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1 **PORE DIFFUSION LIMITS REMOVAL OF MONOCHLORAMINE IN**  
2 **TREATMENT OF SWIMMING POOL WATER USING GRANULAR ACTIVATED**  
3 **CARBON**

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15 **Key words:** swimming pool water; monochloramine; surface catalytic reaction;  
16 granular activated carbon; diffusion control

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18

19

20 **Abstract**

21 Overall apparent reaction rates for the removal of monochloramine (MCA) in granular  
22 activated carbon (GAC) beds were determined using a fixed-bed reactor system and  
23 under conditions typical for swimming pool water treatment. Reaction rates dropped  
24 and quasi-stationary conditions were reached quickly. Diffusional mass transport in the  
25 pores was shown to be limiting the overall reaction rate. This was reflected consistently  
26 in the Thiele modulus, in the effect of temperature, pore size distribution and of grain  
27 size on the reaction rates. Pores <2.5 times the diameter of the monochloramine  
28 molecule were shown to be barely accessible for the monochloramine conversion  
29 reaction. GACs with a significant proportion of large mesopores were found to have  
30 the highest overall reactivity for monochloramine removal.

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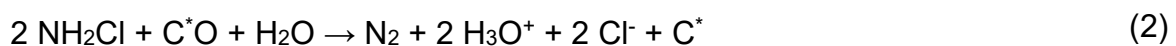
## 32 **1 Introduction**

33 Hypochlorous acid (HOCl) is widely used as disinfectant in swimming pools due to its  
34 effectiveness in disinfection. A major drawback is its reactivity towards inorganic and  
35 organic matter present in pool water, which results in the formation of halogenated  
36 disinfection by-products (DBPs) (Zwiener et al., 2007; Deborde and von Gunten,  
37 2008).

38 The dominating nitrogenous precursors for DBP formation that are introduced into the  
39 pool by bathers are urea and ammonia (Consolazio et al., 1963). Ammonia reacts with  
40 HOCl to progressively form inorganic chloramines (mono-, di- and trichloramine)  
41 (Blatchley and Cheng, 2010; Qiang and Adams, 2004). Urea, however, reacts in pool  
42 water through progressive chlorine addition forming chlorinated urea, which finally  
43 breaks down to form the very volatile trichloramine (Blatchley and Cheng, 2010).

44 Chloramines are known to be irritating to skin and eyes, and are suspected to cause  
45 respiratory problems (Eichelsdörfer et al., 1975). This also includes an increased risk  
46 to children of asthma during adolescence (Bernard et al., 2003; Bernard et al., 2008).  
47 Consequently, the sum of all inorganic chloramines is strictly regulated to a  
48 concentration of  $<0.2 \text{ mg L}^{-1}$  (as  $\text{Cl}_2$ ) in swimming pools in Germany (DIN 19643-2,  
49 2012) and the US (ANSI/APSP-1, 2009). Among the variety of inorganic chloramines,  
50 monochloramine (MCA) is the most dominant species in pool water. A study of 11  
51 swimming pools in the US found concentrations of monochloramine of up to  
52  $1.88 \text{ mg L}^{-1}$  (as  $\text{Cl}_2$ ) (Weaver et al., 2009). Moreover, MCA is particularly of interest  
53 because it has been found to be a precursor for the formation of carcinogenic  
54 N-nitrosodimethylamine (NDMA) (Schreiber and Mitch, 2006), which was also found in  
55 chlorinated pools (Walse and Mitch, 2008).

56 As bathers are introducing urea and ammonia into swimming pools, the removal of the  
57 chloramines formed by the reaction of ammonia can decrease the total concentration  
58 of chloramines in the pool. Activated carbons are widely known as effective adsorbents  
59 (Worch, 2012; Schreiber et al. 2005). However, previous studies have shown that they  
60 are reducing agents for MCA as well. In that, the reduction of MCA primarily proceeds  
61 via a surface chemical reaction, and the capacity of granular activated carbon (GAC)  
62 filters for MCA removal was found to be higher than expected when only adsorption  
63 was considered (Jaguaribe et al., 2005; Bauer and Snoeyink, 1973). Over the course  
64 of the reaction, the reactivity of GACs decreased before stationary conditions were  
65 reached (Scaramelli and Digiano, 1977). As a result, activated carbon filters operated  
66 under swimming pool water conditions have a finite life time and need to be  
67 regenerated or replaced after a certain time in operation. However, the exact  
68 mechanism of the initial decrease of reactivity has not been fully understood until now.  
69 Previous studies proposed a two-step process, where MCA is initially reduced at  
70 neutral pH to  $\text{NH}_4^+$  at free active carbon sites ( $\text{C}^*$ ) (Equation 1). As enough surface  
71 oxides were formed, the authors assume that MCA is oxidised to  $\text{N}_2$  in a second, slower  
72 reaction with surface oxides (Equation 2) (Bauer and Snoeyink, 1973).



73 However, a more recent study indicates that the ratio of transformation products ( $\text{NH}_4^+$   
74 and  $\text{N}_2$ ) formed did not change over the course of the reaction (Fairey et al. 2007),  
75 which contradicts the hypothesis of (Bauer and Snoeyink, 1973). Furthermore, earlier  
76 studies reasoned that the overall reactivity is dominated by pore diffusion, rather than  
77 by the surface chemical reactions (Komorita and Snoeyink, 1985; Fairey and Speitel,  
78 2006). Mass transport in the pore system of GACs is known to be influenced by various

79 parameters such as the grain size, water temperature (Ertl et al., 1997), the presence  
80 of surface oxides (Mangun et al., 1999) and the pore size distribution of GACs (Maia  
81 et al., 2008). To describe theoretically the MCA removal process in GAC filters, the  
82 semi-empirical MCA catalysis model (MCAT model) was established by Kim (1977) to  
83 predict steady-state MCA removal in GAC filters (Kim, 1977). The MCAT model was  
84 complemented recently to take into account the effect of source water type and pH  
85 (Fairey and Speitel, 2006). However, the model estimates the GAC reactivity after  
86 stationary conditions are reached (Komorita and Snoeyink, 1985) using only basic  
87 GAC properties such as the porosity and tortuosity of the GACs. This makes the model  
88 inapplicable in elucidating mass transfer effects in detail, which is important when  
89 considering the potentials for optimisation of the process.

90 Thus, considering the current state of knowledge, the process of decreasing reactivity  
91 of GACs with increasing reaction time, as well as the impact of pore diffusional  
92 limitations on the removal performance of a fixed GAC bed, is still not fully understood.  
93 Conclusively, the aim of this study was to determine the impact of diffusional mass  
94 transport on the overall MCA removal performance of GAC for a wide range of  
95 operation conditions, such as grain size, pore size distribution and water temperature.

## 96 **2 Materials and Methods**

### 97 **2.1 Fixed bed reactor system for determination of the reaction rate** 98 **constants**

99 Reaction rate constants for monochloramine removal in GAC beds were determined  
100 using the fixed bed reactor (FBR) system shown in Figure 1. It included a glass column  
101 of 34 mm inner diameter and a total length of 200 mm, in which the GAC bed is placed  
102 on a glass frit. A glass frit close to the inlet ensures equal distribution of the flow to the  
103 GAC bed. The GAC bed was continuously fed with an MCA solution at a constant flow  
104 rate ( $Q_{\text{bed}}$ ) from a rapidly stirred ( $\sim 250$  rpm), double-walled glass tank of 10 L volume,  
105 equipped with a stainless steel lid and with feedback temperature control (B. Braun,  
106 Germany). The conductivity of the MCA solution was adjusted to  $\sim 400 \mu\text{S cm}^{-1}$  by  
107 dosing  $1 \text{ mol L}^{-1}$  NaCl solution before the experiment was started. The volumetric flow  
108 rate through the bed was maintained by a turbine pump (P1) (BG1-30, Gather  
109 Industries, Germany) and was measured using a magnetic inductive flow meter  
110 (Altoflux IFM 5080 K with IFC 080 transformer, Krohne, Germany).

111 The GAC bed was prepared by initially filling the glass column, disconnected from the  
112 system, with deionised water. Subsequently, GAC that had been wetted and the pores  
113 filled with deionised water by applying a vacuum, was transferred to the column. The  
114 bed was then consolidated by carefully tamping the column to guarantee a  
115 reproducible packing. After connecting the column to the FBR system, the chloramine  
116 solution was circulated through all instruments and tubes while bypassing the column.  
117 The experiment was started then by changing the fluid flow from bypass to the GAC  
118 bed.

119 The effluents from the GAC bed or the bypass were fed back to the tank. The MCA  
120 concentration in the tank was measured by an amperometric chlorine sensor

121 (Dulcotest CTE 1 DMT, Prominent, Germany), mounted in a closed loop and flowed  
122 through by a peristaltic pump (P2) (5004S, Watson Marlow, UK) at a flow rate of 25 L h<sup>-1</sup>.  
123 The pH in the tank was measured continuously and maintained constant at pH 7,  
124 which was in the pH range for swimming pool water (DIN 19643-2, 2012), by dosage  
125 with 0.3 mol L<sup>-1</sup> phosphoric acid or 0.1 mol L<sup>-1</sup> NaOH using feedback-controlled  
126 peristaltic pumps (P4, P5). The level in the tank was kept constant by a peristaltic pump  
127 (P3) triggered by a level electrode.

