even though we like to think that the consequences of natural variation and anthropogenic modifications in flow are relatively well understood, present knowledge on the 96 effects of river flow on aquatic biota is to a large degree based on studies covering a relatively 97 short time-scale (Monk et al., 2008). Such studies predict site-specific short-term effects of 98 river flow, but do not allow inferences to which degree the species pool from which 99 100 recolonization may occur is affected. However, this is important in order to distinguish between short-term effects of disturbances which soon may be ameliorated because 101 102 recolonization is fast, and long-lasting consequences for the ecosystem. Comparative studies on the long-term effects of flow regime on aquatic biota are, however, usually based on 103 spatially diverse datasets. This may lead to covariation between flow regime and other 104 105 potentially influential parameters, e.g. climate and hydrochemistry. Such potentially confounding factors have often been ignored, presumably due to a lack of data (Clausen and 106 107 Biggs, 1997; Petrin et al., 2013). Studies that included river flow as well as water chemistry concluded that both direct changes in river flow or indirect changes in water quality may be 108 109 important for river biota (Sheldon and Thoms, 2006; Greenwood et al., 2016). River regulation does not only modify flow regime, but may also affect water quality due to factors 110 such as the transfer of water between river catchments, or the discharge of hypolimnic 111 reservoir water into rivers (Gracey and Verones, 2016). Consequently, river regulation may 112 113 affect biota via changes in flow regime, or via changes in water quality. For planning effective 114 remediation measures, it is important to distinguish effects of flow regime from effects of water quality on river biota. 115

116 Deterioration and improvement of river ecological status in Europe is determined by comparing the biota that occur at a site with those that occur at unimpacted reference sites 117 (EC, 2000). However, river biota respond to many parameters, including hydrochemistry and 118 different aspects of flow regime. This is particularly relevant in so-called Heavily Modified 119 Water Bodies (HMWB). River reaches can be designated as HMWB if applying the 120 hydromorphological measures to reach good ecological status would significantly affect water 121 uses (e.g. flood protection, hydropower generation). The environmental objectives for 122 123 HMWB can be lowered to good ecological potential (GEP) which corresponds to the state that results from applying all hydromorphological measures that may improve ecological 124 status but at the same time do not significantly affect water uses (Kail and Wolter, 2013). This 125 means that, if river regulation for hydropower generation should consistently affect river 126 biota, the environmental objectives for such rivers could be lowered. We therefore wanted to 127 know (i) whether there occur systematic differences in assemblages of macroinvertebrates and 128 benthic algae, i.e. organisms commonly used for ecological status evaluation, between 129 130 regulated and unregulated rivers, and (ii) how flow regime affects macroinvertebrates and benthic algae. 131

We assessed the impact of streamflow characteristics (calculated from five years of daily 132 averaged discharge data), water chemistry and habitat characteristics on macroinvertebrate 133 and benthic algal structural and functional traits, comparing 20 regulated sites (= modified 134 flow regime) with 20 unregulated sites (= natural flow regime). It has been shown before that 135 disturbance regime affects taxon richness (Townsend et al., 1997) and changes competitive 136 interactions among species and age classes (Feminella and Resh, 1990). We therefore 137 138 hypothesized that (1) regulated sites would have a more stable flow regime than unregulated sites, leading to fewer macroinvertebrate and benthic algal taxa in regulated than in 139 140 unregulated sites, and (2) flow regime would shape macroinvertebrate and benthic algal assemblages, with communities adapted to low flow conditions occurring at sites with a stable 141 142 flow regime.

143

144 **2. Material and Methods**

145 **2.1 Sampling sites**

146 The Norwegian Water Resources and Energy Directorate (NVE) operates a network of

147 hydrological gauging stations (Petterson, 2004). From these sites, we selected 20 which were

situated in regulated rivers in South Norway (Fig. 1). Criteria for site selection were i) 148 availability of daily averaged discharge data since 2008, ii) independence of sites, i.e. no site 149 was located downstream from another regulated site, and iii) accessibility for sampling. All 150 20 sites have been regulated for ≥ 25 years (Table A.1 in the appendix), i.e. we expected 151 riverine biota to have adjusted to the modified flow regime. We then selected 20 unregulated 152 sites, based on the same criteria as the regulated sites, and attempted to match the geographic 153 154 spread of the regulated sites as closely as possible (because climate varies in South Norway, with generally wetter and warmer conditions in the South-West (Moreno and Hasenauer, 155 2016). However, some compromises had to be made, such that two of the unregulated sites 156 lay in the same river (but with a large lake in between, such that these two sites had quite 157 different flow regimes). River regulation is a multifaceted term, and also the 20 regulated sites 158 in our dataset were subject to different main effects of regulation. Our dataset includes so-159 called "minimum discharge" sites, i.e. sites from which stream water is abstracted and 160 bypasses the river, so that the amount of water remaining in the stream is reduced; in addition, 161 out dataset includes sites situated downstream the outlet of hydropower plants and sites that 162 were situated downstream dams. In an earlier version of our manuscript, "minimum 163 discharge" and "downstream outlet hydropower plant" sites were analyzed separately. 164 However, since this did not provide additional important information, the regulated sites were 165 pooled. 166

All 40 sites were visited once between September 2 and September 16, 2013, and samples of 167 stream water, benthic algae and benthic macroinvertebrates were taken. In September, which 168 in Scandinavia is early autumn, benthic algal biomass does not yet show signs of senescence, 169 170 while macroinvertebrate larvae have developed far enough to be countable. Early autumn samples are commonly used for ecological status assessment in Northern European rivers. 171 172 Samples were taken as close as possible to the respective hydrological gauging stations; this was in all cases less than 1 km from the gauging station. No tributaries were present between 173 174 the gauging stations and the respective sites where the samples were collected.

175

176 **2.2 Data collection**

177 Benthic algae

178 At each site, benthic algae were collected from two replicate sub-sites located in riffles,

situated approximately 25 m apart. Chlorophyll *a* (in μ g Chl-*a*/cm²) at each sub-site was

measured from the upper side of five cobbles (with a diameter of approximately 10 cm) using 180 a BenthoTorch, i.e. a Pulse Amplitude Modulated (PAM) fluorimeter developed by BBE 181 Moldaenke GmbH. In Swedish streams, the BenthoTorch has been shown to give similar 182 readings for epilithic Chl a as conventional methods (Kahlert and McKie, 2014). Samples of 183 soft-bodied benthic algae (= algae including cyanobacteria attached to the river bottom or in 184 close contact on or within patches of attached aquatic plants, but excluding diatoms) were 185 taken according to European standard procedures (EN 15708:2009) along an approximately 186 187 10-m length of river bottom using an aquascope (i.e. a bucket with a transparent bottom). At each sub-site, cover (%) of each form of macroscopically visible benthic algae was recorded, 188 and samples were collected and stored separately in vials for species determination. In 189 190 addition, microscopic algae were collected from ten cobbles/stones with diameters ranging between approximately 10 and 20 cm, taken from each site. An area of about 8 x 8 cm from 191 the upper side of each cobble/stone was brushed with a toothbrush to transfer the algae into a 192 beaker containing approximately 1 L of river water from which a subsample was taken. All 193 samples were preserved with a few drops of formaldehyde to a final concentration of 194 approximately 0.5%. The preserved benthic algae samples were later examined under a 195 microscope ($200 - 600 \times magnification$) and all non-diatom algae identified to species, 196 197 wherever possible. For some genera of filamentous green algae whose vegetative forms 198 cannot be determined to species level (e.g. Spirogyra Link or Mougeotia C. Agardh) categories based mainly on filament width were used (see Schneider and Lindstrøm (2009; 199 2011) for further details). The primary identification keys used were Komarek and 200 Anagnostidis (2007), Gutowski and Förster (2009), John et al. (2011) and Komarek (2013). 201 Abundance of each microscopic taxon was estimated in the laboratory as "rare", "common" 202 and "abundant". These estimates were later translated into % cover as 0.001, 0.01 and 0.1%, 203 respectively. Macroscopic algae whose cover was recorded as "<1%" in the field, were noted 204 as "0.1%" for data analysis. For all other taxa, the cover that was estimated in the field was 205 used. Total algal cover was calculated as the sum of cover of all taxa. Note that % algal cover 206 207 includes all types of substrate (including for example algae that grew epiphytic on 208 bryophytes) but does not include diatoms, while Chl a measured with BenthoTorch captured exclusively epilithic algae, but included diatoms. 209

210

211 Macroinvertebrates

At each site, an approximately 50 m long reach was delimited, where we collected ten 212 replicate benthic samples using a Surber net (sampling area: 0.1 m², mesh size: 500 µm). For 213 sampling, the substrate was agitated to a depth of ca. 10 cm for one minute. All benthic 214 samples were immediately preserved in 70 % ethanol and later analyzed in the laboratory. At 215 most sampling locations, the substrate mainly comprised gravel, pebbles, cobbles or small 216 boulders, although at some sites wood, twigs, cones, conifer needles, leaf fragments, aquatic 217 mosses and macrophytes were also recorded. Some of the bed material was partly embedded 218 219 in several reaches, and boulders interspersed the substrate in other reaches. In the laboratory, all benthic samples were sorted using a 500 µm sieve. The benthic macroinvertebrates were 220 classified to the lowest possible taxonomic level, usually species. However, some dipteran 221 222 taxa and microcaddisflies (Hydroptilidae) could only be identified to genus. In addition, bryozoans, nematodes, oligochaetes, water mites, cladocerans, ostracods, non-biting midges 223 and blackflies could not be identified further. 224

225

226 Environmental variables

Hydrological data (discharge in m³ s⁻¹) have been recorded by the Norwegian Water 227 Resources and Energy Directorate (NVE), and are stored in the HYDRA II database. For each 228 site, available discharge data from the five years preceding sampling, i.e. from September 1, 229 230 2008 to August 31, 2013, were extracted from the database as daily averaged values. For one site (site number 25.6, Table S1), data from 2009 were lost, meaning that hydrological 231 characteristics for this site were calculated based on four years of data only. Malfunctioning 232 of the dataloggers caused 13 short gaps in the hydrological data (with an average duration of 9 233 days). Since we had no indication that any extraordinary flow events occurred during these 234 short gaps, the discharge for these days was estimated by linear interpolation between the last 235 day before the onset of malfunctioning and the first day after the malfunctioning datalogger 236 was replaced/fixed. One gap of 172 days was estimated by interpolation from another gauging 237 station close by. Apart from that, the hydrological data for all 40 sites were complete for the 238 period of 5 years. 239

In addition to hydrological regime, we collected data on (i) geographic location and

241 catchment characteristics (latitude, longitude, altitude, catchment size, distance to nearest

lake/reservoir upstream; these data were either taken from Petterson (2004) or recorded from

a digital map of Norway); (ii) water physico-chemistry (Calcium (Ca): NS EN ISO 11885;

total organic carbon (TOC): NS EN 1484; Total phosphorus (TotP): NS EN ISO 15681-2;

- ²⁴⁵ Total nitrogen (TotN): NS 4743); in addition, temperature, pH and conductivity were
- 246 measured with hand-held instruments); and (iii) site characteristics ((a) average water depth
- where the samples were taken; (b) stream width; (c) shading (estimate between 0 = no
- shading and 1 = full shade under trees); (d) % turbulent flow; (e) % cover of boulders (>20
- cm), cobbles (6-20 cm), gravel (2-6 cm), fine gravel (2mm 6 cm), and sand (0.1 2 mm);
- 250 (f) % cover of coarse (> 1 mm) and fine (< 1mm) particulate organic matter (CPOM and
- 251 FPOM) covering the sediment; (g) % cover of bryophytes and macrophytes).
- 252

253 2.3 Data treatment and statistics

254 Benthic algae and macroinvertebrates

255 To explore species composition and abundance of the macroinvertebrate and benthic algal assemblages, respectively, an NMDS (non-metric multidimensional scaling) was computed 256 257 on square-root transformed data. NMDS was used because, in contrast to other ordination methods, it can also handle non-linear responses. The NMDS was computed using the meta 258 259 MDS function in R, version 2.14.2 (R Development Core Team, 2012), extended with the "vegan" package 2.0-4 (Oksanen et al., 2012). Bray–Curtis was used as the dissimilarity 260 measure because it is less dominated by single large differences than many other dissimilarity 261 262 measures (Quinn and Keough, 2002). In addition to NMDS scores, the following response parameters were calculated from the macroinvertebrate and benthic algal taxon lists: (1) taxon 263 richness of macroinvertebrates and benthic algae, respectively; (2) total cover of benthic algae 264 (calculated as sum of cover of all taxa) and density of macroinvertebrates (individuals/m²); 265 (3) cover of cyanobacteria having heterocysts (because they reflect the potential for N-266 fixation (Stancheva et al., 2013); (4) the number of macroinvertebrate individuals in the 267 functional feeding groups shredders (feeding on coarse particulate organic matter (CPOM)), 268 gatherer/collectors (feeding on fine particulate organic matter (FPOM)), grazers/scrapers 269 (feeding on periphyton), and filter feeders (feeding on suspended organic matter), following 270 ASTERICS 4.0.4 (2014), because they provide a link to ecosystem processes; (5) the AIP-271 272 index ("Acidification Index Periphyton"; Schneider and Lindstrøm, 2009) and the acidification index "Raddum 2" (Raddum and Fjellheim, 1984; Raddum 1999) because they 273 274 provide a link to the acidity tolerance of the benthic algal and macroinvertebrate assemblages, 275 respectively; (6) the PIT (Periphyton Index of Trophic Status; Schneider and Lindstrøm,

2011) and ASPT (Average Score Per Taxon; Armitage et al., 1983), because they provide a 276 link to eutrophication and ecological status assessment; (7) the LIFE index (Lotic-invertebrate 277 Index for Flow Evaluation; Extence et al., 1999) was calculated based on macroinvertebrate 278 assemblages using ASTERICS (2014), because it describes flow-preferences of benthic 279 280 invertebrate assemblages. Other response parameters were calculated (e.g. cover of red algae, cover of *Phormidium* sp., diversity indices, relative occurrence of functional feeding groups, 281 282 taxonomic groups such as the number of Ephemeroptera/Plecoptera/Trichoptera, etc.), but omitted from further analysis since they either only occurred in low abundances, or co-varied 283 with other response parameters. After exploratory analysis, data were $\log (x+1)$ -transformed 284 where necessary to improve normality and homoscedasticity (Table 1). For river biota, results 285 of the two benthic algal and ten macroinvertebrate samples per site were averaged, and linear 286 models were computed using the MASS-package in R (Venables and Ripley, 2002). 287 However, we also tested linear mixed models on the complete dataset (including two replicate 288 benthic algal samples per site, and 10 replicate macroinvertebrate samples per site), using the 289 nlme-package in R (Pinheiro et al., 2012), and "site" was included as random factor. In order 290 to enable unbiased comparisons of the response variables between regulated and unregulated 291 sites, their values had to be corrected for the differences in explanatory variables that occurred 292 between regulated and unregulated sites (i.e. catchment size, altitude, TN and TOC; the last 293 three also correlated with each other). In order to do so, we computed a set of multivariate 294 linear models, separately for each response variable that was significantly correlated with one 295 or several of the explanatory variables whose values significantly differed between regulated 296 and unregulated rivers. We then selected, separately for each response variable, the best 297 model by using an information-theoretic approach (Akaike information criterion; AIC), and 298 corrected the value of each response variable based on the slope of the respective best model. 299

300

301 Environmental variables

At one site, we forgot to record conductivity and temperature (NVE number 36.32; Table

A.1). The missing values were estimated from the variables that correlated closest with

304 conductivity and temperature at the remaining 39 sites (i.e. a linear correlation between log

305 (conductivity) and log (Calcium) (Pearson r = 0.94; R²=0.88), as well as temperature and log

306 (TOC) (Pearson r = 0.78; R^2 =0.62)). In order to characterize sediment composition at each

- 307 site, a PCA (principal component analysis) was calculated from the scaled data on % cover of
- 308 boulders, cobbles, gravel, fine gravel, sand, CPOM, FPOM and bryophytes, using the vegan-

- package in R. The first two axes explained 55% of variation; PC1 was positively related with
 boulders and bryophytes, and negatively with cobbles and gravel; PC2 was positively related
- 311 with fine gravel, sand and cover of CPOM (Table A.2).