128 Since MCA was consumed in the GAC bed, the MCA concentration in the tank and  
129 thus the MCA inflow concentration to the GAC bed,  $c_{bed,in}$ , were kept constant by  
130 feeding a cooled (4–6 °C) MCA stock solution to the tank using a feedback-controlled  
131 peristaltic micro pump (P6) (BVK, Ismatec, Germany). That stock solution was placed  
132 on a balance (CA 572, Kern, Germany) and the mass of the stock solution was  
133 recorded continuously. Thus, at any time  $t$ , the mass flow of stock solution to the tank  
134 was known. A SCADA system (TopMessage, Delphin Technologies, Germany) was  
135 used for data recording and feedback control of all relevant process parameters of the  
136 FBR system.



137 Figure 1: Scheme of the laboratory-scale FBR system.

## 138 2.2 Preparation of the monochloramine stock solution

139 Monochloramine stock solutions as used in the FBR experiments were prepared by  
140 drop-wise addition of 750 mL of a  $\text{OCl}^-$  solution (pH 10,  $0.032 \text{ mol L}^{-1}$ , prepared from a  
141 sodium hypochlorite solution (>12 % active chlorine)) to 250 mL of a rapidly stirred  
142  $\text{NH}_4\text{Cl}$  or  $(\text{NH}_4)_2\text{SO}_4$  solution (pH 10,  $0.098 \text{ mol L}^{-1}$  ( $\text{NH}_4\text{Cl}$ ) or  $0.049 \text{ mol L}^{-1}$   
143 ( $(\text{NH}_4)_2\text{SO}_4$ )) at a final molar chlorine-to-ammonia ratio of 1.00:1.03 (Aoki, 1989). All  
144 chemicals were p.a. reagent grade.

## 145 2.3 Analytical quantification of monochloramine and HOCl

146 The total chlorine sensor of the FBR system was calibrated daily using the  
147 spectrophotometric DPD method (DIN EN ISO 7393-2), using a Unicam  
148 UV2-200 UV/VIS spectrophotometer with a 5 cm quartz cuvette.

149 The concentration of the  $\text{OCl}^-$  solution used to prepare the MCA stock solution was  
150 determined spectrophotometrically using a molar absorption coefficient at 294 nm ( $\epsilon_{\text{OCl}^-}$   
151  $_{,294\text{nm}}$ ) of  $348 \text{ mol}^{-1} \text{ cm}^{-1}$  (Hand and Magerum, 1983).

152 The MCA stock solution was determined at least twice a day according to the  
153 spectrophotometric method of Schreiber and Mitch, accounting for the overlapping  
154 absorbance peaks of MCA and dichloramine ( $\text{NHCl}_2$ ) at 245 nm and 295 nm  
155 ( $\epsilon_{\text{NH}_2\text{Cl}, 245 \text{ nm}} = 445 \text{ mol}^{-1} \text{ cm}^{-1}$ ,  $\epsilon_{\text{NHCl}_2, 245 \text{ nm}} = 208 \text{ mol}^{-1} \text{ cm}^{-1}$ ,  $\epsilon_{\text{NH}_2\text{Cl}, 295 \text{ nm}} = 14 \text{ mol}^{-1}$   
156  $\text{cm}^{-1}$ ,  $\epsilon_{\text{NHCl}_2, 295 \text{ nm}} = 267 \text{ mol}^{-1} \text{ cm}^{-1}$ ) (Schreiber and Mitch, 2005). Measurements at  
157 360 nm confirmed the absence of trichloramine ( $\text{NCl}_3$ ) in the MCA stock solution  
158 ( $\epsilon_{\text{NCl}_3, 360 \text{ nm}} = 126 \text{ mol}^{-1} \text{ cm}^{-1}$ ) (Schurter et al., 1995). The yield for transformation of  
159  $\text{NH}_4^+\text{-N}$  to MCA-N was found to be in a range from 95 to 100 %.

## 160 **2.4 Granular activated carbons**

161 Four commercially available granular activated carbons (GACs) were used in this  
162 study, namely Hydriffin 30 N from Donau Carbon GmbH (30 N), Silcarbon K-835 from  
163 Silcarbon Aktivkohle GmbH (K835), Centaur from Chemviron Carbon GmbH (Centaur)  
164 and a spherically shaped activated carbon Type 100058 from Blücher GmbH (100058).  
165 The raw materials of the carbons were anthracite coal (30N), coconut shells (K835),  
166 bituminous coal (Centaur) and non-porous polymer-based spheres (100058). The  
167 GACs, as provided by the manufacturers, are denoted as unfractionated fresh GACs  
168 in the following.

169 Grain size fractions of the 30N, K835 and Centaur with mean grain diameters of  
170 0.30 mm (0.25 - 0.355 mm), 0.57 mm (0.50 - 0.63 mm) and 1.90 mm (1.80 - 2.00 mm)  
171 were prepared by sieving the unfractionated fresh GACs using a sieve tower (AS 200,  
172 Retsch, Germany) and, if necessary, by grinding the unfractionated fresh GACs using  
173 a ball mill (PM100, Retsch, Germany) prior to sieving. The 100058 GAC was used  
174 solely at the original monomodal grain size ( $d_p$ ) provided by the manufacturer  
175 (0.55 mm). Before use, the fresh GACs were treated by: (i) soaking in ultrapure water  
176 for 24 h, (ii) evacuating the soaked GAC using a vacuum chamber until rising air  
177 bubbles could no longer be seen and (iii) washing and decantation of the GAC using  
178 ultrapure water until the supernatant was particle-free.

## 179 **2.5 Physical characterisation of the GAC**

180 Evaluation of the bed volume specific outer surface area  $a_o$  and the representative  
181 hydraulic grain size  $d_{hy}$  of the unfractionated GACs is described in Section A of the  
182 supplementary material (SM).

183 The internal surface area  $a_{\text{BET}}$ , the pore volume of micropores (<2 nm) and mesopores  
184 (2–50 nm), the pore size distribution (PSD) and the tortuosity of the fresh  
185 unfractionated GACs were determined from nitrogen adsorption/desorption isotherms  
186 (BELSORP-max, MicrotracBEL). The pore volume of macropores (>50 nm) was  
187 determined by mercury intrusion porosimetry (Pascal 140/440, Thermo Fisher  
188 Scientific). To further elucidate the morphology of the 100058 GAC, images were taken  
189 from a cross-section of a single GAC grain with a High Resolution Scanning Electron  
190 Microscope (FEI Nova NanoSEM, 5 kV). Further details on the experimental  
191 procedures of physical GAC characterisation are given in Section B of the SM.

## 192 **2.6 Data analysis and calculation of bed-volume-based apparent first order** 193 **rate constants**

194 MCA is consumed in the GAC bed. According to equations (1) and (2) above it is  
195 expected to follow first-order kinetics with respect to MCA if reaction (1) was slower  
196 than reaction (2) and thus rate limiting. If reaction (2) was rate limiting, then second-  
197 order kinetics is expected. Finally, mass transport inside the pores may be rate-limiting  
198 (diffusion control). In that case a first-order reaction will remain first-order and a  
199 second-order reaction will turn into an apparent 1.5-order reaction. To take that into  
200 account, apparent rate constants for a pseudo first-order reaction were determined and  
201 the validity of that approach was verified in separate experiments (see Section 2.10).

202 For a fixed-bed reaction at stationary conditions, the apparent bed-volume-based  
203 reaction rate constant  $k_{\text{app}}$  is obtained according to:

$$k_{app} = \frac{1}{EBCT} \cdot \ln \left( \frac{C_{bed,in}}{C_{bed,out}} \right) \quad (3)$$

204 where  $C_{bed,in}$  and  $C_{bed,out}$  are the GAC bed's MCA in- and outflow concentrations and  
 205 EBCT is the empty bed contact time, given by  $V_{bed}/Q_{bed}$ , where  $V_{bed}$  and  $Q_{bed}$  are the  
 206 bed volume and the volumetric flow rate, respectively.

207 To ensure that all GAC in the filter bed was subjected to the almost same concentration  
 208 of MCA, it was designed such that the change of the concentration from the inflow to  
 209 the outflow of the bed was negligible. Generally, typical outflow concentrations of  
 210 monochloramine over the course of an experiment were about 5% below the inflow  
 211 concentration (see Figure 2).

212 Consequently, the measurement of the outflow concentration would have resulted in  
 213 a lack of precision. To overcome that restriction, the outflow concentration was  
 214 obtained from the decreasing mass of the MCA stock solution. The volumetric dosing  
 215 rate ( $Q_{stock}$ ) for any time  $t$  during an experiment is given by Equation 4:

$$Q_{stock}(t) = \frac{\left( \frac{dm_{stock}}{dt} \right)}{\rho_{stock}(T)} \quad (4)$$

216 Here,  $\rho_{stock}(T)$  is the temperature-dependent density of the stock solution and  $m_{stock}$  is  
 217 the recorded mass of the stock solution on the balance. The time-dependent mass loss  
 218  $dm_{stock}/dt$  was determined in time intervals of 0.5–1 h.

219 A mass balance for MCA around the tank of the FBR system yields:

$$Q_{\text{bed}} \cdot c_{\text{bed,out}}(t) - Q_{\text{bed}} \cdot c_{\text{bed,in}} + Q_{\text{stock}}(t) \cdot c_{\text{stock}} - Q_{\text{stock}}(t) \cdot c_{\text{bed,in}} + R_{\text{sys}}(t) = V_{\text{sys}} \cdot \left( \frac{dc_{\text{bed,in}}}{dt} \right) \quad (5)$$

220 Here  $R_{\text{sys}}$  accounts for a hypothetical loss of MCA in the system excluding the GAC  
 221 bed (e.g. through reaction at the glass or tube walls) and  $c_{\text{stock}}$  is the concentration of  
 222 the MCA stock solution. Preliminary experiments without GAC showed that MCA  
 223 removal in the FBR system without GAC was negligible, thus:

$$R_{\text{sys}} = 0 \quad (6)$$

224 Since  $c_{\text{bed,in}}$  was kept constant, the capacity term in Equation 5 gives:

$$V_{\text{sys}} \left( \frac{dc_{\text{bed,in}}}{dt} \right) = 0 \quad (7)$$

225 Introducing Equations 6 and 7 into Equation 5 and rearranging yields:

$$c_{\text{bed,out}}(t) = \frac{Q_{\text{stock}}(t)}{Q_{\text{bed}}} \cdot (c_{\text{bed,in}} - c_{\text{stock}}) + c_{\text{bed,in}} \quad (8)$$

226 Introducing Equation 8 into Equation 3 gives the bed volume based rate constant  $k_{\text{app}}$   
 227 at any time  $t$ .

## 228 **2.7 Catalyst surface-area-based rate constants**

229 As the reduction of MCA by GAC is a solid surface reaction, catalyst surface-area-  
 230 based apparent rate constants  $k_{\text{app,a}}$  are calculated from the bed-volume-based rate  
 231 constants according to:

$$k_{\text{app,a}} = \frac{k_{\text{app}}}{a_0} \quad (9)$$

232 where  $a_0$  is the bed-volume-specific outer surface area of the GAC in the FBR.

## 233 2.8 Conversion dependence of the overall reaction rate

234 Previous studies showed that the reactivity of GACs for MCA conversion decreased  
235 gradually over time (Scaramelli and Digiano, 1977). However, using the FBR system  
236 described above, the time needed to acquire the data for determination of the rate  
237 constant at a given time  $t$  was short compared to the time needed for a notable  
238 decrease of GAC reactivity. Thus, for each time  $t$ , the assumption of quasi-steady state  
239 was justified.

240 To describe the process of decreasing reactivity of the heterogeneous MCA–GAC  
241 reaction, conversion–time curves were determined in FBR experiments for all GACs  
242 until the reactivity was constant over time. The general shape of the conversion–time  
243 curves can indicate which one of the two widely recognised reaction models, the  
244 Progressive-Conversion Model (PCM) or the Shrinking-Core Model (SCM)  
245 (Levenspiel, 1999) is appropriate.

246 FBR experiments were conducted to determine  $k_{app}$  either at the beginning of the  
247 MCA–GAC reaction or after a certain time of operation, when stationary conditions  
248 were reached. The reactivity at the beginning of the reaction was determined after  
249  $\sim 0.1$  mmol  $g^{-1}$  of MCA per mass of GAC had been converted in the GAC bed. For these  
250 experiments, the fresh GACs were used as provided by the manufacturer.

251 To determine the reactivity under stationary conditions, the fixed-bed experiments  
252 were divided into two consecutive steps: (i) subjecting the GACs to monochloramine  
253 at  $c_{bed,in} = 4.5$  mg  $L^{-1}$  (as  $Cl_2$ ),  $T = 30$  °C and  $Q_{bed} = 40$  L  $h^{-1}$  (equalling a superficial filter  
254 velocity of  $v_{bed} = 44.1$  m  $h^{-1}$ ) until no change in  $k_{app}$  was observed and stationary  
255 conditions were reached and, (ii) determination of  $k_{app}$  under various stationary  
256 operation conditions (e.g. different  $c_{bed,in}$ ,  $T$  or  $v_{bed}$ ). GACs operating under stationary

257 conditions were denoted as altered GACs. A detailed description of the  
258 FBR experiments is provided in the following sections. Table 1 summarises all  
259 experiments conducted.

260 Table 1: Operation conditions of the FBR experiments.

## 261 **2.9 Temperature dependence of rate constants**

262 The temperature dependence of  $k_{app}$  is described using the Arrhenius Equation:

$$k_{app}(T)=k_0 \cdot e^{-\frac{E_A}{R_0 T}} \quad (10)$$

263 Here,  $k_0$  and  $E_A$  are the frequency factor and the activation energy, respectively.  $R_0$  is  
264 the universal gas constant and  $T$  is the temperature.  $E_A$  is derived from the slope of  
265 the linear least-squares best fit of the correlation between  $\ln(k_{app})$  and  $T^{-1}$  (Arrhenius  
266 plots) (Levenspiel, 1999).

267 To describe the temperature dependence of MCA conversion in the GAC filter, the  
268 activation energies  $E_A$  (see Equation 10) of the reaction for both the fresh and the  
269 altered unfractionated GACs: 30N, K835 and Centaur, were determined. In order to  
270 calculate  $E_A$ ,  $k_{app}$  was experimentally determined in FBR experiments in a range of  
271 temperatures of 12 to 45 °C. Other process parameters except the temperature were  
272 kept constant (see Table 1).

## 273 **2.10 Verification of first-order kinetics**

274 To experimentally verify the first-order kinetic approach used in this study, the reactivity  
275 of the altered unfractionated K835 GAC was determined for various MCA inflow  
276 concentrations. The reproducibility of the measurement of  $k_{app}$  was determined from a  
277 set of four repeated experiments.

## 278 **2.11 Impact of external mass transport on the overall reactivity**

279 To assess the impact of pore diffusional mass transport on the overall reactivity of the  
280 GAC, the influence of extra-particle mass transport (film diffusion) on the overall  
281 reaction rate must be excluded (Ertl et al. 1997). To check for the absence of film  
282 diffusion resistance,  $k_{app,a}$  was determined for different superficial flows ( $v_{bed}$ ) for filter  
283 beds of the fresh and altered unfractionated K835 GAC. To guarantee comparable  
284 conditions in all experiments, the flow rate  $v_{bed}$  was adjusted to the actual catalyst  
285 volume ( $V_{bed}$ ) such that the EBCT in the filter column was 2.88 s.

286 The mass transfer coefficient of MCA through the laminar film  
287 layer ( $k_f$ ) was calculated according to Equation 11 (Worch, 2012).

$$k_f = \frac{Sh \cdot D_{bulk,MCA}}{d_{hy}} \quad (11)$$

288 Here Sh is the Sherwood number, which was calculated according to the empirical  
289 approaches of Williamson et al. (1963) and  
290 Gnielinski (1978), and  $D_{bulk,MCA}$  is the bulk diffusion coefficient of MCA in water, which  
291 was calculated using the Wilke-Chang correlation with  $2.03 \times 10^{-5} \text{ cm}^2 \text{ s}^{-1}$  (at 30°C)  
292 (Wilke and Chang, 1955).

293 If the rate of mass transfer through the laminar film layer is much larger than the  
294 surface-based reaction rate constant, thus  $k_f \gg k_{app,a}$ , external mass transfer limitation  
295 can be excluded.



## 296 2.12 Impact of internal mass transport on the overall reactivity

### 297 2.12.1 Impact of grain size

298 To check for the importance of grain size on the overall reactivity,  $k_{app}$  was determined  
299 for different grain size fractions of the altered 30N GAC. To further elucidate the impact  
300 of grain size, the Thiele modulus  $\Phi$  was calculated according to Equation 12  
301 (Levenspiel, 1999). The Thiele modulus is a dimensionless constant representing the  
302 ratio of the overall reaction rate to the diffusion rate of the reactant in the pore system.

$$\Phi = L \cdot \sqrt{\left( \frac{k_{app}}{\eta \cdot D_{E,MCA}} \right)} \quad (12)$$

303 Here,  $L$  is the diffusion path length within the adsorbent, which is  $d_p/6$  assuming a  
304 spherical shape of the GAC grains (Levenspiel, 1999).  $D_{E,MCA}$  is the effective diffusion  
305 coefficient of MCA in the pore system and  $\eta$  is the effectiveness factor.  $D_{E,MCA}$  was  
306 determined based on the pore size distribution of the individual GACs using the random  
307 intersecting pore model (Harriott, 2003) and assuming surface diffusion of MCA in the  
308 pore system of the GACs to be negligible (Fairey and Speitel, 2007):

$$D_{E,MCA} = \frac{1}{\tau} \cdot \frac{\sum_{j=d_{MCA}}^{\infty} D_{bulk, MCA} \cdot \left(1 - \frac{d_{MCA}}{d_{p,j}}\right)^4 \cdot V_{p,j}}{\sum_{j=d_{MCA}}^{\infty} V_{p,j}} \quad (13)$$

309 Here,  $\tau$  is the tortuosity factor of the GACs,  $d_{p,j}$  and  $V_{p,j}$  are the diameter and the  
310 incremental pore volume of a pore in the size fraction  $j$  as derived from the pore size  
311 distribution of the 30N GAC, and  $d_{MCA}$  is the minimal pore diameter of the GAC that is  
312 still accessible for a MCA molecule.