312 Richter et al. (1996) defined several "indicators of hydrologic alteration" to statistically characterize variation in river flow. They are categorized into the following five groups, 313 which are considered useful to quantitatively evaluate the impact of hydrological regime on 314 aquatic biota: (1) mean discharge values, (2) magnitude of annual extremes, (3) timing of 315 annual extremes, (4) frequency and duration of high and low pulses, and (5) rate of change. 316 We calculated 77 variables from the daily averaged discharge values, which were assigned to 317 these five categories (Table 1). In addition, the base flow index (= the ratio of base flow to 318 total streamflow) was calculated using the "lf stat"-package in R (Koffler, 2013). In order to 319 enable comparisons among sites (i.e. independent of river size), the values for the "indicators 320 321 of hydrologic alteration" at each site were calculated relative to the average discharge during the five years preceding sampling. In order to capture effects of both "long-term" flow 322 regime, as well as recent events, all streamflow characteristics were calculated for the total 323 period of five years preceding sampling of benthic algae and macroinvertebrates ("long-324 term"), and in addition for the one year preceding sampling ("recent"). 325

326 Together with site characteristics and water chemistry, 97 environmental variables were compiled for each site. After exploratory analysis, data were transformed where necessary to 327 328 improve normality and homoscedasticity (Table 1). Prior to data analysis, we inspected scatter plots in order to search for possible non-linear (e.g. hump-shaped) relationships. No 329 indications of such patterns were found, however. We used ANOVA to compare regulated 330 with unregulated sites. In order to analyze the influence of overall flow regime on each 331 332 response variable, we summarized the 78 hydrological variables into principal components using the vegan-package in R. However, because each principal component represents a 333 plethora of hydrological variables whose individual importance for the response parameters 334 cannot be deduced, we also calculated a correlation matrix between explanatory and response 335 variables. We then summarized the strongest correlations and interpreted their importance 336 against the background of published information. 337

338

339 **3. Results**

340 **3.1 Differences between regulated and unregulated sites**

We attempted to select our sampling sites in such a way that no environmental variable except 341 flow regime would differ between regulated and unregulated sites. However, this was not 342 possible, since the position of the hydrological gauging stations obviously was tailored to the 343 management needs of the Norwegian Water Resources and Energy Directorate, and not to our 344 project. As a consequence, the regulated sites in our dataset not only differed in flow regime 345 from unregulated sites, but they also had a larger watershed, were situated at a lower altitude, 346 and had slightly higher TN and TOC concentrations (Table 1). Apart from that, only river 347 flow differed between regulated and unregulated sites, with regulated sites having higher 348 relative discharges in winter, lower relative discharges in spring, and smaller differences 349 between upper and lower percentiles (see Table 1 for summary statistics, and Table A.6 for a 350 complete overview over hydrological characteristics at each sampling site). After accounting 351 for the differences in catchment size, altitude, TN and TOC (Table A.3), none of the response 352 variables differed between regulated and unregulated sites, despite the differences that 353 occurred in river flow (Table 1). 354

We then used PCA to summarize the 78 hydrological variables into principal components, 355 reflecting overall flow regime. The first two PCs explained 55% of the variation in 356 hydrological variables (Table A.5). High scores along PC1 corresponded to streams with 357 relatively high winter discharges, generally low 7-day maxima, and small differences between 358 upper and lower percentiles, i.e. high scores along PC1 characterized sites with a 359 comparatively "stable" flow regime. High scores along PC2 corresponded to a hydrological 360 regime dominated by run-off (a low BFI indicates a high contribution of run-off (and a low 361 contribution of base-flow) to total streamflow), steeply rising and falling limbs, and relatively 362 363 high autumn discharges (Table A.5), i.e. high scores along PC2 characterized "flushy" rivers. Higher principal components explained little of the total variation (no axis explained more 364 365 than 10%), and few strong relationships with explanatory variables occurred (data not shown), such that higher PC axes could not be meaningfully interpreted. Although there was 366 367 considerable overlap, regulated rivers had higher scores along PC1, i.e. they had a more "stable" flow regime (Table 1; Fig. 2). 368

369

370 3.2 Effect of flow regime compared to other environmental variables on benthic algal 371 and macroinvertebrate assemblages and traits

In order to separate the effects of flow regime from those of other (correlated) explanatory 372 variables, regulated and unregulated sites were analyzed separately (but PC scores for flow 373 regime were calculated from the pooled dataset, and the results were later separated into 374 regulated and unregulated sites; this was done in order to ensure that characterization of flow 375 regime was comparable between regulated and unregulated sites). In unregulated rivers, flow 376 regime (characterized as PC_{hvdr}1 and 2) was correlated with half of the other explanatory 377 variables, particularly geographic location, the distance to the nearest upstream lake, 378 379 catchment size, some water chemical variables and temperature (Table 2). This was not surprising, since the flow regime of unregulated rivers is determined by catchment 380 characteristics and climate, which in turn are related to water chemistry and geographic 381 location. Likewise, more than half of the response variables were correlated with flow regime 382 Table 2). However, due to the many correlations among flow regime and the other 383 explanatory variables (see above), deducing possibly causal relationships between flow 384 regime and responses was not possible. 385

In contrast, flow regime of the regulated rivers exhibited fewer correlations with other 386 explanatory variables (Table 2). Again, this was not surprising since the flow regime of 387 regulated rivers is tailored to human needs so that climate and geology less affect it. 388 Nevertheless, PC_{hydr}1 was also in regulated rivers correlated with latitude and temperature, 389 and PC_{hydr}2 was correlated with catchment size, % turbulent flow and stream width (Table 2). 390 391 However, in regulated rivers, only PC_{hydr}1 scores correlated with macroinvertebrate species assemblages (reflected as NMDS1 values), as well as with LIFE scores (Table 2; Fig. 3). No 392 other correlations among PC axes and any of the response variables occurred in regulated 393 394 rivers. Because PC_{hydr}1 in regulated rivers correlated with latitude and temperature (Table 2), this indicates that macroinvertebrate species composition and LIFE scores were affected by 395 396 latitude, temperature, or flow regime (if we disregard a possible effect of other variables which we have not measured). The absence of other correlations among PC axes and response 397 398 variables in regulated rivers indicates that all other relationships that occurred in rivers with a natural flow regime, were unlikely to be caused by flow regime, but by one (or several) of the 399 400 explanatory variables that correlated with PC_{hvdr}1 or 2 (Table 2; note that data ranges were comparable between regulated and unregulated rivers (Table 1)). In other words: our results 401 indicate that flow regime may have affected macroinvertebrate species composition and LIFE 402 403 scores, but no other structural or functional characteristics of benthic algae and macroinvertebrates. 404

In order to explore this further, we computed a set of multivariate linear models, separately for LIFE and NMDS1.MI, and selected the best models based on AIC. Although temperature explained most of the variance in NMDS1 scores, and latitude explained most of the variance in LIFE scores, PC_{hydr}1 was retained in both cases (Table 3). This indicates that flow regime significantly affected macroinvertebrate species assemblages, as well as LIFE scores (with lower LIFE scores, indicating a macroinvertebrate assemblage that prefers lower flow, at sites

411 with a "stable" flow regime, i.e. high scores along PC_{hydr}1).

However, PC axes represent summarized descriptors of flow regime. Therefore, instances 412 where one or few individual components of flow regime (e.g. maximum June discharge, 413 number of high pulses, etc.) are influential may be overlooked. To explore which of the 414 explanatory variables, including each of the 78 hydrological variables, were most closely 415 416 related to the response variables, we calculated a correlation matrix and summarized the strongest correlation coefficients (Table 4). Complete results are given in appendix (Table 417 A.4). Regulated and unregulated sites were pooled, because none of the above results 418 indicated a major effect of river regulation, the higher number of sites in the pooled dataset 419 reduced the chance of accidentally significant relationships (false positives), and the different 420 autocorrelations among explanatory variables in regulated and unregulated rivers often 421 prevented a meaningful interpretation of the results from separated datasets. We decided 422 against modelling response variables from the explanatory variables, because the high number 423 of autocorrelations greatly hampered differentiating between possibly causal and random 424 relationships. Instead, we interpreted the results of the correlation matrix against the 425 background of published information (Table 4). 426

427

428 4. Discussion

429 Effects of river flow compared to other environmental variables

Hypothesis 2, which stated that assemblages adapted to low flow conditions would occur at
sites with a more stable flow regime, was accepted for macroinvertebrates, but not for benthic
algae. Overall flow regime, characterized as principal components calculated from 78
hydrological variables over the five years preceding sampling, affected macroinvertebrate
species assemblages, reflected in NMDS and LIFE scores (Table 3). LIFE is based on
macroinvertebrate taxa associated with different "flow groups" (from "rapid" via "slow" to
"standing" and "drought resistant"), and was designed to assess changes in prevailing flow

regimes (Extence et al., 1999). An effect of flow regime on LIFE scores therefore simply 437 meets expectations. Although many benthic macroinvertebrate taxa can live under varying 438 flow regimes (Statzner et al. 1988), some taxa including heptageniid mayfly nymphs and 439 blackfly larvae exhibit behavioural and morphological adaptations to high current velocities 440 (Hart et al. 1991, Weissenberger et al. 1991). This likely explains the change in 441 macroinvertebrate species composition, reflected in NMDS scores, with flow regime. 442 Short-term effects of extreme events on macroinvertebrates and benthic algae are a well-443 known phenomenon (Extence et al., 1999; Monk et al., 2008; Power et al., 2008). However, 444 even though flood scour and dewatering indeed rejuvenate riverine ecosystems, 445 macroinvertebrates and benthic algae rapidly reassemble after such events (Power et al., 446 2013). Rapid reassembly will lead to the absence of correlations between long-term flow 447 regime and response variables. Given rapid recolonization, a relation between overall flow 448 449 regime and a biological response will only emerge once the species pool, from which recolonization occurs, has been affected. Our results indicate that overall flow regime (as 450 characterized by PC1_{hydr}) affected the species pool of macroinvertebrates, but not of benthic 451 algae. This indicates that macroinvertebrate assemblages are more sensitive to long-term 452 overall flow regime than benthic algae. This is in accordance with earlier studies that analyzed 453 flood effects on macroinvertebrates and periphyton, which either reported that both were 454 affected ("high floods"; Danehy et al., 2012; Fuller et al., 2011; Robinson and Uehlinger, 455 2008), or neither of the two was affected ("low floods"; Tonkin and Death, 2014), or that 456 macroinvertebrates were more sensitive than periphyton (Robinson, 2012). 457

458 We have no evidence that overall flow regime affected benthic algal assemblages, taxon

richness, biomass, potential N-fixation, or indices used for ecosystem status assessment

460 (Table 2). Neither did flow regime affect macroinvertebrate taxon richness, overall density,

density of functional feeding groups or indices used for ecosystem status assessment (Table

462 2). This indicates that, though macroinvertebrate species identity shifted with changing flow

regime (along NMDS1), the exchanged taxa had similar functional feeding habits. We would

like to stress that these inferences are only valid for flow regimes that are within the

465 variability we experienced in our dataset (Tables 1, A.6). For example, "extreme" regulation

466 causing streambed drying did not occur at our sites, due to the climatic conditions in Norway,

467 and because of environmental flow regulations aimed at avoiding streambed drying

468 (Alfredsen et al., 2012). If regulation had caused streambed drying, consequences for biota

would probably have been severe (Bonada et al., 2007; Hille et al., 2014; Elias et al., 2015;

470 Verdonschot et al., 2015).

However, overall flow regime is a summary parameter which may overlook potential effects
of individual components of flow regime on river biota. We therefore also analyzed the
effects of each of the 78 hydrological variables which constitute flow regime, and compared
them with the effects of water chemistry and habitat characteristics. Although the large
number of autocorrelations among explanatory variables prevented relating the observed
differences in response variables to single explanatory variables with confidence, the
following inferences were possible (Table 4);

(1)Macroinvertebrate and benthic algal species assemblages were more closely related to 478 water chemical than to hydrological variables; benthic algal assemblages were best 479 explained by water calcium concentrations and conductivity (Tables A.4, 4); Calcium and 480 conductivity were correlated with each other, and their effect on algal assemblages is 481 probably related to the increased availability of inorganic carbon in "hard water"; a 482 relationship between benthic algal assemblages and water calcium concentrations is 483 common and has also been shown in Norway before (Schneider, 2011). Benthic algal 484 assemblages also were related to water TP concentrations (as expected; see Schneider and 485 486 Lindstrøm, 2011), but the correlation was weak due to the low number of sites with high TP concentrations in our data. Macroinvertebrate assemblages were closest related to water 487 temperature and TOC concentrations; temperature and TOC were correlated with each 488 other, but both are well-known to affect macroinvertebrates: the effect of temperature is 489 490 related to species requirements with respect to growth and egg hatching (Lillehammer, 1987; Lillehammer et al., 1989), while TOC has multiple effects, including its use as food 491 for decomposers (Thomas, 1997). 492

(2) flow maxima were related to algal taxon richness, and flow minima to macroinvertebrate 493 taxon richness; however, recovery probably is fast; high June (i.e. three months before 494 sampling) flow maxima were (weakly but significantly) associated with low benthic algal 495 taxon richness (Tables A.4, 4); this may be explained by a short-term effect of flood scour 496 (Biggs and Smith, 2002). However, neither Biggs and Smith (2002) nor our own results 497 with respect to hydrological variables calculated from five years-flow regime (Tables 2, 4, 498 A.4) indicate long-lasting effects of flow regime on benthic algal richness patterns in 499 streams. This suggests that sufficient algae remain after flood scouring to permit rapid 500 501 recolonization. Benthic algal taxon richness was weakly but significantly negatively

502 correlated with water TP-concentrations; such a relationship has been found before (Schneider et al., 2013b) and may be explained by the classical concept of niche theory, 503 where taxon richness decreases with increasing nutrient supply due to the exclusion of taxa 504 by superior competitors (Stevens et al., 2004; Wassen et al., 2005). Macroinvertebrate 505 taxon richness generally increased with increasing minimum discharges, and the strongest 506 relation was with May and November minimum discharges during the year before 507 sampling (Tables 4, A.4). Temporary drying of the riverbed may affect the densities of 508509 benthic macroinvertebrates and hence species diversity (Clarke et al., 2010). Consequently, if lower minimum discharge levels resulted in partial drying of the riverbed, then this may 510 explain the finding of lower macroinvertebrate richness where minimum discharge was 511

512 lowest.

(3) benthic algal biomass and cover was related to water chemistry and river flow, but their 513 relative importance was uncertain; epilithic Chl a and total algal cover were positively 514 correlated with water temperature, TOC concentrations and winter discharges, and 515 negatively with summer discharges (Tables 4, A.4). Since these variables were correlated 516 with each other, their relative importance for benthic algal biomass and cover could not be 517 deduced with confidence. Each of them may in fact be influential: temperature affects algal 518 growth (Piggott et al., 2015), which may lead to a positive relation between temperature 519 and benthic algal biomass in streams (Schneider, 2015); TOC may be beneficial by 520 preventing damage caused by ultraviolet light (Kelly et al., 2001) and by providing a 521 nutrient source that is accessible for some taxa via phosphatase (Whitton et al., 1991); high 522 winter discharges may prevent freezing and drying damage (Lind and Nilsson, 2015), and 523 high summer discharges may harm due to scouring (Francoeur and Biggs, 2006). 524 (4)macroinvertebrate density was poorly related to water chemistry or river flow; this is at 525 odds with earlier studies which observed higher macroinvertebrate densities at phosphorus-526 enriched sites (Rader and Richardson, 1992; McCormick et al, 2004); we suggest that our 527 dataset contained too few clearly nutrient-enriched sites; this may have prevented the 528 529 detection of nutrient effects given that disturbance regime may modify macroinvertebrate responses (Gafner and Robinson, 2007); in our data, the closest relation (Pearson r = 0.47) 530 occurred with autumn minimum discharges (high October and November minimum 531 discharges were associated with higher macroinvertebrate density; Tables 4, A.4); this may 532

be related to partial drying of the riverbed at low minimum discharges (temporary drying

534 of the riverbed affects the densities of benthic macroinvertebrates; Clarke et al. 2010).