313 The effectiveness factor  $\eta$  is defined as the ratio of the reaction rate to the theoretical  
314 reaction rate without diffusional limitations (Ertl et al., 1997). For a first-order reaction  
315 taking place at a spherical grain,  $\eta$  could be determined as follows (Levenspiel, 1999):

$$\eta = \frac{1}{\Phi} \cdot \left( \frac{1}{\tanh 3\Phi} - \frac{1}{3\Phi} \right) \quad (14)$$

316 Numerical solutions of Equations 12 and 14 give the corresponding values for  $\Phi$  and  
317  $\eta$ .

318 **2.12.2 Impact of pore size distribution**

319 In this set of experiments,  $k_{app}$  was determined for the altered 30N, K835, Centaur and  
320 100058 GACs. To exclude the influence of GAC grain size on  $k_{app}$  and to allow  
321 comparison between the GACs, filter beds of equal grain size fractions were used  
322 (0.50 mm (100058) and 0.57 mm (30N, K835, Centaur)) to study the effect of the pore  
323 size distribution. It will be discussed and proven in detail in Sections 3.6 and 3.7 below  
324 that the prerequisites for the comparison between carbons of different surface  
325 chemical properties are fulfilled.

## 326 **3 Results and Discussion**

### 327 **3.1 Physical characterisation of the GACs**

328 The physical characteristics of the GACs tested in this study are summarised in Table  
329 2. More details and further discussion of the results are given in Section B of the SM.  
330 There, N<sub>2</sub> adsorption-desorption isotherms of the fresh unfractionated GACs and the  
331 corresponding pore size distributions (PSD) are shown in Figure B.1 and the grain size  
332 distribution of the fresh unfractionated GACs are shown in Figure B.2.

333

334 Table 2: Physical characterisation of the fresh unfractionated GACs 30N, K835,  
335 Centaur and 100058.

### 336 **3.2 Verification of the first-order reaction kinetics**

337 Figure 2 exemplarily shows time plots of  $C_{bed,in}$  and  $C_{bed,out}$  of one of the four repeated  
338 verification experiments, as well as  $k_{app}$ , as calculated according to Equation (3).  
339 Figure 2(A) shows the rapid decrease of the reaction rate constant until stationary  
340 conditions were reached after ~90 h. This is much shorter than the time reported in a  
341 previous study (1250–3000 h) (Fairey and Speitel, 2007), although comparable MCA  
342 inflow concentrations were used in both studies. This discrepancy can be attributed to  
343 the significant differences in EBCT of the respective GAC beds (30 s in Fairey and  
344 Speitel (2007) and 2.8 s in this work) and thus, the difference in the time needed to  
345 oxidise the GAC in the bed.

346 Figure 2(B) shows the determination of  $k_{app}$  at three different inflow concentrations that  
347 were adjusted after stationary conditions had been reached.

348 Figure 2: Reaction rate constant  $k_{app}$  for MCA removal at the K835 GAC at  
349  $C_{bed,in}$  of 4.5 mg L<sup>-1</sup> (as Cl<sub>2</sub>) until stationary conditions were reached  
350 (A) and at three additional inflow concentrations (9.0, 1.6 and 0.9  
351 mg L<sup>-1</sup> (as Cl<sub>2</sub>)) (B). Error bars represent the standard deviation of  
352  $k_{app}$  as determined by Gaussian error propagation (n = 50).

353 Figure 3 shows the linear least-squares regression analysis between  $k_{app}$  and  $C_{bed,in}$ .  
354 The slope of the linear regression was almost zero ( $[0.4 \pm 0.21] \times 10^{-3} \text{ L mg}^{-1} \text{ s}^{-1}$ ), which  
355 shows that the reaction rate constant was independent of the inflow concentration. This  
356 confirms the validity of the first-order approach used in Equation 3. This was also true  
357 for MCA concentrations that are as low as typically found in swimming pool water  
358 (e.g. 0–1.8 mg L<sup>-1</sup> as Cl<sub>2</sub>, (Weaver et al., 2009)). The repeatability standard deviation  
359 of  $k_{app}$  for the altered GAC K835 determined by four identical repeated validation  
360 experiments was  $\pm 0.006 \text{ s}^{-1}$  ( $\sim \pm 4\%$ ), which indicates a high reproducibility of the  
361 experimental method used.

362 Figure 3: Effect of MCA inflow concentration on the apparent reaction rate  
363 constant for MCA conversion at the altered unfractionated K835  
364 GAC. The solid line represents the linear least-squares best fit and  
365 dashed lines represent the 95% confidence band of the regression.  
366 Errors bars represent the repeatability standard deviation (n = 63).

367

368 Additionally, to check for heterogeneity of the reaction, the Maitlis' test was conducted  
369 (Crabtree, 2012). In that, the loss of MCA in the tank of the FBR system was measured  
370 once during the course of an experiment while bypassing the GAC filter column.

371 Results confirmed that no MCA was degraded and thus, no reactive agents leached  
372 from the GAC.

### 373 **3.3 Exclusion of extra-particle mass transfer limitations**

374 Figure 4 shows the dependence of the apparent surface-related reaction rate constant  
375  $k_{app,a}$  (in  $m\ s^{-1}$ ), determined experimentally, and of the theoretical mass transfer  
376 coefficient of MCA in the laminar film layer  $k_f$  (in  $m\ s^{-1}$ ), on the filter velocity  $v_{bed}$  for the  
377 fresh and altered unfractionated GAC K835. It showed that  $k_{app,a}$  was independent of  
378 the superficial flow rate  $v_{bed}$  for both, the fresh and altered GAC ( $p < 0.001$ ). As  
379 mentioned above, two empirical approaches for the calculation of the Sherwood  
380 number were used. It appeared that  $k_{app,a}$  was  $\sim 4$  times lower than  $k_f$  when the  
381 Sherwood number was approximated according to Williamson et al. (1963) and  
382  $\sim 10$  times lower when the approximation according to Gnielinski (1978) was used.  
383 Similar results were obtained for the fresh 30N GAC (results not shown). This proves  
384 that external mass transport is not limiting the overall reactivity (Worch, 2008) for MCA  
385 removal in GAC filters over a wide range of superficial filter velocities for both fresh  
386 and altered GACs.

387 In previous work it was assumed that the negligible impact of film diffusion on the  
388 removal of dichloramine at GACs would also apply to the removal of MCA (Fairey and  
389 Speitel, 2007; Kim and Snoeyink, 1980). Our results prove that this assumption is  
390 justified.

391 Figure 4: Impact of filter velocity on the experimentally determined apparent  
392 surface-related reaction rate constant  $k_{app,a}$  for MCA conversion and  
393 the theoretical mass transfer coefficient  $k_f$  of the unfractionated K835  
394 GAC. Error bars represent the 95% confidence interval. Mass  
395 transfer coefficients  $k_f$  were calculated using a mean grain diameter  
396  $d_{hy,K835}$  of 1.39 mm. The dashed vertical line represents the  
397 recommended filter velocity for GAC filters in swimming pool water  
398 treatment in Germany ( $\leq 30 \text{ m h}^{-1}$ ) (DIN 19643-2, 2012).

### 399 3.4 Progress of the reaction until stationary conditions were reached

400 Figure 5(A) shows the conversion–time curves of MCA in the GAC filter for the  
401 0.57 mm grain size fractions of the 30N, K835, Centaur and 100058 GAC. The  
402 respective reaction rates are given in Figures 5(B) and 5(C).

403 It is apparent that the conversion–time plot of MCA removal for the conventional GACs  
404 (30N, K835, Centaur), which comprise a homogeneously distributed and strongly  
405 microporous pore size distribution, is hyperbolic, while that of the 100058 GAC is  
406 sigmoid (S-shaped). The conversion–time behaviour is reflected in the behaviour of  
407 the reaction rate over time.  $k_{app}$  of the GACs 30N, K835 and Centaur drop continuously.  
408 Instead,  $k_{app}$  of the 100058 first shows an increase and then decreases continuously  
409 after a maximum has been reached after about 15 h.

410 The sigmoid conversion–time behaviour is often associated with a shift in the reaction-  
411 controlling mechanism from chemical reaction rate control to diffusional control and  
412 can be described by the Shrinking-Core Model (SCM) (Levenspiel, 1999). The SCM  
413 describes a reaction that starts first at the outer surface of the GAC grains, and the  
414 reaction front then moves towards the centre of the grains with ongoing reaction time

415 (Levenspiel, 1999). The difference in conversion–time behaviour between the  
416 microporous GACs and the 100058 GAC is explained in more detail in Section C of  
417 the SM.

418  
419 Figure 5: MCA removed (A) and bed volume-based reaction rate constant  $k_{app}$   
420 for MCA removal (B, C) over the filter run time. All GACs had the  
421 same grain size of ~0.55 mm (0.57 mm for the 100058 GAC).  
422 Dashed horizontal lines in (B) represent the GAC reactivity when  
423 stationary conditions were reached.

424

### 425 **3.5 Impact of grain size**

426 Figure 6(A) presents the time course of the bed-volume based first-order reaction rate  
427 constant  $k_{app}$  for MCA removal by different grain size fractions of the 30N GAC until  
428 stationary conditions were reached. The bed-volume-based reaction rate constants at  
429 stationary conditions are displayed in 6(B). They decrease linearly with increasing  
430 grain diameter. These results are in agreement with previous studies which, however,  
431 were obtained before stationary conditions were reached (Komorita and Snoeyink,  
432 1985).



433  
434 Figure 6: (A) Reaction rate constant  $k_{app}$  over the filter runtime for different  
435 grain size fractions of the 30N GAC. Dashed horizontal lines  
436 represent the level of GAC reactivity at stationary conditions. (B)  $k_{app}$   
437 and (C)  $k_{app,a}$  at stationary conditions as function of the grain  
438 diameter. Open circle symbols in (B) and (C) represent  $k_{app}$  for the  
439 unfractionated 30N GAC using  $d_{hy,30N} = 1.18$  mm as representative  
440 grain size (data not shown in A). Error bars in (B) and (C) represent  
441 the 95% confidence intervals.

442

443 Figure 6(C) shows the catalyst surface-area-related rate constants for different grain  
444 size fractions of the 30N carbon. If the reaction only took place at the outer surface of  
445 the GAC grains, then  $k_{app,a}$  would be independent of the grain diameter. The decrease  
446 indicates that diffusional resistances in the pore systems are limiting the overall  
447 reaction rate.

448 To assess further the influence of pore diffusion on the overall reactivity, the Thiele  
449 modulus  $\Phi$  and the effectiveness factor  $\eta$  were calculated using the stationary effective  
450 bed-volume-based reaction rate constants. The minimal pore diameter of the GAC that  
451 is accessible for the MCA molecule,  $d_{MCA}$ , was needed in equation (13). As will be  
452 discussed further in Section 3.6,  $d_{MCA}$  was assumed to be approximately twice the  
453 molecular diameter of the MCA molecule. The individual tortuosity factors  $\tau$  of the  
454 carbons needed to determine  $d_{MCA}$  were calculated using the CSTM model and are  
455 given in Section B of the SM.

456

457 Figure 7 shows  $\Phi$  as function of  $\eta$ . The Thiele modulus of the grain size fractions  
458 1.9 mm, 1.18 mm (unfractionated GAC) and 0.57 mm were  $>4$ , which confirmed that  
459 pore diffusion strongly controls the overall reaction. The Thiele modulus of the smallest  
460 grain size fraction (0.30 mm) was 3, which is related to the transient region with  
461 moderate pore diffusional influence (Levenspiel, 1999). In a previous study (Scaramelli  
462 and Digiano, 1977) a Thiele modulus of 0.51 was found for a GAC with a grain size of  
463  $\sim 0.5$  mm. The Thiele modulus of the 0.57 mm fraction of the 30N GAC observed in this  
464 study was higher by a factor of  $\sim 15$ . This indicates that the importance of pore diffusion  
465 on the overall reactivity found in this study was higher compared to the previous results.  
466 This discrepancy is due to the fact that  $D_{E,MCA}$  of the 30N GAC ( $9.43 \times 10^{-7} \text{ cm}^2 \text{ s}^{-1}$ )  
467 used to calculate  $\Phi$  in the present study was lower by a factor of  $\sim 4$  compared to that  
468 used previously, where an estimated  $D_{E,MCA}$  of  $5.484 \times 10^{-6} \text{ cm}^2 \text{ s}^{-1}$  was used  
469 (Scaramelli and Digiano, 1977). However, they had no data available regarding the  
470 pore size distribution and based their estimation on the GACs porosity and roughly  
471 estimated tortuosity.

472  
473 Figure 7: Effectiveness factor  $\eta$  of the MCA-GAC reaction at stationary  
474 conditions for different grain size fractions of the 30N GAC as  
475 function of the Thiele modulus  $\Phi$ . Open triangle symbols represent  
476  $k_{app}$  for the unfractionated 30N GAC using  $d_{hy,30N} = 1.18$  mm. Filled  
477 triangle symbols represent  $k_{app}$  for the fractionated 30N GAC. The  
478 dashed black line is for orientation only and represents the expected  
479 relationship between  $\Phi$  and  $\eta$  as discussed elsewhere (Ertl et al.,  
480 1997). Error bars represent the 95% confidence intervals

### 481 3.6 Impact of pore size distribution

482 The magnitudes of  $\Phi$  and  $\eta$  (Figure 7) demonstrate a very strong limitation of the  
483 overall reaction rate by diffusion. Consequently, it is concluded that  $k_{app}$  is not, or only  
484 to a minor extent, affected by the intrinsic chemical reaction, which most likely differs  
485 among the different types of carbons tested.

486 Given the strong pore diffusional control of the overall reaction, conclusively, values of  
487  $k_{app}$  for the given set of equally sized GACs (Figure 5) are primarily affected by the  
488 effective diffusion coefficient of MCA in the pore system ( $D_{E,MCA}$ ), which in turn is a  
489 function of the pore size distribution (Equation 13). Thus, the GAC that exhibits the  
490 highest  $D_{E,MCA}$  should exhibit the highest  $k_{app}$  and vice versa.

491 Using the effective molecular diameter of MCA of  $\sim 0.5$  nm (calculated using the  
492 RasMol visualisation tool (Sayle and Milnerwhite, 1995)) as  $d_{MCA}$  in Equation 13 to  
493 calculate  $D_{E,MCA}$  for the different GACs) gives the following order of  $D_{E,MCA}$ :  $2.28 \times 10^{-6}$   
494  $\text{cm}^2 \text{s}^{-1}$  (30N)  $> 1.42 \times 10^{-6} \text{cm}^2 \text{s}^{-1}$  (K835)  $> 1.26 \times 10^{-6} \text{cm}^2 \text{s}^{-1}$  (Centaur). Obviously,  
495 this order does not comply with the order of  $k_{app}$  shown in Figure 8  
496 ( $0.036 \text{s}^{-1}$  (Centaur)  $> 0.025 \text{s}^{-1}$  (K835)  $> 0.017 \text{s}^{-1}$  (30N)  $> 0.016 \text{s}^{-1}$  (100058)).

497 By calculating  $D_{E,MCA}$  as a function of  $d_{MCA}$ , it appeared that the orders of  $D_{E,MCA}$  and  
498  $k_{app}$  correspond when  $d_{MCA}$  is equal or larger than  $\sim 1.3$  nm. This leads to the conclusion  
499 that pores  $< 1.3$  nm in width, which is about 2.5 times the molecular diameter of a MCA  
500 molecule, might be barely accessible for MCA. This can be explained by constrictions  
501 at the pore mouth caused by water molecules that adsorb onto oxygen-containing  
502 functional groups present at the GAC surface by hydrogen bonding (McCallum et al.,  
503 1999). These findings are in agreement with those found in previous studies, where it

504 was assumed that certain pores might be barely accessible for MCA molecules due to  
505 diffusional limitations (Fairey et al., 2006).

506 The significantly higher  $D_{E,MCA}$  found for the Centaur GAC for  $d_{MCA} > 1.3$  nm compared  
507 to the 30N and K835 GAC results mainly from the large number of mesopores ( $> 7$  nm),  
508 which the other GACs did not contain. It must be noted here that calculation of  $D_{E,MCA}$   
509 for the 100058 GAC according to Equation 13 is not eligible due to the structured pore  
510 size distribution of the GAC grains (Harriott, 2003).

### 511 **3.7 Impact of water temperature**

512 Figure 8 presents Arrhenius plots for MCA removal in GAC beds of the 30N, K835 and  
513 Centaur GAC, determined at temperatures between 12 and 45 °C. The activation  
514 energies  $E_A$  of the fresh and altered GACs derived from the Arrhenius plots ranged  
515 between 20.4 and 29.8 kJ mol<sup>-1</sup>. For typical swimming pool water temperatures (20 to  
516 35 °C ) an increase of activation energy from 20 to 30 kJ mol<sup>-1</sup> results in an increase  
517 in the reaction rates of 40% (fresh GACs) and 80% (altered GACs).

518 For the fresh unfractionated GACs investigated (30N and K835), activation energies  
519 were between 20.4 and 21.4 kJ mol<sup>-1</sup>. For the altered unfractionated GACs  
520 investigated (30N, K835 and Centaur) they were slightly higher, namely between 26.0  
521 and 29.3 kJ mol<sup>-1</sup>. However, the differences were not significant ( $p < 0.001$ ).

522 The activation energies determined are those for the combined pore diffusion – surface  
523 reaction process. It is interesting to note that temperature sensitivity was higher when  
524 the GACs had been altered, i.e. for lower overall reaction rates. For most chemical  
525 reactions, the activation energy is in the range of 50 to 250 kJ mol<sup>-1</sup>. For diffusional  
526 mass transport in water, the diffusion coefficient is approximately linearly proportional

527 to T. From that, an activation energy for diffusion in water of approximately 5 kJ mol<sup>-1</sup>  
528 can be calculated

529 For the temperature dependence of the combined processes of diffusion and chemical  
530 surface reaction in pores, it will consequently be lower for diffusion-controlled  
531 processes, i.e. when the chemical surface reaction is faster, and will be higher for  
532 chemical surface reaction-controlled processes. Thus, as the alteration of the GACs  
533 brings about a decrease in a chemical reaction rate, the higher activation energies  
534 found for the altered GACs (although not significant due to the small number of data  
535 points available) support the conclusions on diffusion control drawn above, as diffusion  
536 limits less for slower chemical reactions.