535 (5) we were unable to confidently establish relationships between species traits (related to potential nitrogen fixation, grazing, filtering, degradation of CPOM and FPOM) and 536 water chemistry or river flow; The abundance of N-fixing algae has earlier been shown to 537 be related to water nitrate (plus nitrite) concentrations (Stancheva et al., 2013; Gillett et al., 538 539 2016), a parameter which we have not measured (only total N). We therefore cannot exclude that a relation between water nitrate concentrations and the abundance of N-fixing 540 541 algae existed also in our dataset. Low flow minima indeed tended to decrease the number of filter feeders (Tables 4, A.4), which may be explained by their dependence on a 542 minimum flow to transport food particles. However, many autocorrelations occurred 543 among hydrological variables, and – most importantly – there was also a negative 544 relationship between the number of filter feeders and the distance between the sampling 545 site and the nearest upstream lake/reservoir (Table A.4). An enhanced number of filter 546 feeders in lake outlets is a well-known phenomenon (Malmqvist and Eriksson, 1995). We 547 therefore deem a relationship between the number of filter feeders and flow minima 548 uncertain, and request further studies before conclusions may be drawn with confidence. 549 There was a weak but significant trend that more grazers and more collectors occurred at 550 high pH (Table A.4), but autocorrelations occurred with geographic position. The absence 551 of strong relationships between water chemistry, hydrological variables and 552 macroinvertebrate functional feeding groups may be related to many macroinvertebrate 553 species showing flexible feeding habits (Rawer-Jost et al., 2000), but also to the 554 overarching effect of riparian vegetation on stream food webs, via litter input (Wallace et 555 556 al., 1997), as well as to the manifold interactions between hydrochemistry, flow, primary producers and consumers (Lamberti et al., 1991; Wallace et al., 1997) which may 557 confound straightforward relationships. 558

(6) We found no indications that river flow affected macroinvertebrate and benthic algal
 acidification indices; both acidification indices (AIP for benthic algae and Raddum 2 for
 macroinvertebrates) were most closely related to pH. These indices were designed to

reflect pH (Raddum and Fjellheim, 1984; Raddum 1999; Schneider and Lindstrøm, 2009),

- and we therefore suggest that all other relationships among these indices and other
 explanatory variables (Table A.4) were due to their autocorrelation with pH.
- 565 (7) the LIFE index was useful for characterizing overall flow regime; The LIFE index was
- 566 most closely related to latitude, but it also was correlated with overall flow regime
- 567 (characterized as principal components; Table 2). Across our sampling sites, highest
- 568 precipitation generally occurred at the southernmost sites, and precipitation changed

roughly linearly with latitude (www.met.no). Consequently, latitude correlated with overall 569 flow regime, and may - in our dataset - indeed be a surrogate variable for long-term flow 570 regime. The other environmental variables that were related to the LIFE index (Table A.4) 571 also were correlated with latitude (data not shown). There are several arguments which 572 together suggest that the LIFE index indeed may be useful for characterizing overall flow 573 regime (also in Norway where it previously has not been tested): (i) the LIFE index was 574 designed to asses changes in prevailing flow regimes (Extence et al., 1999), and a recent 575 adaptation of the LIFE index to New Zealand also primarily correlated with hydrological 576 variables instead of water chemistry (Greenwood et al., 2016); (ii) the LIFE index 577 correlated with PC1_{hydr}, and (iii) among 97 environmental variables in our dataset, the 578 579 LIFE index was most closely related to latitude, which in our dataset likely reflects overall flow regime. 580

581

582 *Effects of river regulation*

583 Hypothesis 1, which stated that regulated sites would have a more stable flow regime than unregulated sites and that this would lead to fewer macroinvertebrate and benthic algal taxa in 584 regulated than in unregulated sites, was only partly accepted. Regulated sites indeed had a 585 more stable flow regime (Table 1), but this was not associated with reduced taxon richness. 586 587 Neither have we found differences in macroinvertebrate and benthic algal assemblages and functional traits between regulated and unregulated rivers (Table 1). The absence of 588 systematic differences in aquatic biota between regulated and unregulated sites may at first 589 sight be surprising, but is in line with results of Poff and Zimmerman (2010), who were 590 591 unable to develop general relationships between flow alteration and ecological response. River regulation may have manifold consequences, affecting not only river flow, but also 592 water temperature, nutrient concentrations, organic matter and alkalinity/pH, among others 593 (reviewed by Gracey and Verones, 2016). These water physico-chemical variables were 594 among those that explained most of the variability in benthic algal and macroinvertebrate 595 596 assemblages and biomass (Table 4). To which degree and in which direction water quality 597 and quantity are affected by an individual hydropower plant depends on the location, design and management of the dam/power plant, such that effects vary between sites (Gracey and 598 Verones, 2016). For example, river regulation may increase or decrease water temperature 599 (Gracey and Verones, 2016). This will lead to different responses among river biota. For 600 example, mass developments of macrophytes and benthic algae may occur downstream the 601

outlet of some but not other hydropower plants (Johansen et al., 2000). Also, the severity and 602 direction of responses in fish communities and traits to river regulation vary widely (Murchie 603 et al. 2008). The absence of systematic differences between regulated and unregulated rivers 604 therefore does not contradict observed differences between upstream and downstream 605 locations of dams (Lessard and Hayes, 2003), or before and after river regulation (Dejalon 606 and Sanchez, 1994) at specific river sites. The question is, however, whether these observed 607 differences at specific sites were caused by the changes in river flow, or by concomitant 608 609 changes in water physico-chemistry.

Our results indicate that overall flow regime affected macroinvertebrate assemblages 610 611 (reflected in NMDS and LIFE scores), but the difference in flow regime between regulated and unregulated sites was not sufficiently large to be reflected in macroinvertebrate 612 613 assemblages. The results also indicated that many of our response variables primarily respond 614 to water physico-chemical variables (Table 4). Together with the fact that river regulation may affect both flow regime and water physico-chemistry to various degrees, this may 615 explain the absence of consistent differences between regulated and unregulated sites. It also 616 explains the observed wide variations in the severity and direction of biological responses in 617 regulated rivers (Murchie et al., 2008). Our data carefully suggest that changes in water 618 physico-chemistry caused by river regulation may be equally important for benthic algae and 619 macroinvertebrates than changes in river flow. Understanding these relationships is essential 620 for improvement of river management practices, and for planning remediation measures to 621 minimize effects of river regulation on aquatic biota. In addition, using data on river flow for 622 relating observed changes in riverine biota to river regulation may lead to misleading results 623 624 when concomitant changes in water physico-chemical parameters are not taken into account.

625 We observed no differences between regulated and unregulated rivers in any of the indices used for ecological status assessment (Raddum 2 and AIP for acidification, ASPT and PIT for 626 eutrophication/organic pollution; Table 1). Both acidification indices responded closely to pH, 627 irrespective of river regulation (a similar analysis for PIT and ASPT was not possible because 628 too few eutrophic sites occurred in our dataset, preventing a meaningful interpretation of 629 correlations). Consequently, our results (i) give no reason for defining "good ecological 630 potential" in regulated rivers differently than "good ecological status" in unregulated rivers, 631 and (ii) suggest that the existing assessment systems for macroinvertebrates and benthic algae 632 with respect to acidification and eutrophication (Raddum 2, AIP, ASPT, PIT) may also be 633 applicable in regulated rivers. 634

- 635 Our results indicate that long-term modification of flow regime towards more "stable"
- 636 conditions (as characterized by PC1_{hydr}) may lead to changes in macroinvertebrate
- assemblages, which are reflected in the LIFE index (Extence et al., 1999). The LIFE index
- therefore seems a suitable response parameter for monitoring long-term changes in flow
- 639 regime (time series data).
- 640

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- 647

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Fig. 1: map of 40 sampling sites in Norway; $\hat{o} =$ regulated (modified flow regime), $\hat{o} =$ unregulated (natural flow regime)



Fig. 2. PCA of 78 hydrological variables (Table 1) characterizing river flow at regulated (ô)

887 and unregulated (Ô) river sites



Fig. 3. Scatter plots of response and explanatory variables that were significantly correlated
with flow regime at both regulated (ô) and unregulated (ô) river sites (Table 2). NMDS.MI =
non-metric multidimensional scaling scores (along axes 1) for macroinvertebrates, LIFE =
LIFE index for macroinvertebrates.

895 Table headings

- **Table 1.** Summary statistics for regulated and unregulated sites, and p-values for differences
- 898 between groups (t-test). Significant differences are marked in bold. Underlined p-values were
- 899 calculated from corrected values of the response variables, i.e. which were corrected for
- 900 differences in explanatory variables other than flow regime (by using the models given in
- 901 Table A.3).