537 Additionally, the conclusions on diffusion control are supported by the fact that  
538 activation energies found for MCA removal in GAC filters in this study were significantly  
539 lower than those reported for the removal of dichloramine (35.6 kJ mol<sup>-1</sup> at pH 10 (Kim  
540 and Snoeyink, 1978)) and free chlorine (43.9 kJ mol<sup>-1</sup> at pH 7.6 (Suidan and Snoeyink,  
541 1977)). The lower activation energies are in agreement with the higher reactivity of  
542 GACs for the removal of dichloramine and free chlorine as compared with MCA.

543 As shown previously in this study, the impact of pore diffusion increases with increasing  
544 grain size of the GACs used. Consequently, this increase in diffusional resistance  
545 should affect the temperature dependence of the overall reaction as well (Levenspiel,  
546 1999). However, the difference in the experimentally determined E<sub>A</sub> for the  
547 unfractionated 30 N carbon (1.18 mm grain diameter) and the 0.5 mm size fraction of  
548 the 30N GAC was not significant (p < 0.001). This discrepancy can be explained by the  
549 limited resolution of the method used to determine k<sub>app</sub>, which in turn is used to  
550 calculate E<sub>A</sub>.

551

552 Figure 8: Impact of water temperature on the apparent reaction rate constant  
553 for MCA conversion at the GACs 30N, K835 and Centaur. Solid lines  
554 represent linear least-squares best fit and dashed lines represent the  
555 95% confidence band of the fitted regression. Errors for  $E_A$  represent  
556 the standard deviation of the slope of the linear regression ( $n = 2 - 6$ ,  
557 depending on the carbon). (n.d. = not determined)

558

## 559 **4 Summary and conclusions**

560 Using a fixed bed reactor system, the thorough investigation of apparent reaction rates  
561 of the removal of monochloramine at GAC surfaces showed a rapid decrease with  
562 ongoing exposure of the GAC to monochloramine. Quasi-stationary conditions were  
563 reached after less than 100 h, when the GAC had been exposed to MCA  
564 concentrations of 4.5 mg L<sup>-1</sup>.

565 When quasi-stationary conditions had been reached, film diffusion definitely is not rate-  
566 limiting. However, diffusion of monochloramine in the pores, for the GACs investigated,  
567 limits the overall reaction rate. It was shown that the overall apparent reaction including  
568 diffusion and chemical reaction is first-order. As diffusion limitation of a second-order  
569 chemical reaction will result in an apparent 1.5-order overall reaction, and diffusion  
570 limitation of a first-order reaction will result in an apparent first-order overall reaction  
571 rate, it is concluded that the reaction according to equation (2) is much faster than the  
572 reaction according to equation (1) or not taking place at all.

573 The dependence of the apparent overall reaction rate on temperature yielded  
574 activation energies  $E_A$ , according to the Arrhenius relation, in the range of 20 – 27 kJ  
575 mol<sup>-1</sup>. This additionally supports the conclusion that diffusion is rate-limiting as  
576 chemical reactions usually show activation energies of 50 kJ mol<sup>-1</sup> or higher, while  
577 temperature dependence of diffusion is equivalent to activation energies of about  
578 10 kJ mol<sup>-1</sup>.

579 Also, the Thiele modulus (3.0 – 42.1), derived from experiments with different grain  
580 sizes, indicates that the overall reaction rate is strongly controlled by diffusional mass  
581 transport in the pore system of the GACs considered. Conclusively, the pore size

582 distribution of the GACs was found to have a significant impact on the overall reaction  
583 rate constant of MCA removal in GAC filters.

584 Analysis of the impact of the pore size distribution indicates that pores  $<1.3$  nm, which  
585 is about 2.5 times the molecular diameter of MCA, are not accessible for the MCA–  
586 GAC reaction. Instead, the number of large mesopores is suggested to be of  
587 importance for a high overall reactivity of a GAC bed for MCA removal.

588 The use of smaller grain size fractions resulted in an increase of the overall reactivity  
589 of the GAC bed, which is attributed to a lower diffusion path length in the single GAC  
590 grains. If for example swimming pool water with a MCA concentration of  $0.2 \text{ mg L}^{-1}$  is  
591 treated in a GAC bed of 1 m height of the 30N GAC and of the largest grain size  
592 investigated (1.9 mm), the effluent concentration will be  $0.038 \text{ mg L}^{-1}$ . For the smallest  
593 grain size (0.3 mm) this effluent concentration can be achieved with a GAC bed of 0.56  
594 m.

595 However, it must be considered that the use of smaller GAC grain sizes for enhanced  
596 MCA removal in full-scale applications would bring about a considerably high pressure  
597 loss in fixed-bed GAC filters. For the example described above, using the empirical  
598 relationship described by Ergun (Ergun, 1952) and 1 m bed of 1.9 mm grains, the  
599 pressure loss will be 1.6 mbar. However, for a 0.56 m bed of the 0.3 mm grains, the  
600 pressure loss will 24 mbar, i.e. higher by a factor of 15.

601

602 From the shape of the conversion–time curves of the reaction it was concluded that  
603 the decrease of the reactivity of the GACs observed in the initial phase of the MCA–  
604 GAC reaction is due to an increasing pore diffusional resistance that developed with



605 increasing reaction time. The increasing diffusional control is explained by the fact that  
606 the reaction front starts at the outer surface of the grains and then moves towards the  
607 centre of the GAC grains.

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618 **Abbreviations**

619	EBCT	Empty bed contact time
620	FBR	Fixed-bed reactor
621	GAC	Granular activated carbon
622	MCA	Monochloramine
623	NDMA	N-Nitrosodimethylamine
624	PCM	Progressive Conversion Model
625	PSD	Pore size distribution
626	QSDFT	Quenched solid density functional theory
627	SCM	Shrinking-Core Model
628	SEM	Scanning Electron Microscopy

629

630 **Symbols**

631	$a_o$	bed volume specific outer surface area ( $m^2 m^{-3}$ )
632	$C_{bed,in}$	MCA in-flow concentrations of a GAC filter ( $mg L^{-1}$ as $Cl_2$ )
633	$C_{bed,out}$	MCA out-flow concentrations of a GAC filter ( $mg L^{-1}$ as $Cl_2$ )
634	$C_{stock}$	Concentration of the MCA stock solution ( $mg L^{-1}$ as $Cl_2$ )
635	$D_{bulk,MCA}$	Bulk diffusion coefficient of MCA in water ( $m^2 s^{-1}$ )
636	$D_{E,MCA}$	Effective diffusion coefficient of MCA in the pore system of a GAC grain
637		( $m^2 s^{-1}$ )
638	$d_{hy}$	Hydraulic diameter (m)

639	$d_{MCA}$	Minimum pore diameter of a GAC that is still accessible for an MCA molecule (nm)
640		
641	$d_{p,i}$	Pore diameter of pore size $i$ (nm)
642	$E_A$	Activation energy ( $J\ mol^{-1}$ )
643	$k_{app}$	Observable pseudo-first-order reaction rate constant ( $s^{-1}$ )
644	$k_{app,a}$	Pseudo-first-order reaction rate constant normalised by the specific outer surface $a_o$ of the GAC in the filter bed ( $m\ s^{-1}$ )
645		
646	$k_0$	Frequency factor ( $s^{-1}$ )
647	$k_f$	Mass transfer coefficient of MCA through the laminar film layer ( $m\ s^{-1}$ )
648	$L$	Diffusion path length within the adsorbent (m)
649	pH	pH value
650	$Q$	Volumetric fluid flow rate ( $L\ s^{-1}$ )
651	$Q_{bed}$	Volumetric fluid flow through a GAC bed ( $L\ s^{-1}$ )
652	$Q_{stock}$	Volumetric dosing rate of an MCA stock solution ( $L\ s^{-1}$ )
653		
654	$R_0$	Universal gas constant ( $R = 8.314\ J\ K^{-1}\ mol^{-1}$ )
655	$R_{sys}$	Time-dependent loss of MCA in the system without GAC filter ( $mg\ s^{-1}\ as\ Cl_2$ )
656		
657	$Sh$	Sherwood number (dimensionless)
658	$T$	Temperature (K)
659	$t_{EBCT}$	Empty bed contact time (EBCT) (s)
660	$V_{bed}$	Bed volume (L)
661	$v_{bed}$	Superficial filter velocity ( $m\ s^{-1}$ )

662  $V_{\text{sys}}$  Water volume in the FBR system (L)

663  $Z_{\text{bed}}$  Bed depth (m)

664

665 **Greek symbols**

666  $\varepsilon$  Molar absorption coefficient ( $\text{mol}^{-1} \text{cm}^{-1}$ )

667  $\eta$  Effectiveness factor of a heterogeneous reaction (dimensionless)

668  $\rho_{\text{stock}}(T)$  Temperature-dependent density of an MCA stock solution  
669 ( $\text{g L}^{-1}$ )

670  $\tau$  Tortuosity factor of a GAC (dimensionless)

671  $\Phi$  Thiele modulus of a heterogeneous reaction (dimensionless)