	unregulated							p-value for difference between groups			
	Ν	Mean	Std.Dev.	5 percentile	95 percentile	Ν	Mean	Std.Dev.	5 percentile	95 percentile	р
explanatory variables longitude (east; UTM 32)	20	496827	79387.4	347093	610168	20	476339	89443.1	351318	614134	0.448
latitude (north; UTM 32) log (x+1) distance to nearest lake/reservoir upstream (km)	20 20	6735184 0.93	143972.1 0.57	6507868 0.08	6917091 2.00	20 20	6669749 0.89	129931.3 0.47	6512172 0.18	6917171 1.59	0.140 0.781
log catchment size (km2) altitude (m asl)	20 20	2.23 508.9	0.57 270.5	1.08 121.5	2.89 987.5	20 20	2.71 310.4	0.51 216.0	1.72 17.5	3.34 680.0	0.007 0.014
Shading (%) log (x+1) Tot-P/L [µg P/I]	20 20	0.29 0.83	0.24 0.55	0.00 0.30	0.75 2.22	20 20	0.29 0.70	0.17 0.23	0.06 0.39	0.63 1.17	1.000 0.337
Tot-N/L [µg N/I] log (x+1) TOC [mg C/I]	20 20	158.0 0.44	82.0 0.19	63.0 0.18	315.0 0.75	20 20	233.1 0.57	93.5 0.19	101.0 0.22	417.5 0.86	0.010 0.039
log (x+1) Ca [mg/l] log conductivity (µs/cm)	20 20	0.45 1.23	0.23 0.25	0.18 0.87	0.82 1.68	20 20	0.52 1.33	0.32 0.33	0.12 0.96	1.20 2.06	0.439 0.296
temperature (degree C) pH	20 20	10.84 6.83	2.76 0.43	6.65 5.89	15.85 7.30	20 20	12.30 6.81	2.36 0.64	8.10 5.36	16.00 7.69	0.081 0.879
% turbulent flow average depth (m)	20 20	75.75 0.35	34.57 0.13	6.25 0.15	100.00 0.55	20 20	62.50 0.36	37.20 0.12	1.25 0.19	100.00 0.56	0.251 0.663
width (m) sediment PC1	20 20	22.08 -0.01	11.83 0.72	6.25 -0.96	47.50 1.15	20 20	27.45 0.01	14.34 0.65	10.00 -0.97	52.50 0.99	0.204 0.945
sediment PC2 mean discharge	20	0.01	0.44	-0.57	0.76	20	-0.01	0.86	-0.65	2.06	0.961
log (x+1) mean discharge january relative to mean (%)	20	1.38	0.24	1.01	1.87	20	1.69	0.30	1.24	2.18	0.001
log (x+1) mean discharge march relative to mean (%) log (x+1) mean discharge march relative to mean (%)	20	1.42	0.22	1.04	1.84	20	1.73	0.32	1.34	2.10	0.000
mean discharge may relative to mean (%)	20	222.8	81.1	106.1	357.1	20	160.4	40.56	68.1	280.3	0.012
mean discharge july relative to mean (%) mean discharge july relative to mean (%)	20	149.83	62.99	67.36	274.95	20	132.05	52.44	44.78	222.25	0.114
mean discharge august relative to mean (%) mean discharge september relative to mean (%)	20	138.74 116.33	38.08 20.62	69.39 72.94	195.99	20	123.08 119.39	37.10 37.55	54.12 72.57	184.15 204.09	0.196
mean discharge october relative to mean (%) log (x+1) mean discharge november relative to mean (%)	20	92.69	30.86 0.27	1.41	2.24	20	99.04 1.93	31.42 0.23	42.98	153.02 2.21	0.523 0.162
log (x+1) mean discharge december relative to mean (%) one year before sampling	20	1.54	0.22	1.14	1.91	20	1.77	0.22	1.39	2.15	0.002
log (x+1) mean discharge january 1 ybs relative to mean log (x+1) mean discharge february 1 ybs relative to mean	20 20	0.11	0.08	0.03	0.30	20	0.19	0.11	0.04	0.40	0.009
log (x+1) mean discharge march 1 ybs relative to mean log (x+1) mean discharge april 1 ybs relative to mean	20 20	0.04 0.17	0.02	0.00	0.06 0.45	20 20	0.14 0.26	0.13 0.11	0.01 0.08	0.43 0.42	0.001 0.035
mean discharge may 1 ybs relative to mean mean discharge june 1 ybs relative to mean	20 20	3.82 2.13	1.33 0.84	1.85 0.57	5.91 3.34	20 20	2.66 1.55	1.45 0.71	0.57 0.41	5.16 2.60	0.012 0.023
mean discharge july 1 ybs relative to mean mean discharge august 1 ybs relative to mean	20 20	0.85 1.14	0.49 0.47	0.24 0.40	1.83 2.07	20 20	0.93 1.04	0.56 0.47	0.31 0.34	2.09 1.82	0.666 0.497
mean discharge september 1 ybs relative to mean mean discharge october 1 ybs relative to mean	20 20	0.94 0.87	0.40 0.46	0.53 0.32	1.78 1.82	20 20	0.95 1.03	0.50 0.54	0.26 0.31	1.93 2.21	0.942 0.335
log (x+1) mean discharge november 1 ybs relative to mean log (x+1) mean discharge december 1 ybs relative to mean	20 20	0.28 0.12	0.17 0.04	0.08	0.58 0.18	20 20	0.34 0.20	0.14 0.10	0.11 0.09	0.60 0.39	0.229
magnitude of extremes max relative to mean (%)	20	1299.2	298.7	786.6	1774.9	20	1234.1	813.1	275.4	2660.9	0.739
min relative to mean (%) 95 perc. relative to mean (%)	20 20	5.1 346.8	3.7 41.9	0.2 270.1	12.3 422.1	20 20	6.8 282.0	6.5 72.8	0.0 164.5	19.9 392.2	0.312
log (x+1) 5 perc. relative to mean (%) difference min-max relative to mean (%)	20 20	0.9 1294.2	0.2 298.6	0.6 784.1	1.2 1771.6	20 20	1.2 1227.3	0.3 816.1	0.7 256.4	1.7 2657.4	0.004 0.733
difference 95-5 percentile relative to mean (%) difference 99-1 percentile relative to mean (%)	20 20	338.4 635.6	44.8 107.3	254.7 463.8	417.5 807.9	20 20	263.2 492.4	83.7 205.3	117.5 156.1	382.8 793.2	0.001 0.009
75 perc. relative to mean (%) 25 perc. relative to mean (%)	20 20	129.8 20.2	14.0 6.4	109.0 8.8	155.9 31.5	20 20	128.7 35.7	24.2 20.0	102.8 16.0	178.8 73.5	0.861
average yearly max relative to mean discharge (%) coefficient of variation yearly max	20 20	910.4 0.54	148.5	638.8 0.46	1125.3	20 20	796.1 0.52	472.9	218.4	1777.9	0.309
average yearly min relative to mean discharge (%) coefficient of variation yearly min	20 20	9.01	4.16	3.96	17.59	20	13.99	9.91 0.31	2.78	34.62	0.045
7 day max 5 years relative to mean discharge 7 day min 5 years relative to mean discharge	20 20	8.59	2.30	5.27	12.40	20	6.41	2.93	2.08	10.78	0.013
log (x+1) max discharge january 1 ybs relative to annual mean log (x+1) max discharge january 1 ybs relative to annual mean	20	0.22	0.23	0.03	0.75	20	0.30	0.21	0.06	0.77	0.265
log (x+1) max discharge march 1 ybs relative to annual mean log (x+1) max discharge march 1 ybs relative to annual mean	20	0.04	0.02	0.00	0.07	20	0.17	0.15	0.02	0.48	0.001
max discharge may 1 ybs relative to annual mean maximum max discharge may 1 ybs relative to annual mean max discharge inc. 1 ybs relative to annual mean	20	10.52	3.68	4.38	16.84	20	9.21	6.81	1.09	20.79	0.456
max discharge july 1 ybs relative to annual mean	20	2.05	0.76	0.58	3.19	20	1.96	1.21	0.50	4.74	0.782
max discharge adgust i yus relative to annual mean max discharge september 1 ybs relative to annual mean	20	2.29	1.52	1.00	6.51	20	2.44	2.17	0.85	8.38	0.806
log (x+1) max discharge october 1 yos relative to annual mean log (x+1) max discharge november 1 ybs relative to annual mean	20	0.45	0.31	0.13	4.92	20	0.52	0.26	0.16	1.03	0.456
log (x+1) min discharge becenter i yos relative to annual mean	20	0.24	0.03	0.00	0.09	20	0.35	0.24	0.00	0.34	0.009
log (x+1) min discharge reoruary 1 ybs relative to annual mean log (x+1) min discharge march 1 ybs relative to annual mean	20	0.04	0.02	0.00	0.06	20	0.12	0.11	0.01	0.35	0.003
log (x+1) min discharge april 1 ybs relative to annual mean log (x+1) min discharge may 1 ybs relative to annual mean	20 20	0.03	0.02	0.00	0.05	20	0.07	0.06	0.01	0.18	0.002
min discharge june 1 ybs relative to annual mean log (x+1) min discharge july 1 ybs relative to annual mean	20 20	0.97 0.14	0.52 0.10	0.21 0.02	1.94 0.33	20 20	0.74 0.15	0.50 0.11	0.09	1.63 0.38	0.164 0.646
log (x+1) min discharge august 1 ybs relative to annual mean min discharge september 1 ybs relative to annual mean	20 20	0.13 0.46	0.10	0.03	0.29 0.84	20 20	0.15 0.41	0.11 0.26	0.02	0.37 0.87	0.575 0.500
log (x+1) min discharge october 1 ybs relative to annual mean log (x+1) min discharge november 1 ybs relative to annual mean	20 20	0.14	0.06	0.06	0.26	20 20	0.16	0.09	0.05	0.33	0.505 0.675
timing of extransional days of examples a second of the second se	20	0.19	0.09	0.04	0.32	20	0.38	0.30	0.05	1.02	0.010
days between sampling and last maximum	20	132	71.4	92	299	20	120	69.8	55	317	0.608
average Julian day maximum	20	152	45.1	73	218	20	175	72.7	39	307	0.236
month with highest discharge	20	6	1.9	4	11	20	5	2.2	1	10	0.358
number of bids released to the fore sampling	20	40	7.3	28	49	20	39 10	10.3	19	56	0.699
total number of high pulses in 5 years	20	44	4.7	23	81	20	46	19.9	21	87	0.678
rate of change	20	0.00	0.12	0.67	1.05	20	0.82	0.24	0.25	1 22	0.454
 log (x+r) maximum name minimum falling limb relative to average outcoming 5 years log (sqrt of quadrat of minimum falling limb relative to average discharge 5 years average rising limb relative to average discharge 5 years 	20 20 20	-0.72	0.13	-1.01	-0.36	20 20 20	-0.66	0.43	-1.37	0.00	0.580
average falling limb relative to average discharge 5 years base flow index	20	-0.19	0.08	-0.36	-0.11	20	-0.19	0.11	-0.45	-0.06	0.903
BFI 5 years BFI 1 years before sampling	20 20	0.519 0.517	0.1 0.1	0.32 0.31	0.70 0.69	20 20	0.579 0.582	0.2 0.2	0.25 0.22	0.85 0.87	0.198 0.206
Principal components of hydrological variables PC1 hydr	20	-0.56	1.3	-1.73	0.77	20	0.56	1.3	-1.12	3.13	0.002
PC2 hydr Response variables	20	0.01	1.1	-1.28	2.03	20	-0.01	1.3	-1.54	2.85	0.978
species assemblages NMDS1 algae	19	0.06	0.56	-0.61	1.93	20	-0.06	0.61	-1.56	0.52	0.517
NMUS2 algae number of taxa algae	19 20	-0.03 17.58	0.43	-0.86 1.50	1.13	20 20	U.03 18.40	0.40 5.30	-0.75 9.00	0.53 26.75	0.626
NMDS2 MI	20	-0.15	0.40	-0.76	0.55	20 20	0.15	0.46	-0.78	0.79	0.820
number of taxa Mi abundance	20	10.82	3.89	4.75	17.00	20	14.01	7.11	4.05	26.50	<u>u.426</u>
log (x+1) chi a µg/cm² log (x+1) % cover algae	20 20	0.36 1.02	0.17	0.05 0.06	0.69 1.91	20 20	0.46 1.07	0.23	0.12	0.83 1.91	0.657
log density MI [ind/m²] ecosystem processes	20	1.82	0.36	1.13	2.34	20	1.98	0.50	0.97	2.71	0.255
log (x+1) % cyanobacteria with heterocsts log (x+1) number of grazers / m2	20 20	0.27 1.38	0.58 0.43	0.00 0.47	1.90 1.95	20 20	0.19 1.42	0.26 0.54	0.00 0.44	0.77 2.28	0.550 0.780
log (x+1) number of shredders / m2 log (x+1) number of filter feeders / m2	20 20	0.61 0.63	0.38 0.32	0.06 0.15	1.31 1.20	20 20	0.67 0.92	0.36 0.59	0.14 0.02	1.40 2.17	0.613 0.059
log (x+1) number of gatherers/collectors / m2 indices	20	1.00	0.58	0.12	1.90	20	0.82	0.58	0.04	1.91	0.158
AIP Raddum 2	18 20	6.66 2.10	0.41 1.13	5.66 0.50	7.13 4.00	19 20	6.60 1.71	0.48 1.01	5.78 0.50	7.17 3.66	0.708 0.267
PIT ASPT	18 20	6.37 6.05	1.20	4.67 5.12	9.32	20 20	7.43 5.89	4.39 0.74	4.56	18.02 6.97	1.000 0.428
	20	7 9 2	0.50	6.08	8 59	20	7.61	0.60	6 76	872	0.545

- 904 **Table 2.** Correlations (Pearson r) among flow regime (calculated as principal components
- 905 (PC_{hydr}) from 78 hydrological variables) and other explanatory variables as well as response
- variables, separately for regulated and unregulated sites. Variables were transformed as
- 907 described in Table 1. Significant (Pearson; p<0.05) correlations with PC axes are marked in
- 908 bold. Note that PC_{hydr} axes were calculated from the pooled dataset, and the results were later

- 909 separated into regulated and unregulated sites. This was done in order to ensure that
- 910 characterization of flow regime was comparable between regulated and unregulated sites.

	unreg	ulated	regu	ated	
	PC1 hydr	PC2 hydr	PC1 hydr	PC2 hydr	
explanatory variables other than flow regime					
longitude (east; UTM 32)	-0.211	-0.516	-0.125	-0.313	
latitude (north; UTM 32)	-0.663	-0.876	-0.495	-0.434	
dist. to lake/reservoir upstream	-0.498	-0.508	-0.390	-0.123	
catchment size	-0.496	-0.507	-0.206	-0.585	
altitude (m asl)	-0.412	-0.644	-0.197	-0.295	
Shading (%)	0.302	0.499	-0.420	0.372	
Tot-P/L [µg P/I]	-0.483	-0.262	-0.031	0.060	
Tot-N/L [µg N/I]	0.662	0.792	0.075	0.331	
TOC [mg C/l]	0.478	0.565	0.256	0.401	
Ca [mg/l]	-0.419	-0.358	-0.409	-0.016	
conductivity (μs/cm)	-0.146	-0.039	-0.323	0.061	
temperature (degree C)	0.531	0.631	0.460	0.216	
рН	-0.606	-0.574	-0.436	0.057	
% turbulent flow	-0.147	-0.287	-0.222	0.490	
average depth (m)	-0.106	-0.083	0.326	-0.209	
width (m)	-0.185	-0.215	0.217	-0.602	
sediment PC1	0.045	0.144	0.119	0.268	
sediment PC2	-0.333	-0.378	0.397	-0.208	
response variables					
species assemblages					
NMDS1 algae	-0.352	-0.409	0.024	-0.018	
NMDS2 algae	0.258	0.149	0.252	-0.093	
number of taxa algae	0.232	0.099	0.114	0.073	
NMDS1 MI	0.571	0.475	0.536	0.075	
NMDS2 MI	0.322	0.463	0.356	-0.061	
number of taxa MI	0.326	-0.118	0.301	0.000	
abundance/biomass					
ChI a µg/cm²	0.558	0.372	0.199	0.403	
% cover algae	0.561	0.458	0.211	0.287	
density MI [ind/m ²]	-0.050	-0.495	0.182	0.024	
ecosystem functions					
% cyanobacteria with heterocsts	0.597	0.617	0.219	0.178	
number of grazers / m2	-0.214	-0.538	-0.006	0.069	
number of shredders / m2	0.285	-0.001	0.044	-0.208	
number of filter feeders / m2	0.177	-0.206	0.443	-0.092	
number of gatherers/collectors / m2	-0.455	-0.741	-0.321	-0.200	
ecosystem assessment					
AIP	-0.603	-0.641	-0.392	-0.021	
Raddum 2	-0.527	-0.599	-0.236	-0.090	
PIT	-0.345	-0.518	-0.099	0.174	
ASPT	-0.141	-0.459	0.105	-0.162	
LIFE	-0.616	-0.680	-0.607	-0.177	

913 **Table 3.** Multivariate linear models for NMDS1.MI and LIFE (interactions were tested, but

914 not significant)

formula	Adjust	ed R2	F-statistic	q
NMDS1.MI = -1.06 + 0.09*temperature + 0.12*PChydr1	0.55	545	25.28 on 2 and 37 DF	1.20E-07
Analysis of Variance	sum of squares	mean squares	F value	Р
temperature	3.9305	3.93	43.9016	8.98E-08
PC1hydr	0.5953	0.60	6.6496	0.01403
Residuals	3.3126	0.09		
formula	Adjust	ed R2	F-statistic	р
LIFE = -8.05 + 2.360e-06*latitude - 0.146*PC1hydr	0.61	151	32.17 on 2 and 37 DF	8.05E-09
Analysis of Variance	sum of squares	mean squares	F value	Р
latitude	7.1541	7.15	57.7778	4.55E-09
PC1hydr	0.8114	0.81	6.5529	0.01469
Residuals	4.5814	0.12		

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- 917 **Table 4.** Summary of correlation matrix between 97 explanatory variables, and the response
- variables; only strong correlations (Pearson r > 0.5 or < -0.5) are listed; + indicates positive, -

negative correlations, q = discharge, MI = macroinvertebrates, CPOM = coarse particulate

- 920 organic matter, FPOM = fine particulate organic matter; TP = total phosphorus; PIT and
- 921 ASPT indices were excluded from this analysis because there occurred too few
- 922 eutrophic/polluted sites, which prevented a meaningful interpretation of the results.

	response	calculated as	best explained by	interpretation
	algal species assemblage	NMDS 1 and 2	Ca, conductivity	algal species assemblages were mainly related to water hardness/alkalinity
ages	algal species richness	number of algal taxa	no correlations with r>0.5, but weak correlations with TP (-) and maximum June q (-) $% \left({{\rm{P}}} \right)$	high water TP and high maximum June discharges (i.e. 3 months before sampling) slightly decreased algal taxon richness
assembla	MI species assemblage	NMDS 1 and 2	longitude, latitude, TN, TOC, temperature, pH, average discharges (particularly in winter and June/July/August), min q June and November	the strongest correlations occurred with temperature and TOC, and most other variables were correlated with these; it is therefore likely that macroinvertebrate species assemblages were mainly affected by temperature and TOC
	MI species richness	number of MI taxa	5- and 25 percentile of discharge (+), 7-day min q (+), min q May and November (+)	strongest relation with May and November minimum discharges during the year before sampling; fewer macroinvertebrate species occurred in streams with low May and November minimum discharges
ass	epilithic algal biomass	epilithic Chl a	TOC (+), mean q June (-), mean and max q November (+)	these variables were correlated with each other, and their relative importance for algal biomass is uncertain; TOC may be beneficial because many algae may use organic P via phosphatase; high summer discharges may have a negative effect due to scouring, high winter discharges (before snow falls) may be beneficial because they prevent freezing damage
biom	total algal cover	% algal cover	temperature (+), June and July discharges (-)	these variables were correlated with each other; high summer discharges may have a negative effect due to scouring, temperature may increase algal growth
	MI density	number of MI individuals /m²	no correlations with r>0.5, but weak correlations with October and November minimum q (+)	high autumn discharges slightly increased MI abundance; this may be due to less drying of the river bed
	potential N-fixation	% cover of cyanobacteria having heterocysts	Julian day of max q (+), month with highest q (+), max q January (+)	autocorrelations occurred with geographic location, temperature and pH, which in turn may be related to nitrogen deposition and nitrogen cycling; we were unable to find arguments for causal relationships between river flow and the abundance of cyanobacteria having heterocysts
Inctions	grazing	number of grazers /m²	longitude (+)	there also occurred a weaker but significant relation with pH (which was correlated with longitude); many grazers, e.g. snails, tend to be acid sensitive; we were unable to separate a possible effect of longitude from an effect of pH; we have no evidence for an effect of river flow characteristics
em fu	degradation of CPOM	number of shredders /m ²	no correlations with r>0.5	the number of shredders was no straightforward response to any of the measured variables; we have no evidence for an effect of river flow characteristics
cosyst	degradation of FPOM	number of collectors /m ²	longitude (+), latitude (+), pH (+), mean q January, March and November (-), max q January and November (-), mean q June (+)	strongest relation occurred with latitude, then with longitude and pH; climate (which changes with geographic position) and pH seemed to be more influential than river flow characteristics
W	filtering	number of filter feeders /m²	25 percentile discharge (+)	there occurred several autocorrelations; the number of filter feeders was also negatively related to the distance between the sampling site and the nearest lake/reservoir upstream; a higher number of filter feeders at lake outlets is well-known; flow minima may reduce the number of filter feeders, but more data are needed before confident conclusions may be drawn;
ŧ	acid sensitivity of algal assemblage	AIP	longitude , latitude, Ca, conductivity, pH (all +), mean winter discharges (-), mean June discharge (+)	strongest relation occurred with pH, all other variables were correlated with pH; most likely pH was causal; we have no evidence for an effect of river flow characteristics
sessmer	acid sensitivity of MI assemblage	Raddum 2	longitude , latitude, pH (all +), winter discharge (-), may discharge (+)	strongest relation occurred with pH, all other variables were correlated with pH; most likely pH was causal; we have no evidence for an effect of river flow characteristics
as	flow preference of MI assemblage	LIFE	latitude (+), pH (+), summer discharges (+), TOC (-), temp (-), winter discharges (-)	strongest relation occurred with latitude; latitude was correlated with overall flow regime (PC1 hydrology, Table 3); most likely, overall flow regime was influential

924 Appendix

Table A.1. List of sampling sites.

NVE number	name	regu- lated since	east (UTM32)	north (UTM32)	average discharge (m ³) (Sept. 2008 - Aug. 2013)
2.129	Dølplass	1916	575519	6896441	24.79
2.267	Mistra Bru		618518	6844041	13.31
2.268	Akslen		471000	6852350	26.50
2.303	Dombås		505319	6883891	10.29
2.32	Atnasjø		564319	6858291	11.38
2.434	Ofossen	1979	463919	6861292	55.79
2.439	Kvarstadseter		601818	6784141	9.19
2.479	Li Bru		552376	6875695	4.04
2.592	Fokstua		515128	6886690	0.71
2.611	Storsjøen ndfØra	1940	628518	6803191	99.37
6.1	Gryta		600551	6651559	0.15
6.9	Maridalsvatn ndf.	1956	599750	6649300	3.22
8.2	Bjørnegårdsvingen	1968	584400	6640500	3.90
12.137	Gjærdeslåtten	1957	485118	6739392	23.62
12.2	Kolbjørnshus	1988	558318	6743592	24.04
12.207	Vinde-elv		504069	6779692	5.77
12.7	Etna		533918	6757592	11.35
12.8	Grønvold bru	1988	558918	6759891	8.68
16.1	Omnesfoss	1958	499618	6608170	24.46
16.128	Austbygdåi		490345	6650892	9.34
16.132	Gjuvå		488518	6624192	1.18
16.155	Sønnlandsvatn	1986	492020	6618490	4.32
16.193	Hørte		507618	6588192	4.77
16.51	Hagadrag	1944	492895	6588165	23.81
19.72	Jørundland	1963	456850	6528550	12.01
20.2	Austenå		448084	6522544	10.26
21.21	Hoslemo	1918	409604	6589839	5.59
25.6	Homstølvatn ndf.	1925	380400	6507550	1.08
27.13	Maudal	1942	347768	6516793	4.44
27.15	Austrumdal		339468	6507943	5.55
27.16	Bjordal		354718	6507793	10.65
30.8	Øvstabøstøl	1986	360100	6527850	1.38
35.2	Hauge bru	1981	354868	6579542	5.72
36.31	Kvilldal	1985	365918	6598992	0.79
36.32	Lauvastøl		370168	6598600	1.92
50.11	Høel	1968	404069	6699542	6.83
50.13	Bioreio		411569	6695392	10.68
109.2	Grensehølen	1973	508200	6937900	29.30
109.21	Svoni		528519	6902891	3.39
109.9	Risefoss		530519	6931291	17.99

Importance of components	PC1	PC2	PC3
Eigenvalue	2.472	1.958	1.205
Proportion Explained	0.309	0.245	0.151
Cumulative Proportion	0.309	0.554	0.704
PC scores	PC1	PC2	PC3
% bolders (>20cm)	1.458	-0.160	0.063
% cobbles (6-20cm)	-1.039	-0.623	-0.444
log (x+1) % gravel (2-6cm)	-1.264	0.165	-0.007
% fine gravel (2mm-2cm)	-0.233	1.222	0.243
log (x+1) % sand (0.1 mm-2mm)	-0.077	1.275	0.068
log (x+1) sediment cover CPOM (> 1mm)	0.020	0.843	-0.465
log (x+1) sediment cover FPOM (<1mm)	0.221	-0.029	1.094
log (x+1) % cover bryophytes	0.738	0.226	-0.992

Table A.2. PCA for sediment composition, calculated from the averaged values per site;
 significant correlations with PC axes are marked in bold.

Table A.3. Regulated sites differed from unregulated sites in catchment size, altitude, TN and
TOC (Table 2). Altitude, TN and TOC also correlated with each other. In order to enable
unbiased comparisons between regulated and unregulated sites, the values of the response
variables were corrected for these differences. The correction was done based on multivariate
linear models which were computed using the MASS package in R, with forward entering of
variables and model selection based on AIC. All models were significant at p<0.05.

model used for correction of response variable	Adjusted R ²
NMDS1.MI=-0.7968+1.5874*TOC	0.486
n.taxa.MI=7.7134+0.024*TN	0.128
Chla=0.11017+0.59898*TOC	0.311
perc.cover.algae=0.4087+1.27*TOC	0.179
n.collectors=0.4954+0.3588*catchm.size-0.941*TOC	0.191
PIT=3.9167+0.015*TN	0.169
LIFE=8.6196-1.7053*TOC	0.343

Table A.4. Correlation matrix among explanatory and response variables; regulated and 951

unregulated sites were pooled; correlations marked in red were significant (Pearson; p<0.05), 952 correlations additionally shimmered in red were strong (Pearson r > 0.5 or < -0.5). 953

- correlations marked in red are significant at p < 0.6; boxes mark strong correlations (coefficients <0.5 or <-0.5) NMDS1 NMDS2 number (NMDS2 number log (x+1) log (x+1 AIP Raddum PIT site ASPT LIFE 0.548 0.330 0.129 0.092 -0.038 0.031 -0.065 0.136 0.086 0.451 0.399 0.009 0.500
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 0.460

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 -0.411
 -0.032

 -0.311
 -0.187
 -0.083

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 -0.280

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 -0.094
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0.388 0.051 0.388 0.051 0.388 0.051 0.388 0.051 0.388 0.051 0.388 0.051 0.388 0.051 0.388 0.051 0.388 0.051 0.388 0.051 0.388 0.051 0.388 0.051 0.388 0.051 0.055 0.032 0.035 0.035 0.032 0.032 0.032 0.032 0.032 0.035 0.032 0.035 0.032 0.032 0.035 0.032 0.035 0.032 0. -0.149 -0.221 -0.161 -0.280 -0.165 -0.185 -0.185 -0.185 -0.185 -0.291 -0.401 -0.359 -0.401 -0.401 -0.401 -0.401 -0.401 -0.401 -0.492 -0.491 -0.492 -0.491 -0.492 -0.492 -0.495 -0.491 -0.492 -0.292 -0.292 -0.295 -0 0.359 0.515 0.515 0.515 0.251 0.251 0.251 0.251 0.251 0.250 0.250 0.250 0.420 0.421 0.422 0.4210 0.292 0.147 0.082 -0.083 -0.400 -0.138 -0.183 -0.238 0.065 0.050 0.123 -0.331 0.022 -0.096 0.056 -0.139 0.069 0.154 0.177 0.207 0.233 0.054 -0.141 -0.070 -0.329 -0.198 -0.174 0.095 0.032 0.110 0.081 0.095 -0.014 -0.002 0.081 -0.138 -0.066 -0.366 -0.272 -0.176 -0.152 -0.265 -0.148 -0.032 0.123 0.123 0.123 0.027 -0.095 -0.381 -0.327 -0.381 0.227 -0.181 0.210 0.045 -0.210 0.045 0.027 0.203 0.090 -0.061 -0.470 -0.396 -0.116 0.402 0.501 0.411 0.088 -0.029 -0.088 0.045 0.013 -0.181 -0.095 -0.191 0.167 0.035 -0.073 -0.305 0.042 0.056 0.185 0.043 0.071 0.262 -0.223 0.208 0.005 -0.063 -0.040 0.053 0.115 -0.051 0.053 -0.063 0.036 -0.042 0.040 0.078 0.139 0.145 -0.044 -0.170 -0.145 -0.108 0.311 -0.021 0.010 0.198 0.380 0.373 0.130 0.411 0.173 0.082 0.348 0.120 -0.020 -0.300 0.180 0.094 -0.027 0.035 0.106 -0.097 -0.218 -0.176 -0.089 -0.017 0.107 0.014 -0.042 0.053 -0.057 0.083 0.144 -0.025 -0.020 -0.104 0.209 0.058 -0.010 0.024 -0.147 0.551 0.286 0.371 0.228 -0.152 -0.021 -0.059 0.001 -0.049 -0.038 -0.328 0.086 umber of days with high pulses year 1 number of high pulses 1 years before sampling otal number of high pulses in 5 years average duration of high pulses -0.444 -0.313 0.159 -0.352 -0.245 0.202 -0.399 -0.313 0.190 0.207 -0.180 -0.271 0.313 0.201 0.099 -0.316 -0.324 0.094 0.158 0.014 -0.030 average duration of high pulses go (x+1) maximum insing limb relative to average discharge 5 years - log (sqrt of quadrat of minimum failing limb relative to average discharge 5 years) average finiting limb relative to average discharge 5 years moves of the limb relative to average discharge 5 years moves of the limb relative to average discharge 5 years -0.152 0.043 -0.022 0.222 -0.177 -0.279 -0.276 0.159 0.228 -0.200 -0.363 0.404 0.342 0.279 0.158 -0.141 0.135 -0.193 0.195 0.136 0.147 -0.259 0.295 -0.074 0.026 0.038 0.019 -0.121 -0.212 0.141 0.173 -0.229 -0.144 -0.062 -0.005 -0.157 0.207 -0.250 0.246 0.255 0.158 -0.023 -0.058 0.175 -0.031 -0.034 -0.004 -0.070 -0.295 0.244 -0.095 0.037 -0.045 -0.134 0.203 0.007 0.123 0.219 -0.208 -0.206 -0.206 0.151 -0.064 -0.205 0.232 0.271 0.183 0.083 -0.086 0.016 0.023 0.010 -0.041 0.064 -0.004 -0.263 0.320 0.259 0.214 0.202 -0.215 -0.148 0.209 0.217 0.223
- 954

month with highest discharge

BFI all years BFI 1 years before sampling

Shading log (x+1) Tot-P/L [µg P/I] Tot-N/L [µg N/I] log (x+1) TOC [mg C/I] log (x+1) Ca/ICP [mg/I] log conductivity tomporture

mperature

pH % turbulent flow

- 955
- 956
- 957
- 958

- **Table A.5.** Reduction of the 78 hydrological variables into principal components. Variables
- 960 that are strongly related to PC axis 1 and 2 (PC scores >0.6 or <-0.6) are marked.