672

## 673 References

- 674 Aleghafouri, A., Mohsen-Nia, M., Mohajeri, A., Mahdyarfar, M., Asghari, M., 2012.  
675 Micropore size analysis of activated carbons using nitrogen, carbon dioxide and  
676 methane adsorption isotherms: Experimental and theoretical studies. *Adsorption*  
677 *Science and Technology* 30, 307-316.
- 678 ANSI/APSP-1, 2009. American National Standard for Water Quality in Public Pools  
679 and Spas. The Association of Pool and Spa Professionals. Alexandria.
- 680 Aoki, T., 1989. Continuous-Flow Method for Simultaneous Determination of  
681 Monochloramine, Dichloramine, and Free Chlorine - Application to a Water-  
682 Purification Plant. *Environmental Science and Technology* 23, 46-50.
- 683 Bashkova, S., Baker, F.S., Wu, X.X., Armstrong, T.R., Schwartz, V., 2007. Activated  
684 carbon catalyst for selective oxidation of hydrogen sulphide: On the influence of  
685 pore structure, surface characteristics, and catalytically-active nitrogen. *Carbon*,  
686 45, 1354-1363.
- 687 Bauer, R.C., Snoeyink, V.L., 1973. Reactions of Chloramines with Active Carbon.  
688 *Journal of the Water Pollution Control Federation* 45, 2290-2301.
- 689 Bernard, A., Carbonnelle, S., Michel, O., Higuët, S., de Burbure, C., Buchet, J.P.,  
690 Hermans C., Dumont, X., Doyle, I., 2003. Lung hyperpermeability and asthma  
691 prevalence in schoolchildren: Unexpected associations with the attendance at  
692 indoor chlorinated swimming pools. *Occupational and Environmental Medicine*  
693 60, 385-394.
- 694 Bernard, A., Nickmilder, M., Voisin, C., 2008. Outdoor swimming pools and the risks of  
695 asthma and allergies during adolescence. *European Respiratory Journal* 32, 979-  
696 988.
- 697 Blatchley, E.R., Cheng, M.M., 2010. Reaction Mechanism for Chlorination of Urea.  
698 *Environmental Science and Technology* 44, 8529-8534.
- 699 Consolazio, C.F., Johnson, R.E., Pecora, L.J., 1963. *Physiological Measurements of*  
700 *Metabolic Functions in Man*. McGraw-Hill, New York.
- 701 Crabtree, R.H., 2012. Resolving Heterogeneity Problems and Impurity Artifacts in  
702 Operationally Homogeneous Transition Metal Catalysts. *Chemical Reviews* 112,  
703 1536-1554.
- 704 Deborde, M., von Gunten, U., 2008. Reactions of chlorine with inorganic and organic  
705 compounds during water treatment - Kinetics and mechanisms: A critical review.  
706 *Water Research* 42, 13-51.
- 707 DIN 19643-2, 2012. Treatment of water of swimming pools and baths, Part 2:  
708 Combinations of process with fixed bed filters and precoat filters (in German:  
709 Aufbereitung von Schwimm- und Badebeckenwasser, Teil 2:  
710 Verfahrenskombinationen mit Festbett- und Anschwemmfiltern).  
711 Normenausschuss Wasserwesen (NAW) im DIN Deutsches Institut für Normung.  
712 Beuth Verlag, Berlin.
- 713 DIN EN ISO 7393-2, 2000. Water quality - Determination of free chlorine and total  
714 chlorine, Part 2: Colorimetric method using N,N-diethyl-1,4-phenylenediamine,  
715 for routine control purposes (in German: Wasserbeschaffenheit - Bestimmung  
716 von freiem Chlor und Gesamtchlor, Teil 2: Kolorimetrisches Verfahren mit

- 717 N,N-Diethyl-1,4-Phenylendiamin für Routinekontrollen). Normenausschuss  
718 Wasserwesen (NAW) im DIN Deutsches Institut für Normung. Beuth Verlag,  
719 Berlin.
- 720 Eichelsdörfer, D., Slovak, D., Dirnagl, D., Schmidt, K., 1975. Zur Reizwirkung  
721 (Konjunctivitis) von Chlor und Chloraminen im Schwimmbeckenwasser (in  
722 German). Vom Wasser 45, 17-28.
- 723 Ergun S., 1952. Fluid Flow through Packed Columns. Chemical Engineering Progress  
724 48, 89-94.
- 725 Ertl, G., Knözinger, H., Weitkamp, J., 1997. Handbook of heterogeneous catalysis.  
726 VCH Verlagsgesellschaft mbH, Weinheim.
- 727 Fairey, J.L., Speitel, G.E., Katz, L.E., 2006. Impact of natural organic matter on  
728 monochloramine reduction by granular activated carbon: The role of porosity and  
729 electrostatic surface properties. Environmental Science and Technology 40,  
730 4268-4273.
- 731 Fairey, J.L., Speitel, G.E., Katz, L.E., 2007. Monochloramine destruction by GAC -  
732 Effect of activated carbon type and source water characteristics. Journal of the  
733 American Water Works Association 99, 110-120.
- 734 Gnielinski, V., 1978. Gleichungen zur Berechnung des Wärme- und Stoffaustausches  
735 in durchströmten ruhenden Kugelschüttungen bei mittleren und großen  
736 Pecletzahlen (in German). Verfahrenstechnik 12, 363-366.
- 737 Hand, V.C., Margerum, D.W., 1983. Kinetics and Mechanisms of the Decomposition  
738 of Dichloramine in Aqueous-Solution. Inorganic Chemistry 22, 1449-1456.
- 739 Harriott, P., 2003. Chemical reactor design. Marcel Dekker, New York.
- 740 Jaguaribe, E.F., Medeiros, L.L., Barreto, M.C.S., Araujo, L.P., 2005. The performance  
741 of activated carbons from sugarcane bagasse, babassu, and coconut shells in  
742 removing residual chlorine. Brazilian Journal of Chemical Engineering 22, 41-47.
- 743 Kim, B.R., 1977. Analysis of batch and packed bed reactor models for the carbon-  
744 chloramine reactions. Doctoral thesis. University of Illinois at Urbana-Champaign,  
745 Illinois.
- 746 Kim, B.R., Snoeyink, V.L., Schmitz, R.A., 1978. Removal of dichloramine and ammonia  
747 by granular carbon. Journal of the Water Pollution Control Federation 50, 122-  
748 133.
- 749 Kim, B.R., Snoeyink, V.L., 1980. The Monochloramine-GAC Reaction in Adsorption  
750 Systems. Journal of the American Water Works Association 72, 488-490.
- 751 Komorita, J.D., Snoeyink, V.L., 1985. Monochloramine Removal from Water by  
752 Activated Carbon. Journal of the American Water Works Association 77, 60-60.
- 753 Levenspiel, O., 1999. Chemical reaction engineering, Third Edition. John Wiley &  
754 Sons, New York.
- 755 Mangun, C.L., Braatz, R.D., Economy, J., Hall, A.J., 1999. Fixed bed adsorption of  
756 acetone and ammonia onto oxidized activated carbon fibers. Industrial and  
757 Engineering Chemistry Research 38, 3499-3504.

- 758 Maia, F., Silva, R., Jarrais, B., Silva, A.R., Freire, C., Pereira, M.F.R., Figueiredo, J.L.,  
759 2008. Pore tuned activated carbons as supports for an enantioselective  
760 molecular catalyst. *Journal of Colloid and Interface Science* 328, 314-323.
- 761 McCallum, C.L., Bandosz, T.J., McGrother, S.C., Muller, E.A., Gubbins, K.E., 1999. A  
762 molecular model for adsorption of water on activated carbon: Comparison of  
763 simulation and experiment. *Langmuir* 15, 533-544.
- 764 Qiang, Z.M., Adams, C.D., 2004. Determination of monochloramine formation rate  
765 constants with stopped-flow spectrophotometry. *Environmental Science and  
766 Technology* 38, 1435-1444.
- 767 Radian, A., Carmeli, M., Zadaka-Amir, D., Nir, S., Wakshal, E., Mishael, Y.G., 2011.  
768 Enhanced removal of humic acid from water by micelle-montmorillonite  
769 composites: Comparison to granulated activated carbon. *Applied Clay Science*  
770 54, 258–263.
- 771 Sayle, R.A., Milnerwhite, E.J., 1995. Rasmol - Biomolecular Graphics for All. *Trends in  
772 Biochemical Sciences* 20, 374-376.
- 773 Scaramelli, A.B., Digiano, F.A., 1977. Effect of sorbed organics on efficiency of  
774 ammonia removal by chloramine-carbon surface-reactions. *Journal of the Water  
775 Pollution Control Federation* 49, 693-705.
- 776 Schreiber, B., Brinkmann, T., Schmalz, V., Worch, E., 2005. Adsorption of dissolved  
777 organic matter onto activated carbon - the influence of temperature, absorption  
778 wavelength, and molecular size. *Water Research* 39, 3449-3456.
- 779 Schreiber, I.M., Mitch, W.A., 2005. Influence of the order of reagent addition on NDMA  
780 formation during chloramination. *Environmental Science and Technology* 39,  
781 3811-3818.
- 782 Schreiber, I.M., Mitch, W.A., 2006. Nitrosamine formation pathway revisited: The  
783 importance of chloramine speciation and dissolved oxygen. *Environmental  
784 Science and Technology* 40, 6007-6014.
- 785 Schurter, L.M., Bachelor, P.P., Margerum, D.W., 1995. Nonmetal Redox Kinetics -  
786 Monochloramine, Dichloramine, and Trichloramine Reactions with Cyanide Ion.  
787 *Environmental Science and Technology* 29, 1127-1134.
- 788 Sparks, D.L., 1989. *Kinetics of Soil Chemical Processes*. Academic Press, New York.
- 789 Suidan, M.T., Snoeyink, V.L., Schmitz, R.A., 1977. Reduction of Aqueous Free  
790 Chlorine with Granular Activated Carbon, pH and Temperature Effects.  
791 *Environmental Science and Technology* 11, 785-789.
- 792 Walse, S.S., Mitch, W.A., 2008. Nitrosamine carcinogens also swim in chlorinated  
793 pools. *Environmental Science and Technology* 42, 1032-1037.
- 794 Weaver, W.A., Li, J., Wen, Y.L., Johnston, J., Blatchley, M.R., Blatchley, E.R., 2009.  
795 Volatile disinfection by-product analysis from chlorinated indoor swimming pools.  
796 *Water Research* 43, 3308-3318.
- 797 Williamson, J.E., Bazaire, K.E., Geankoplis, C.J., 1963. Liquid-Phase Mass Transfer  
798 at Low Reynolds Numbers. *Industrial and Engineering Chemistry Fundamentals*  
799 2, 126-129.
- 800 Wilke, C.R., P. Chang, 1955. Correlation of Diffusion Coefficients in Dilute Solutions.  
801 *American Institute of Chemical Engineers Journal* 1 264-270.