Eigenvalue Proportion Explained	DC1	
Eigenvalue Proportion Explained	24.45	10.15
Proportion Explained	24.45	19.15
	0.31	0.24
Cumulative Proportion	0.31	0.55
	0.51	0.55
Importance of components	PC3	PC4
Eigenvalue	7.27	5.15
Proportion Explained	0.09	0.07
Cumulative Proportion	0 64	0.71
Cumulative Proportion	0.04	0.71
PC scores	PC1	PC2
mean discharge		
average E vegra		
uveruge 5 years		
January	0.746	0.174
february	0.733	0.008
march	0.685	0.326
	0.005	0.520
april	0.167	0.599
may	-0.552	-0.153
iune	-0.530	-0.528
j	0.519	0 511
July	-0.518	-0.511
august	-0.505	-0.388
september	-0.004	0.211
ortobor	0 202	0 504
octobel	0.352	0.554
november	0.499	0.601
december	0.735	0.307
one vear before sampling		
ionuoni	0.753	0.170
January	0.752	0.179
february	0.733	-0.250
march	0.666	-0.261
april	0 365	0 587
	0.505	0.40
may	-0.556	-0.124
june	-0.556	-0.543
july	-0.165	-0.515
august	-0.401	-0 /00
augusi	-0.401	-0.408
september	-0.028	0.351
october	0.449	0.398
november	0.429	0.609
december	0.423	0.005
aecember	0.688	-0.009
magnitude of extremes		
max	-0.581	0.304
min	0 /12	-0 303
	0.412	-0.393
95 percentile	-0.589	0.366
5 percentile	0.518	-0.358
difference min-max	-0 583	0 306
	0.505	0.000
difference 95-5 percentile	-0.602	0.374
difference 99-1 percentile	-0.627	0.345
75 percentile	0.034	-0.284
25 parcentile	0 505	0.222
zopercentile	0.393	-0.232
average yearly max	-0.558	0.468
coefficient of variation yearly max	-0.456	-0.052
average yearly min	0.456	-0.367
average yearly min	0.450	0.507
coefficient of variation yearly min	-0.065	0.311
7 day max	-0.652	-0.046
7 day min	0.505	-0.408
menthly maximum and year hefere complian	0.505	0.100
montiny muximum one year before sumpling		
january	0.404	0.589
february	0.698	-0.090
march	0.664	-0 241
	0.470	0.000
april	0.178	0.693
may	-0.531	0.015
iune	-0.509	-0.261
iuly	0.252	0.240
july	-0.252	-0.240
august	-0.412	0.279
september	-0.030	0.638
october	0.001	0.633
november .	0.011	0.032
november	U.240	0.725
december	0.297	0.564
monthly minimum one year before sampling	0.602	-0.272
ianuary	0.052	0.272
monthly minimum one year before sampling january foly unary	C	
monthy minimum one year before sampling january february	0.684	-0.276
monthiy minimum one year before sampling january february march	0.684 0.653	-0.276
monthiy minimum one year before sampling january february march april	0.684 0.653 0.581	-0.276 -0.237 -0.242
monthiy minimum one year before sampling january february march april may	0.684 0.653 0.581	-0.276 -0.237 -0.242
monthiy minimum one year before sampling january february march april may	0.684 0.653 0.581 0.264	-0.276 -0.237 -0.242 0.262
monthiy minimum one year before sampling january february march april may june	0.684 0.653 0.581 0.264 -0.444	-0.276 -0.237 -0.242 0.262 -0.565
monthiy minimum one year before sampling january february march april may june july	0.684 0.653 0.581 0.264 -0.444 -0.135	-0.276 -0.237 -0.242 0.262 -0.565 -0.487
monthiy minimum one year before sampling january february march april may june july aurust	0.684 0.653 0.581 0.264 -0.444 -0.135	-0.276 -0.237 -0.242 0.262 -0.565 -0.487
monthiy minimum one year before sampling january february march april may june july august	0.684 0.653 0.581 0.264 -0.444 -0.135 -0.109	-0.276 -0.237 -0.242 0.262 -0.565 -0.487 -0.508
monthly minimum one year before sampling january february march april may june july august september	0.684 0.653 0.581 0.264 -0.444 -0.135 -0.109 -0.031	-0.276 -0.237 -0.242 0.262 -0.565 -0.487 -0.508 -0.308
monthy minimum one year before sampling january february march april may june july august september october	0.684 0.653 0.581 0.264 -0.444 -0.135 -0.109 -0.031 0.515	-0.278 -0.237 -0.242 0.262 -0.565 -0.487 -0.508 -0.308 -0.140
monthly minimum one year before sampling january february march april may june july august september october november	0.684 0.653 0.581 0.264 -0.444 -0.135 -0.109 -0.031 0.515 0.495	-0.278 -0.237 -0.242 0.262 -0.565 -0.487 -0.508 -0.308 -0.308 -0.140 0.006
monthly minimum one year before sampling january february march april may june july august september october november	0.684 0.653 0.581 0.264 -0.444 -0.135 -0.109 -0.031 0.515 0.643	-0.278 -0.237 -0.242 0.262 -0.565 -0.487 -0.508 -0.308 -0.308 -0.140 0.006
monthly minimum one year before sampling january february march april may june july august september october november december	0.684 0.653 0.581 0.264 -0.444 -0.135 -0.109 -0.031 0.515 0.495 0.647	-0.278 -0.237 -0.242 0.262 -0.565 -0.487 -0.508 -0.308 -0.140 0.006 -0.276
monthly minimum one year before sampling january february march april may june july august september october november december timing of extremes	0.684 0.653 0.581 0.264 -0.444 -0.135 -0.109 -0.031 0.515 0.495 0.647	-0.278 -0.237 -0.242 0.262 -0.565 -0.487 -0.508 -0.308 -0.140 0.006 -0.276
monthly minimum one year before sampling january february march april may june july august september october november december timing of extremes Julian day of max 1 year before sampling	0.684 0.653 0.581 0.264 -0.444 -0.135 -0.109 -0.031 0.515 0.495 0.647 0.316	-0.278 -0.237 -0.242 0.262 -0.565 -0.487 -0.508 -0.308 -0.308 -0.140 0.006 -0.276
monthly minimum one year before sampling january february march april may june july august september october november december timing of extremes Julian day of max 1 year before sampling days between samoling and last maximum	0.684 0.653 0.581 0.264 -0.444 -0.135 -0.109 -0.031 0.515 0.495 0.647 0.316 0.233	-0.276 -0.237 -0.242 0.262 -0.565 -0.487 -0.508 -0.308 -0.140 0.006 -0.276
monthly minimum one year before sampling january february march april may june july august september october november december timing of extremes Julian day of max 1 year before sampling days between sampling and last maximum biliae dura fei artic una bif func article	0.684 0.653 0.581 0.264 -0.444 -0.135 -0.109 -0.031 0.515 0.495 0.647 0.316 0.233 0.232	-0.276 -0.237 -0.242 -0.262 -0.365 -0.487 -0.508 -0.308 -0.140 0.006 -0.276
monthly minimum one year before sampling january february march april may june july august september october november december timing of extremes Julian day of max 1 year before sampling days between sampling and last maximum Julian day of min 1 year before sampling	0.684 0.653 0.581 0.264 -0.444 -0.135 -0.109 -0.031 0.515 0.495 0.647 0.316 0.233 0.238	-0.276 -0.277 -0.242 -0.262 -0.565 -0.487 -0.508 -0.308 -0.308 -0.140 0.006 -0.276 0.351 0.448 -0.091
monthly minimum one year before sampling january february march april may june july august september october november december timing of extremes Julian day of max 1 year before sampling days between sampling and last maximum Julian day of min 1 year before sampling average Julian day maximum	0.684 0.653 0.561 0.264 -0.444 -0.135 -0.109 -0.031 0.515 0.495 0.647 0.316 0.233 0.238 0.011	-0.276 -0.242 -0.262 -0.565 -0.487 -0.508 -0.308 -0.140 0.006 -0.276 0.351 0.448 -0.091 -0.138
monthly minimum one year before sampling january february march april may june july august september october november december timing of extremes Julian day of mix 1 year before sampling days between sampling and last maximum Julian day of min 1 year before sampling average Julian day maximum average Julian day minimum	0.684 0.653 0.581 0.264 -0.444 -0.135 -0.109 -0.031 0.515 0.495 0.647 0.316 0.233 0.238 0.011 0.231	-0.276 -0.242 -0.262 -0.565 -0.487 -0.508 -0.508 -0.508 -0.508 -0.508 -0.508 -0.508 -0.508 -0.508 -0.276 0.351 0.448 -0.0448 -0.0351 -0.138
monthly minimum one year before sampling january february march april may june july august september october november december timing of extremes Julian day of max 1 year before sampling days between sampling and last maximum Julian day of min 1 year before sampling average Julian day minimum average Julian day minimum mooth with bidheet direcharge	0.684 0.653 0.581 0.264 -0.444 -0.135 -0.109 -0.031 0.515 0.495 0.495 0.495 0.647 0.316 0.233 0.238 0.011 0.291	-0.276 -0.242 -0.262 -0.565 -0.487 -0.508 -0.308 -0.140 0.006 -0.276 0.351 0.448 -0.091 -0.138 -0.138
monthly minimum one year before sampling january february march april may june july august september october november december timing of extremes Julian day of min 1 year before sampling days between sampling and last maximum Julian day of min 1 year before sampling average Julian day minimum average Julian day minimum month with highest discharge	0.684 0.653 0.581 0.264 -0.444 -0.135 -0.109 -0.031 0.515 0.495 0.647 0.316 0.233 0.238 0.011 0.291 -0.203	-0.27/6 -0.237 -0.242 -0.262 -0.365 -0.487 -0.508 -0.308 -0.140 -0.276 -0.351 -0.448 -0.091 -0.138 -0.165 -0.237
monthly minimum one year before sampling january february march april may june july august september october november december timing of extremes Julian day of max 1 year before sampling days between sampling and last maximum Julian day of min 1 year before sampling average Julian day minimum average Julian day minimum month with highest discharge frequency and duration of high pulses (high pulse is > 0.9 percentile)	0.684 0.653 0.581 0.264 -0.444 -0.135 -0.109 -0.031 0.515 0.495 0.495 0.495 0.431 0.233 0.238 0.011 0.291 -0.203	-0.217 -0.242 -0.262 -0.562 -0.487 -0.508 -0.308 -0.140 0.006 -0.276 -0.351 0.448 -0.091 -0.138 -0.165 0.237
monthly minimum one year before sampling january february march april may june july august september october november december timing of extremes Julian day of max 1 year before sampling days between sampling and last maximum Julian day of max 1 year before sampling average Julian day minimum average Julian day minimum month with high est discharge frequency and duration of high pulses (high pulse is > 0.9 percentile) number of days with high outses 1 year before sampling	0.684 0.653 0.581 0.264 -0.444 -0.135 -0.109 -0.031 0.515 0.495 0.647 0.316 0.233 0.238 0.011 0.291 -0.203 -0.099	-0.27/6 -0.237 -0.242 0.262 -0.565 -0.487 -0.508 -0.308 -0.308 -0.308 -0.308 -0.308 -0.308 -0.276 0.351 0.448 -0.351 0.448 -0.351 0.351 0.351 0.337
monthy minimum one year before sampling january february march april may june july august september october november december timing of extremes Julian day of min 1 year before sampling days between sampling and last maximum Julian day of min 1 year before sampling days between sampling and last maximum Julian day of min 1 year before sampling days between sampling and last maximum Julian day of min 1 year before sampling for extremes frequency and duration of hing hulses (high pulses is > 0.9 percentile) number of days with high pulses (high pulses is year before sampling pumber of binb muleon 1 unco high and programming	0.684 0.653 0.581 0.264 -0.464 -0.135 -0.109 -0.031 0.515 0.495 0.647 0.316 0.233 0.238 0.011 0.291 -0.203 -0.099	-0.276 -0.242 -0.262 -0.562 -0.487 -0.508 -0.308 -0.140 -0.308 -0.140 -0.276 -0.276 -0.351 -0.448 -0.091 -0.138 -0.165 -0.237
monthly minimum one year before sampling january february march april may june july august september october november december timing of extremes Julian day of max 1 year before sampling days between sampling and last maximum Julian day of max 1 year before sampling average Julian day minimum month with high pulses fligh pulse is > 0.9 percentile) number of days with high pulses 1 year before sampling number of high pulses 1 year before sampling number of high pulses 1 year before sampling	0.684 0.653 0.581 0.264 -0.444 -0.135 -0.109 -0.031 0.515 0.495 0.647 0.316 0.233 0.238 0.011 0.291 -0.203 -0.099 0.133	-0.278 -0.237 -0.242 0.262 -0.487 -0.487 -0.508 -0.308 -0.308 -0.308 -0.306 -0.276 0.351 0.448 -0.091 -0.135 0.237
monthy minimum one year before sampling january february march april may june july august september october november december timing of extremes Julian day of min 1 year before sampling days between sampling and last maximum Julian day of min 1 year before sampling average Julian day maximum average Julian day maximum average Julian day maximum month with high est discharge frequency and duration of high pulses in \$> 0.9 percentile) number of days with high pulses 1 year before sampling total number of high pulses 1 year before sampling total number of high pulses 1 year before sampling total number of high pulses 1 year before sampling	0.684 0.653 0.581 0.264 -0.464 -0.135 -0.109 -0.031 0.515 0.495 0.647 0.316 0.233 0.238 0.011 0.291 -0.203 -0.203 -0.033 -0.047	-0.276 -0.227 -0.262 -0.265 -0.487 -0.508 -0.308 -0.308 -0.140 -0.276 -0.351 0.448 -0.091 -0.138 -0.138 -0.165 0.237 0.267 0.567
monthly minimum one year before sampling january february march april may june july august september october november december timing of extremes Julian day of max 1 year before sampling days between sampling and last maximum Julian day of max 1 year before sampling average Julian day minimum month with high est before sampling average Julian day minimum month with high pulses (high pulse is > 0.9 percentile) number of days with high pulses 1 year before sampling number of high pulses 1 year before sampling	0.684 0.684 0.581 0.264 -0.444 -0.135 -0.109 -0.031 0.515 0.495 0.495 0.495 0.495 0.495 0.495 0.495 0.495 0.495 0.495 0.233 0.238 0.011 0.291 -0.203 -0.099 0.133 -0.047 0.036	-0.278 -0.227 -0.262 -0.487 -0.487 -0.508 -0.308 -0.308 -0.300 -0.276 0.351 0.448 -0.091 -0.276 0.351 0.448 -0.091 -0.138 0.237 0.267 0.591 0.591 0.591
monthy minimum one year before sampling january february march april may june july august september october november december timing of extremes Julian day of min 1 year before sampling days between sampling and last maximum Julian day of min 1 year before sampling average Julian day maximum average Julian day minimum month with highe st discharge <i>frequency and duration of high pulses</i> in <i>Sears</i> number of days with high pulses in <i>Sears</i> total number of high pulses in <i>Sears</i> average duration of high pulses (days)	0.684 0.684 0.581 0.264 -0.444 -0.135 -0.109 -0.031 0.515 0.495 0.647 0.316 0.233 0.238 0.011 0.291 -0.203 -0.099 0.133 -0.047 0.036	-0.276 -0.237 -0.242 -0.565 -0.565 -0.487 -0.508 -0.308 -0.308 -0.308 -0.308 -0.308 -0.308 -0.308 -0.308 -0.3091 -0.351 -0.438 -0.091 -0.138 -0.655 -0.557 -0.567 -0.418
monthly minimum one year before sampling january february march april may june july august september october november december timing of extremes Julian day of max 1 year before sampling days between sampling and last maximum Julian day of max 1 year before sampling average Julian day maximum average Julian day minimum month with highest discharge frequency and duration of high pulses i year before sampling number of days with high pulses 1 year before sampling number of high pulses 1 year before sampling number of high pulses 1 year before sampling total number of high pulses in 5 years average duration of high pulses (days) rate of change	0.684 0.684 0.581 0.264 -0.444 -0.135 -0.109 -0.031 0.515 0.495 0.647 0.345 0.233 0.238 0.011 0.291 -0.203 -0.099 0.133 -0.047 0.036	-0.27/8 -0.237 -0.242 -0.565 -0.487 -0.508 -0.308 -0.140 0.006 -0.276 0.351 0.448 -0.091 -0.135 0.237 0.267 0.2591 0.567 -0.418
monthy minimum one year before sampling january february march april may june july august september october november december timing of extremes Julian day of min 1 year before sampling days between sampling and last maximum Julian day of min 1 year before sampling average Julian day maximum average Julian day minimum month with highest discharge frequency and duration of high pulses in 5 × 0.9 percentile) number of days with high pulses 1 year before sampling total number of high pulses 1 year set of change maximum rising limb	0.684 0.684 0.581 0.264 -0.444 -0.135 0.031 0.515 0.647 0.316 0.233 0.238 0.011 0.291 -0.203 -0.099 0.133 -0.047 0.036 -0.550	-0.276 -0.237 -0.242 -0.565 -0.487 -0.508 -0.308 -0.308 -0.308 -0.308 -0.308 -0.308 -0.308 -0.308 -0.308 -0.308 -0.276 -0.276 -0.276 -0.2351 -0.438 -0.591 -0.567 -0.418 -0.433
monthly minimum one year before sampling january february march april may june july august september october november december timing of extremes Julian day of max 1 year before sampling days between sampling and last maximum Julian day of max 1 year before sampling average Julian day maximum average Julian day maximum average Julian day minimum month with highest discharge frequency and duration of high pulses i year before sampling number of days with high pulses 1 year before sampling number of high pulses 1 year before sampling number of high pulses 1 year before sampling total number of high pulses in 5 years average duration of high pulses (adys) rate of change maximum rising limb	0.684 0.684 0.581 0.264 -0.444 -0.135 -0.109 -0.031 0.515 0.495 0.647 0.316 0.233 0.238 0.011 0.293 -0.203 -0.099 0.133 -0.047 0.047 0.047 0.047	-0.27/6 -0.237 -0.242 -0.565 -0.487 -0.508 -0.308 -0.140 0.006 -0.276 0.351 0.448 -0.091 -0.135 0.237 0.267 0.291 0.567 -0.418 0.433 -0.439
monthy minimum one year before sampling january february march april may june july august september october november december december timing of extremes Julian day of min 1 year before sampling days between sampling and last maximum Julian day of min 1 year before sampling average Julian day maximum average Julian day minimum month with highest discharge frequency and duration of high pulses in 5 × 0.9 percentile) number of days with high pulses 1 year before sampling total number of high pulses (days) rate of change maximum rising limb minimum falling limb average tising limb	0.684 0.684 0.581 0.264 -0.444 -0.135 0.031 0.515 0.495 0.647 0.316 0.233 0.238 0.011 0.291 -0.039 0.037 0.036 -0.0550 0.552 0.5550	-0.277 -0.242 -0.565 -0.487 -0.508 -0.308 -0.308 -0.308 -0.308 -0.308 -0.308 -0.308 -0.308 -0.308 -0.308 -0.308 -0.308 -0.308 -0.308 -0.308 -0.309 -0.351 0.448 -0.091 -0.165 0.237 -0.267 0.567 -0.418 -0.567 -0.418 -0.567 -0.418 -0.567 -0.418 -0.567 -0.418 -0.567 -0.418 -0.555 -0.427 -0.555 -0.427 -0.555 -0.487 -0.140 -0.555 -0.487 -0.555 -0.487 -0.555 -0.487 -0.555 -0.487 -0.557 -0.557 -0.488 -0.557 -0.557 -0.488 -0.557 -0.557 -0.488 -0.557 -0.557 -0.448 -0.557 -0.557 -0.448 -0.557 -0.557 -0.448 -0.557 -0.557 -0.448 -0.557 -0.557 -0.448 -0.557 -0.577 -0.448 -0.557 -0.448 -0.557 -0.577 -0.448 -0.577 -0.448 -0.577 -0.448 -0.577 -0.448 -0.577 -0.448 -0.577 -0.448 -0.433 -0.777 -0.777 -0.777 -0.777 -0.777 -0.418 -0.777 -0.418 -0.777 -0.418 -0.777 -0.418 -0.777 -0.418 -0.777 -0.777 -0.418 -0.777 -0.777 -0.777 -0.418 -0.7777 -0.777 -0.777 -0.7777 -0.7777 -0.7777 -0.7777 -0.7777 -0.7777 -0.7777 -0.77777 -0.77777 -0.77777 -0.7777777777
monthly minimum one year before sampling january february march april may june july august september october november december timing of extremes Julian day of max 1 year before sampling days between sampling and last maximum Julian day of max 1 year before sampling average Julian day maximum average Julian day minimum month with high pulses in for a precentile) number of days with high pulses 1 year before sampling number of high pulses 1 year before sampling maximum rising limb average fising limb average fising limb	0.684 0.653 0.581 0.264 -0.444 -0.135 -0.109 -0.031 0.515 0.495 0.495 0.495 0.495 0.495 0.495 0.495 0.495 0.495 0.495 0.495 0.495 0.495 0.495 0.233 0.238 0.011 0.291 -0.203 -0.099 0.133 -0.047 0.047 0.047 0.0550 0.552 -0.550	-0.27/6 -0.237 -0.242 -0.565 -0.487 -0.508 -0.487 -0.508 -0.140 -0.276 -0.276 -0.276 -0.275 -0.448 -0.091 -0.135 -0.448 -0.091 -0.165 -0.237 -0.267 -0.267 -0.591 -0.591 -0.567 -0.413 -0.433 -0.439 -0.427
monthy minimum one year before sampling january february march april may june july august september october november december timing of extremes Julian day of min 1 year before sampling days between sampling and last maximum Julian day of min 1 year before sampling average Julian day maximum average Julian day maximum month with highest discharge frequency and duration of high pulses in 5 × 0.9 percentile) number of days with high pulses 1 year before sampling total number of high pulses 1 years average duration of high pulses (days) rate of change maximum rising limb minimum falling limb average falling limb	0.684 0.684 0.581 0.264 -0.444 -0.135 0.031 0.515 0.495 0.647 0.316 0.233 0.238 0.011 0.291 -0.039 -0.099 0.133 -0.047 0.036 -0.550 0.542 -0.550 0.202	0.277 0.242 0.265 0.565 -0.487 -0.508 -0.308 -0.308 -0.308 -0.308 -0.308 -0.308 -0.308 -0.351 0.448 -0.091 -0.351 0.448 -0.165 0.237 0.267 0.567 -0.418 0.433 -0.433 -0.439 0.727 -0.598
monthly minimum one year before sampling january february march april may june july august september october november december timing of extremes Julian day of max 1 year before sampling days between sampling and last maximum Julian day of max 1 year before sampling ayerage Julian day minimum month with high pulses ingh pulse is > 0.9 percentile) frequency and duration of high pulses (high pulse is > 0.9 percentile) number of days with high pulses 1 year before sampling number of high pulses 1 year before sampling frequency and duration of high pulses (days) rate of change maximum rising limb average fising limb average falling limb base flow index	0.684 0.653 0.581 0.264 -0.444 -0.135 -0.109 -0.031 0.515 0.495 0.495 0.495 0.495 0.495 0.495 0.495 0.495 0.495 0.495 0.495 0.495 0.495 0.495 0.233 0.238 0.011 0.291 -0.203 -0.099 0.133 -0.047 0.036 0.550 0.552 -0.270 0.202	-0.27/6 -0.237 -0.242 -0.565 -0.487 -0.508 -0.487 -0.508 -0.140 -0.0276 -0.276 -0.276 -0.275 -0.448 -0.091 -0.135 -0.448 -0.091 -0.165 -0.237 -0.591 -0.591 -0.591 -0.433 -0.439 -0.439 -0.439 -0.277 -0.698
monthly minimum one year before sampling january february march april may june july august september october november december timing of extremes Julian day of min 1 year before sampling days between sampling and last maximum Julian day of min 1 year before sampling average Julian day maximum average Julian day maximum month with highest discharge frequency and duration of high pulses in 5 × 0.9 percentile) number of days with high pulses 1 year before sampling total number of high pulses 1 years average duration of high pulses (days) rate of change maximum rising limb minimum falling limb average falling limb average falling limb Bese flow index BET 5 vears	0.684 0.684 0.581 0.264 -0.444 -0.135 0.515 0.515 0.647 0.316 0.233 0.238 0.011 0.291 -0.203 -0.099 0.133 -0.047 0.036 -0.550 0.542 -0.252 0.202	-0.276 -0.237 -0.242 -0.565 -0.487 -0.508 -0.308 -0.308 -0.308 -0.308 -0.308 -0.308 -0.308 -0.308 -0.308 -0.276 -0.351 0.448 -0.091 -0.351 0.448 -0.091 -0.165 0.237 -0.267 0.567 -0.418 0.567 -0.418 -0.567 -0.418 -0.567 -0.418 -0.567 -0.418 -0.567 -0.418 -0.565 -0.410 -0.555 -0.427 -0.557 -0.448 -0.557 -0.448 -0.557 -0.448 -0.557 -0.448 -0.557 -0.448 -0.557 -0.448 -0.557 -0.448 -0.557 -0.448 -0.557 -0.448 -0.557 -0.448 -0.557 -0.448 -0.557 -0.448 -0.557 -0.448 -0.557 -0.448 -0.557 -0.448 -0.557 -0.557 -0.455 -0.557
monthly minimum one year before sampling january february march april may june july august september october november december <i>timing of extremes</i> Julian day of max 1 year before sampling days between sampling and last maximum Julian day of max 1 year before sampling days between sampling and last maximum Julian day of max 1 year before sampling days between sampling and last maximum Julian day of max 1 year before sampling days between sampling and last maximum average Julian day minimum month with high pelses (high pulse is > 0.9 percentile) number of days with high pulses 1 year before sampling number of high pulses 1 year before sampling frequency and duration of high pulses (days) rate of change maximum rising limb average failing limb base flow index BFI 5 years DFI 4 uses before accenting	0.684 0.653 0.581 0.264 -0.444 -0.135 -0.109 -0.031 0.515 0.495 0.495 0.495 0.495 0.495 0.495 0.495 0.495 0.495 0.495 0.495 0.495 0.495 0.495 0.495 0.233 -0.291 -0.203 -0.099 0.133 -0.047 0.036 -0.550 0.552 -0.270 0.202 0.201	-0.27/8 -0.242 0.262 -0.565 -0.487 -0.508 -0.308 -0.140 0.006 -0.276 -0.351 0.448 -0.091 -0.138 -0.165 0.237 0.267 0.591 0.567 -0.418 -0.433 -0.443 -0.445 -0.448 -0.455 -0.448 -0.455 -0.557 -0.448 -0.455 -0.448 -0.455 -0.448 -0.455 -0.455 -0.448 -0.455 -

- **Table A.6.** hydrological characteristics at 20 regulated and 20 unregulated sites. Site codes
- 963 refer to Table A.1.

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 | regulat | ed sites

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95 perc. relative to mean (%)	109.2 355.5

 | 12.137
228.8
40.2 | 12.2
330.5

 | 12.8
294.5
9.2
 | 16.1
242.8
21.7

 | 16.155
360.9

 | 16.51
195.0

 | 19.72
191.2 | 2.129
 | 2.434
283.1
12.9 | 2.611

 | 21.21
242.5
27.1 | 25.6
319.3 | 27.13
 | 30.8
345.2 | 35.2
400.6
 | 36.31
295.2
5.1 | 50.11
224.9
11.6 | 6.9
343.3
29.2
 | 8.2
383.8
13.0 |
| max relative to mean (%)
min relative to mean (%) | 2245.3
7.1

 | 915.3
5.7 | 2076.0
10.3

 | 1935.6
0.0
 | 797.2
10.9

 | 1334.7
2.3

 | 654.6
19.7

 | 316.3
5.0 | 2081.3
5.2
 | 674.0
5.9 | 296.4
18.0

 | 1570.7
2.6 | 379.4
0.1 | 254.5
20.0
 | 3076.5
0.0 | 2197.8
0.2
 | 806.7
0.4 | 572.6
3.8 | 932.3
12.4
 | 1565.2
6.0 |
| difference min-max relative to mean (%)
difference 95-5 percentile relative to mean (%)
difference 99-1 percentile relative to mean (%) | 2238.3
345.6
577.3

 | 909.6
188.6
491.9 | 2065.7
317.2
735.9

 | 1935.6
285.3
531.3
 | 786.3
221.1
392.3

 | 1332.5
350.9
702.7

 | 634.9
160.6
333.7

 | 311.3
179.6
205.0 | 2076.1
263.8
600.0
 | 668.0
270.3
408.4 | 278.3
94.7
146.4

 | 1568.1
205.4
451.7 | 379.3
314.7
342.8 | 234.4
140.3
165.7
 | 3076.5
341.1
793.9 | 2197.6
394.7
792.6
 | 806.3
290.1
484.4 | 568.8
213.3
300.2 | 920.0
315.1
716.3
 | 1559.2
370.9
675.3 |
| 75 perc. relative to mean (%)
25 perc. relative to mean (%) | 125.1
20.9

 | 100.0
67.6 | 113.0
22.7

 | 131.7
17.7
 | 111.4
59.3

 | 122.7
26.8

 | 112.1
66.8

 | 167.8
21.8 | 136.3
23.7
 | 158.8
34.1 | 118.0
79.3

 | 107.5
50.2 | 160.4
16.4 | 145.2
58.3
 | 108.0
21.9 | 117.6
15.6
 | 129.8
26.9 | 189.7
22.4 | 112.4
36.7
 | 105.6
25.1 |
| average yearly max relative to mean discharge (%)
average yearly min relative to mean discharge (%) | 1131.7
13.5

 | 604.9
15.8 | 1126.6
11.9

 | 913.1
5.8
 | 637.5
12.9

 | 994.9
9.3

 | 446.2
24.6

 | 275.5
8.2 | 1242.1
13.5
 | 495.4
10.6 | 232.4
39.9

 | 799.0
26.3 | 361.6
1.9 | 204.4
29.4
 | 2035.0
3.8 | 1520.8
3.7
 | 697.2
5.1 | 422.3
8.3 | 620.5
23.0
 | 1161.1
12.4 |
| mean discharge january relative to mean (%) | 20.5

 | 73.4 | 20.2

 | 13.5
 | 84.2

 | 22.7

 | 82.0

 | 166.1 | 19.9
 | 41.5 | 84.9

 | 47.3 | 107.8 | 138.1
 | 49.7 | 81.6
 | 54.7 | 37.3 | 35.1
 | 30.0 |
| mean discharge replicary relative to mean (%)
mean discharge march relative to mean (%) | 27.4 62.4

 | 64.5
58.7 | 23.1
111.7

 | 26.6
125.6
 | 78.5
122.8

 | 39.7
144.0

 | 85.5
102.9

 | 141.2
83.6 | 18.6
107.8
 | 33.6
32.3 | 76.4
78.0

 | 55.8
109.2 | 90.9
51.4 | 110.1
88.4
 | 45.6
145.0 | 87.9
128.2
 | 68.1
122.6 | 37.2
63.1 | 36.9
161.3
 | 50.0
252.1 |
| mean discharge may relative to mean (%)
mean discharge june relative to mean (%)
mean discharge july relative to mean (%) | 219.9
311.0
214.8

 | 178.0
171.3 | 300.5
144.0

 | 260.1
123.2
 | 129.8
75.1

 | 227.4
138.4

 | 127.3
73.2

 | 62.1
45.0 | 237.0
212.4
 | 138.2
239.7
229.7 | 128.5
138.9

 | 224.7
140.2 | 86.8
196.2 | 74.1
62.0
 | 188.6
79.8
77.0 | 104.8
44.5
64.2
 | 159.1
95.3
97.1 | 157.8
208.3
200.9 | 93.2
68.9
81.4
 | 111.0
73.4 |
| mean discharge august relative to mean (%)
mean discharge august relative to mean (%) | 131.5
81.9

 | 114.1
102.6 | 164.7
133.9

 | 153.2
123.3
 | 111.3
105.2

 | 137.4
117.0

 | 124.3
116.4

 | 49.9
63.3 | 163.0
108.2
 | 173.3
115.7 | 118.0
110.6

 | 105.9
108.3 | 140.5
83.9 | 58.4
109.9
 | 100.8
112.1 | 77.1
154.8
 | 107.9
140.4 | 195.0
130.0 | 99.6
253.4
 | 135.6
116.7 |
| mean discharge october relative to mean (%)
mean discharge november relative to mean (%)
mean discharge december relative to mean (%) | 52.7
32.0
20.6

 | 94.6
79.7
78.8 | 75.1
55.3
30.4

 | 85.3
92.6
60.1
 | 101.2
108.3
83.6

 | 101.0
70.6
33.1

 | 106.6
105.5
90.7

 | 118.6
129.3
150.9 | 78.8
47.5
26.5
 | 67.3
42.8
42.3 | 102.9
88.8
72.4

 | 104.0
84.1
50.7 | 33.3
41.2
78.3 | 124.7
137.3
128.1
 | 128.5
140.3
67.2 | 160.3
164.9
85.1
 | 120.2
138.0
61.9 | 62.4
37.6
36.0 | 145.8
145.3
46.7
 | 117.8
154.8
56.5 |
| max discharge january 1 ybs relative to annual mean | 0.17

 | 0.79 | 0.31

 | 0.24
 | 1.09

 | 0.39

 | 0.96

 | 1.91 | 0.11
 | 0.55 | 0.94

 | 0.63 | 1.54 | 1.61
 | 2.63 | 5.10
 | 4.72 | 0.45 | 0.71
 | 0.64 |
| max discharge tebruary 1 ybs relative to annual mean
max discharge march 1 ybs relative to annual mean
max discharge april 1 ybs relative to annual mean | 0.11
0.11
0.94

 | 0.76
0.73
0.62 | 0.19
0.14
0.85

 | 0.12
0.11
2.39
 | 0.95
0.99
1.91

 | 0.26
0.13
2.71

 | 0.99
0.95
1.14

 | 1.90
1.81
1.74 | 0.07
0.08
3.08
 | 0.42
0.44
0.20 | 0.84
0.85
1.14

 | 0.52
0.52
1.53 | 2.06
2.30
2.02 | 1.46
0.89
1.12
 | 0.00
0.01
8.28 | 0.40
0.14
6.73
 | 1.53
0.11
5.08 | 0.28
0.24
2.37 | 0.36
0.42
3.27
 | 0.21
0.15
6.04 |
| max discharge may 1 ybs relative to annual mean
max discharge june 1 ybs relative to annual mean | 9.42
6.06

 | 7.91
4.50 | 20.76

 | 19.36
4.45
 | 7.97

 | 13.35
5.27

 | 5.50
1.81

 | 2.64
1.80 | 20.81
 | 4.66 | 2.38

 | 15.71
3.77 | 0.79 | 1.39
 | 16.69
1.97 | 4.25
0.69
 | 7.51 | 5.73
2.32 | 1.81
6.04
 | 15.65
7.95 |
| max discharge august 1 ybs relative to annual mean
max discharge august 1 ybs relative to annual mean
max discharge september 1 ybs relative to annual mean | 2.95

 | 1.43 | 3.95
1.40

 | 2.03
 | 2.36

 | 3.20

 | 1.65

 | 1.29 | 5.17
 | 2.78 | 1.30

 | 4.27 | 3.09 | 1.09
 | 5.25 | 3.30
 | 5.01 | 2.28 | 0.49
 | 2.54 |
| max discharge october 1 ybs relative to annual mean
max discharge november 1 ybs relative to annual mean
max discharge december 1 ybs relative to annual mean | 0.56
0.45
0.24

 | 1.30
1.03
0.83 | 1.33
1.60
0.66

 | 2.01
2.34
5.29
 | 2.41
2.50
1.05

 | 3.12
3.54
0.55

 | 1.52
1.86
1.24

 | 2.23
2.49
1.91 | 1.00
0.70
0.34
 | 0.43
0.47
0.49 | 1.30
1.55
1.01

 | 2.91
2.52
0.62 | 0.81
0.96
1.42 | 1.59
1.99
1.55
 | 6.69
7.17
2.25 | 2.43
6.47
7.52
 | 2.65
3.77
4.79 | 1.14
1.22
0.39 | 3.84
4.64
0.33
 | 9.98
12.99
0.62 |
| min discharge january 1 ybs relative to average | 0.09

 | 0.70 | 0.18

 | 0.07
 | 0.52

 | 0.18

 | 0.50

 | 1.51 | 0.07
 | 0.37 | 0.83

 | 0.48 | 0.09 | 0.88
 | 0.00 | 0.06
 | 0.02 | 0.25 | 0.25
 | 0.19 |
| min discharge rebruary 1 yos relative to average
min discharge march 1 ybs relative to annual mean
min discharge april 1 ybs relative to annual mean | 0.08

 | 0.42 | 0.14
0.11
0.10

 | 0.10
 | 0.55
0.43
0.13

 | 0.08

 | 0.62
0.39
0.42

 | 1.64
1.56
0.31 | 0.05
 | 0.37
0.10
0.10 | 0.53

 | 0.48 | 1.20 | 0.88
 | 0.00 | 0.06
 | 0.06 | 0.18 | 0.28
0.28
0.42
 | 0.14
0.12
0.17 |
| min discharge may 1 ybs relative to annual mean
min discharge june 1 ybs relative to annual mean | 0.14

 | 0.15 | 0.70

 | 0.69
 | 0.60

 | 1.02

 | 0.85

 | 0.12 | 0.50
 | 0.13 | 1.16

 | 0.46 | 0.11 | 0.35
 | 0.28 | 0.22
 | 0.81 | 0.11 | 0.54
 | 0.69 |
| min discharge july 1 yds relative to annual mean
min discharge august 1 ybs relative to annual mean
min discharge september 1 ybs relative to annual mean | 0.46
0.42

 | 0.38
0.43
0.53 | 0.29
0.30
0.47

 | 0.16
0.14
0.48
 | 0.27
0.22
0.13

 | 0.02
0.04
0.30

 | 0.48

 | 0.11
0.12 | 1.00
0.72
 | 0.96 | 0.80

 | 0.81
0.82 | 0.05 | 0.32
0.32
0.68
 | 0.29
0.25
0.16 | 0.09
0.20
 | 0.74 0.72 0.75 | 1.68
0.27 | 0.38
0.36
0.42
 | 0.13
0.25 |
| min discharge october 1 ybs relative to annual mean
min discharge november 1 ybs relative to annual mean
min discharge december 1 ybs relative to annual mean | 0.27
0.24
0.15

 | 0.34
0.47
0.74 | 0.52
0.51
0.30

 | 0.55
0.65
0.18
 | 0.20
0.71
0.55

 | 0.48
0.62
0.30

 | 0.88
0.87
0.59

 | 1.02
0.49
1.18 | 0.49
0.34
0.22
 | 0.13
0.13
0.37 | 0.87
0.82
0.70

 | 0.49
0.51
0.44 | 0.15 | 1.26
1.27
0.86
 | 0.12
0.10
0.07 | 0.19
0.21
0.08
 | 0.24
0.25
0.12 | 0.17
0.13
0.18 | 0.42
 | 0.58
0.65
0.25 |
| mean discharge january 1 ybs relative to average | 0.13

 | 0.74 | 0.23

 | 0.15
 | 0.87

 | 0.26

 | 0.83

 | 1.74 | 0.09
 | 0.43 | 0.88

 | 0.54 | 0.95 | 1.29
 | 0.38 | 0.86
 | 0.71 | 0.31 | 0.34
 | 0.29 |
| mean discharge tebruary 1 ybs relative to average
mean discharge march 1 ybs relative to average
mean discharge april 1 ybs relative to average | 0.09
0.09
0.27

 | 0.74
0.70
0.35 | 0.16
0.12
0.37

 | 0.12
0.09
0.65
 | 0.81
0.70
0.75

 | 0.13
0.10
0.87

 | 0.93
0.69
0.78

 | 1.78
1.72
1.16 | 0.07
0.06
1.26
 | 0.39
0.33
0.13 | 0.81
0.74
0.59

 | 0.50
0.47
0.63 | 1.20
1.72
1.14 | 1.04
0.80
0.69
 | 0.00
0.00
1.48 | 0.13
0.09
1.41
 | 0.17
0.07
1.08 | 0.26
0.21
0.43 | 0.33
0.33
1.21
 | 0.16
0.13
1.81 |
| mean discharge may 1 ybs relative to average
mean discharge june 1 ybs relative to average | 3.49
2.58

 | 3.20
2.06 | 5.51
2.30

 | 4.38
 | 2.51

 | 4.81
2.52

 | 2.45
1.04

 | 0.79 | 4.21
2.62
 | 2.11 2.36 | 1.47
1.44

 | 4.31 | 0.43 | 0.71
 | 3.25
0.74 | 1.43
0.29
 | 2.35
0.94 | 1.86
1.93 | 1.37
1.59
 | 2.53
1.32 |
| mean discharge july 1 yos relative to average
mean discharge august 1 ybs relative to average
mean discharge september 1 ybs relative to average | 1.09
0.69

 | 0.82
1.02
0.73 | 0.83
1.30
0.68

 | 0.63
 | 0.80

 | 0.64

 | 0.82

 | 0.32 | 1.67
 | 1.85 | 0.97

 | 1.24 | 1.74
0.24 | 0.34
0.36
1.26
 | 0.40
0.98
1.43 | 0.29
0.64
2.26
 | 1.28
1.60 | 1.80 | 0.43
1.35
 | 0.58 |
| mean discharge october 1 ybs relative to average
mean discharge november 1 ybs relative to average
mean discharge december 1 ybs relative to average | 0.38
0.30
0.19

 | 0.78
0.84
0.80 | 0.81
0.88
0.41

 | 0.92
 | 1.17
1.37
0.86

 | 1.16
1.49
0.35

 | 1.17
1.37
0.94

 | 1.74
1.70
1.69 | 0.68
0.57
0.26
 | 0.24
0.25
0.43 | 1.11
1.11
0.77

 | 0.92 | 0.42
0.57
0.83 | 1.33
1.53
1.28
 | 1.19
1.69
0.26 | 0.72 1.78 0.61
 | 0.90 | 0.44
0.42
0.25 | 2.29
3.15
0.31
 | 2.13
2.81
0.36 |
| average Julian day of max
days between sampling and last maximum | 164.9
104.5

 | 146.7
102.5 | 162.1
100.5

 | 180.5
101.5
 | 187.3
107.5

 | 185.1
107.5

 | 204.1
104.5

 | 243.7
107.5 | 158.3
100.5
 | 187.6
102.5 | 180.5
98.5

 | 180.7
103.5 | 229.0
48.5 | 201.5
 | 260.5
105.5 | 240.5
351.5
 | 181.1
103.5 | 135.7
105.5 | 228.5
61.5
 | 250.1
100.5 |
| BFI all years
BFI 1 years before campling | 0.61

 | 0.66 | 0.56

 | 0.53
 | 0.58

 | 0.48

 | 0.74

 | 0.61 | 0.67
 | 0.67 | 0.88

 | 0.68 | 0.40 | 0.81
 | 0.29 | 0.21
 | 0.46 | 0.74 | 0.61
 | 0.40 |
| total number of high pulses in 5 years | 50.0

 | 56.5 | 52.5

 | 69.5
 | 66.5

 | 68.0

 | 34.0

 | 78.0 | 47.0
 | 43.0 | 50.5

 | 95.0 | 36.3 | 53.0
 | 114.0 | 109.0
 | 134.5 | 82.0 | 28.5
 | 75.5 |
| average duration of high pulses
7 day max 5 years relative to mean discharge | 3.7
9.54

 | 3.2
6.73 | 3.5

 | 2.6
8.96
 | 2.8
3.99

 | 2.7
8.33

 | 5.4
4.64

 | 2.1 | 3.9
9.72
 | 4.3
5.60 | 3.6
2.42

 | 1.9 | 5.0
3.45 | 3.5
2.03
 | 1.6
8.72 | 1.7
 | 1.4
4.29 | 2.2
3.97 | 6.4
8.89
 | 2.4
6.26 |
| 7 day min 5 years relative to mean discharge | 0.07

 | 0.18 | 0.10

 | 0.05
 | 0.12

 | 0.03

 | 0.21

 | 0.09 | 0.05
 | 0.06 | 0.24

 | 0.05 | 0.02 | 0.22
 | 0.00 | 0.01
 | 0.03 | 0.05 | 0.15
 | 0.08 |
| minimum falling limb relative to average discharge 5 years
average rising limb relative to average discharge 5 years | -13.00
0.30

 | -4.86
0.19 | -9.91
0.31

 | -11.63
0.29
 | -3.69
0.24

 | -4.15
0.38

 | -1.38
0.09

 | -1.89
0.24 | -13.63
0.28
 | -2.09 | -0.99

 | -6.14
0.25 | -2.42 | -1.00 0.12
 | -28.83
0.77 | -19.41
0.83
 | -5.47
0.47 | -1.96
0.18 | -2.24 0.15
 | -8.80 |
| anara ao kallion linak calakina ka anarana disakaran l'inaara | .0.10

 | -0.15 | -0.16

 | -0.17
 | -0.18

 | -0.21

 | -0.07

 | -0.20 | -0.16
 | -0.13 | -0.04

 | -0.20 | -0.20 | -0.09
 | -0.48 | -0.41
 | -0.32 | -0.12 | -0.14
 | -0.28 |
| average raining inno relative to average discharge 3 years | 0.10

 | |

 |
 |

 |

 |

 | |
 | |

 | | |
 | |
 | | |
 | |
| average raining initial relative to average discharge a years | 109.21

 | 109.9 | 12.207

 | 12.7
 | 16.128

 | 16.132

 | 16.193

 | 2.267 | 2.268
 | unregula
2.303 | 2.32

 | 2.439 | 2.479 | 2.592
 | 20.2 | 27.15
 | 27.16 | 36.32 | 50.13
 | 6.1 |
| average raning immutedure to average unchange a years
95 perc. relative to mean (%)
5 perc. relative to mean (%)
max relative to mean (%) | 109.21
327.6
11.4
1461.7

 | 109.9
385.3
4.7
1863.1 | 12.207
344.8
5.2
1686.7

 | 12.7
354.3
5.7
1681.7
 | 16.128
380.8
6.0
1264.3

 | 16.132
335.6
11.0
1177.1

 | 16.193
382.7
6.2
1560.0

 | 2.267
306.5
12.8
1013.4 | 2.268
364.1
2.8
1365.4
 | unregula
2.303
365.3
8.4
1449.7 | 2.32
260.9
17.0
1191.4

 | 2.439
332.2
9.2
1389.0 | 2.479
279.2
13.7
1354.3 | 2.592
337.0
12.9
1448.0
 | 20.2
327.3
12.4
765.9 | 27.15
324.3
6.9
837.4
 | 27.16
354.5
6.5
1418.4 | 36.32
341.3
3.2
1156.6 | 50.13
459.0
4.4
807.4
 | 6.1
374.2
8.6
1093.3 |
| arenage samily mini resure us arenage uncluing a years
95 perc. relative to mean (N)
perc. relative to mean (N)
min relative to mean (N)
differences (N) security relative (N)
differences (N)
differences (N)
differences (N)
differences (N)
differences (N)
differences (N)
differenc | 109.21
327.6
11.4
1461.7
1.0
1460.7
316.1

 | 109.9
385.3
4.7
1863.1
3.8
1859.3
380.5 | 12.207
344.8
5.2
1686.7
2.7
1683.9
339.6

 | 12.7
354.3
5.7
1681.7
4.2
1677.5
348.7 | 16.128
380.8
6.0
1264.3
4.1
1260.2
374.8

 | 16.132
335.6
11.0
1177.1
7.6
1169.4
324.6

 | 16.193
382.7
6.2
1560.0
3.8
1556.3
376.5

 | 2.267
306.5
12.8
1013.4
9.1
1004.3
293.7 | 2.268
364.1
2.8
1365.4
1.5
1363.9
361.3 | unregula
2.303
365.3
8.4
1449.7
6.1
1443.6
356.9
 | 2.32
260.9
17.0
1191.4
14.0
1177.4
243.9
 | 2.439
332.2
9.2
1389.0
6.0
1383.0
323.0
 | 2.479
279.2
13.7
1354.3
10.2
1344.2
265.5 | 2.592
337.0
12.9
1448.0
10.5
1437.5
324.1 | 20.2
327.3
12.4
765.9
5.1
760.7
314.8
 | 27.15
324.3
6.9
837.4
1.3
836.1
317.4 | 27.16
354.5
6.5
1418.4
4.7
1413.7
348.0
 | 36.32
341.3
3.2
1156.6
0.5
1156.1
338.1 | 50.13
459.0
4.4
807.4
0.0
807.4
454.5
 | 6.1
374.2
8.6
1093.3
4.9
1088.4
365.6 |
| Strenge Lang mini resure us arronge sacuage 3 years
95 perc. relative to mean (%)
5 perc. relative to mean (%)
min relative to mean (%)
difference sind a percentile relative to mean (%)
difference sind a percentile relative to mean (%)
difference in man xelative to mean (%)
difference in man (%)
2 perc. relative to mean (%)
2 perc. relative to mean (%)
2 perc. relative to mean (%) | 109.21
327.6
11.4
1461.7
1.0
1460.7
316.1
595.9
133.4

 | 109.9
385.3
4.7
1863.1
3.8
1859.3
380.5
668.3
137.4 | 12.207
344.8
5.2
1686.7
2.7
1683.9
339.6
808.1
120.7
18.7

 | 12.7
354.3
5.7
1681.7
4.2
1677.5
348.7
804.5
125.1
 | 16.128
380.8
6.0
1264.3
4.1
1260.2
374.8
782.5
115.6
15.2

 | 16.132
335.6
11.0
1177.1
7.6
1169.4
324.6
665.6
120.6
24.1

 | 16.193
382.7
6.2
1560.0
3.8
1556.3
376.5
670.1
119.4

 | 2.267
306.5
12.8
1013.4
9.1
1004.3
293.7
650.8
119.5
20.4 | 2.268
364.1
2.8
1365.4
1.5
1363.9
361.3
560.8
163.2
 | unregula
2.303
365.3
8.4
1449.7
6.1
1443.6
356.9
640.3
142.4
14.6 | 2.32
260.9
17.0
1191.4
14.0
1177.4
243.9
454.4
1454.6
27.7

 | 2.439
332.2
9.2
1389.0
6.0
1383.0
323.0
807.7
117.5 | 2.479
279.2
13.7
1354.3
10.2
1344.2
265.5
493.1
147.0
25.2 | 2.592
337.0
12.9
1448.0
10.5
1437.5
324.1
618.8
126.8
21.0
 | 20.2
327.3
12.4
765.9
5.1
760.7
314.8
473.2
136.1
24.4 | 27.15
324.3
6.9
837.4
1.3
836.1
317.4
514.9
129.1
23.5
 | 27.16
354.5
6.5
1418.4
4.7
1413.7
348.0
571.9
131.0 | 36.32
341.3
3.2
1156.6
0.5
1156.1
338.1
634.1
138.1
634.1 | 50.13
459.0
4.4
807.4
0.0
807.4
454.5
641.4
102.4
 | 6.1
374.2
8.6
1093.3
4.9
1088.4
365.6
656.4
120.9
23.4 |
| Strenger Jaming minur resource us wronger Jamins ger 3 years
95 perc. relative to mean (N)
5 perc. relative to mean (N)
man relative to mean (N)
difference SS - sper-calive to mean (N)
difference SS - sper-calive to mean (N)
25 perc. relative to mean (N)
32 perc. relative to mean (N)
34 perc. relative | 109.21
327.6
11.4
1461.7
1.0
1460.7
316.1
595.9
133.4
19.9
985.7

 | 109.9
385.3
4.7
1863.1
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1859.3
380.5
668.3
137.4
9.9 | 12.207
344.8
5.2
1686.7
2.7
1683.9
339.6
808.1
120.7
18.7
1126.3

 | 12.7
354.3
5.7
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4.2
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804.5
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14.9
1031.9 | 16.128
380.8
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1264.3
4.1
1260.2
374.8
782.5
115.6
15.3
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 | 16.132
335.6
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1177.1
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324.6
665.6
120.6
24.1
866.2

 | 16.193
382.7
6.2
1560.0
3.8
1556.3
376.5
670.1
119.4
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 | 2.267
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 | 2.479
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21.9
974.0 | 20.2
327.3
12.4
765.9
5.1
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 | 27.15
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6.5
1418.4
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 | 36.32
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| The approximation of the second secon | 109.21
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 | 2.479
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| The approximation of the second secon | 109.21
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115.6
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911.0
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19.4

 | 16.132
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11.0
1177.1
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1169.4
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27.8
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 | 16.193
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1560.0
3.8
1556.5
670.1
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 | 2.267
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650.8
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unregul:
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