- 802 Worch, E., 2008. Fixed-bed adsorption in drinking water treatment: a critical review on  
803 models and parameter estimation. *Journal of Water Supply Research and*  
804 *Technology-Aqua* 57, 171-183.
- 805 Worch, E., 2012. *Adsorption Technology in Water Treatment: Fundamentals,*  
806 *Processes, and Modeling.* Walter de Gruyter GmbH & Co. KG, Berlin/Boston.
- 807 Zwiener, C., Richardson, S.D., De Marini, D.M., Grummt, T., Glauner, T., Frimmel,  
808 F.H., 2007. Drowning in disinfection byproducts? Assessing swimming pool  
809 water. *Environmental Science and Technology* 41, 363-372.
- 810

811 **Tables**

812

813 Table 1: Operation conditions of the FBR experiments.

Type of study	GACs	Parameters varied	Parameters kept constant
Verification of the kinetic first-order approach	K835 ( <i>unfractionated</i> )	$C_{bed, in} = 0.5 - 9 \text{ mg L}^{-1}$ (as $\text{Cl}_2$ )	$v_{bed} = 44.1 \text{ m h}^{-1}$ $T = 30 \text{ }^\circ\text{C}$ $V_{bed} = 32 \text{ mL}$ $h_{bed} = 3.5 \text{ cm}$ $t_{EBCT} = 2.88 \text{ s}$
Extra-particle mass transfer	K835, 30N ( <i>unfractionated</i> )	$v_{bed} = 8.3 - 132.2 \text{ m h}^{-1}$ $V_{bed} = 6 - 96 \text{ mL}$ $h_{bed} = 0.6 - 94 \text{ cm}$	$T = 30 \text{ }^\circ\text{C}$ $C_{bed, in} = 4.5 \text{ mg L}^{-1}$ (as $\text{Cl}_2$ ) $t_{EBCT} = 2.88 \text{ s}$
Impact of grain size	30N ( <i>size fractions and unfractionated</i> )	$d_{grain} = 0.30 - 1.9 \text{ mm}$	$v_{bed} = 44.1 \text{ m h}^{-1}$ $T = 30 \text{ }^\circ\text{C}$ $C_{bed, in} = 4.5 \text{ mg L}^{-1}$ (as $\text{Cl}_2$ ) $V_{bed} = 32 \text{ mL}$ $h_{bed} = 3.5 \text{ cm}$ $t_{EBCT} = 2.88 \text{ s}$
Temperature dependence	30N, K835, Centaur ( <i>unfractionated</i> )	$T = 12 - 45 \text{ }^\circ\text{C}$	$v_{bed} = 44.1 \text{ m h}^{-1}$ $C_{bed, in} = 4.5 \text{ mg L}^{-1}$ (as $\text{Cl}_2$ ) $V_{bed} = 32 \text{ mL}$ $h_{bed} = 3.5 \text{ cm}$ $t_{EBCT} = 2.88 \text{ s}$
Conversion dependence and impact of pore size distribution	30N, K835, Centaur and 100058 ( <i>0.57 mm size fraction</i> )	-	$v_{bed} = 44.1 \text{ m h}^{-1}$ $T = 30 \text{ }^\circ\text{C}$ $C_{bed, in} = 4.5 \text{ mg L}^{-1}$ (as $\text{Cl}_2$ ) $V_{bed} = 32 \text{ mL}$ $h_{bed} = 3.5 \text{ cm}$ $t_{EBCT} = 2.88 \text{ s}$

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816

817 Table 2: Physical characterisation of the fresh unfractionated GACs 30N, K835, Centaur and 100058.

GAC type	Hydraulic diameter $d_{hy}$ , in mm	Outer surface $a_o$ in $m^2 m^{-3}$ <sup>a</sup>	Inner surface $a_{BET}$ in $m^2 g^{-1}$ <sup>a,b</sup>	Specific pore volume in $cm^3 g^{-1}$ (%)				
				Total	Micropores	Mesopores (total)	Mesopores >7 nm <sup>d</sup>	Macro-pores
30N	1.18	5444	1105 (900 <sup>e</sup> )	0.522	0.411 (79)	0.092 (18)	0.003 (0.6)	0.019 (4)
K835	1.39	4480	1073 (1170 <sup>f</sup> )	0.446	0.419 (94)	0.020 (5)	0.001 (0.2)	0.007 (2)
Centaur	1.00	5532	895 (815 <sup>g</sup> )	0.408	0.353 (87)	0.046 (11)	0.015 (3.7)	0.009 (2)
100058	0.55	17418	1291 (1350 <sup>c</sup> )	0.605	0.496 (82)	0.102 (17)	0.013 (2.1)	0.007 (1)

818 <sup>a</sup>... All pores of the activated GACs were potentially accessible for the MCA conversion reaction (differences in BET surfaces between granular GACs and their powdered  
819 counterparts were <5%).

820 <sup>b</sup>... Values in brackets represent data as found in the literature.

821 <sup>c</sup>... As reported by the manufacturer.

822 <sup>d</sup>... As derived from the PSD.

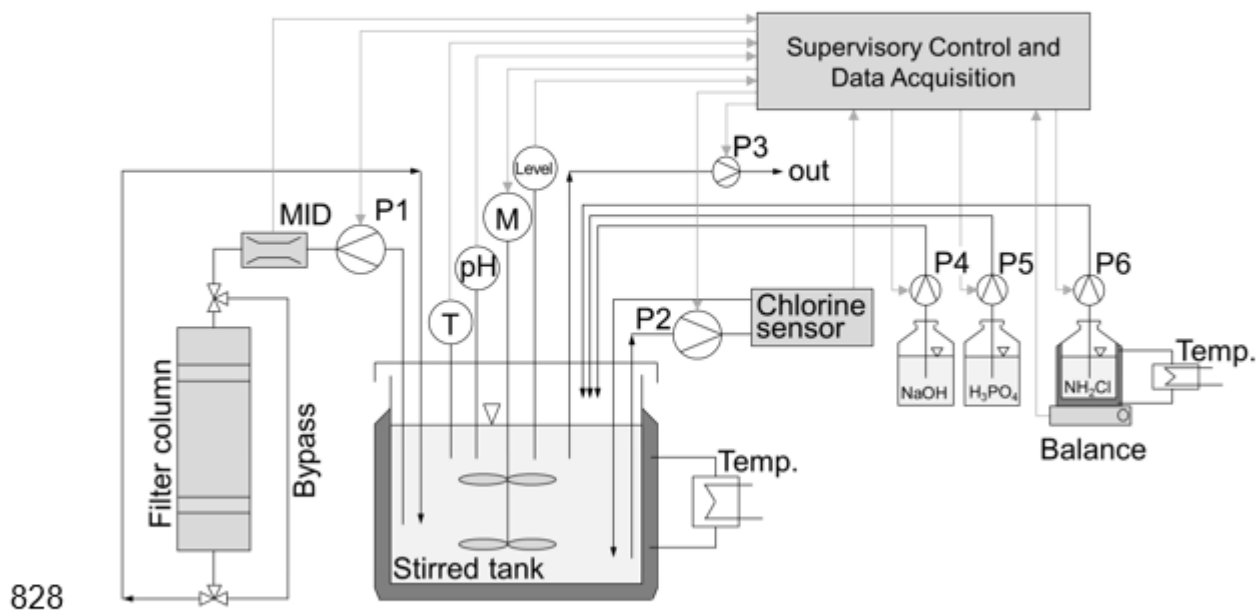
823 <sup>e</sup>... Radian et al., 2011.

824 <sup>f</sup>... Aleghafouri et al., 2012.

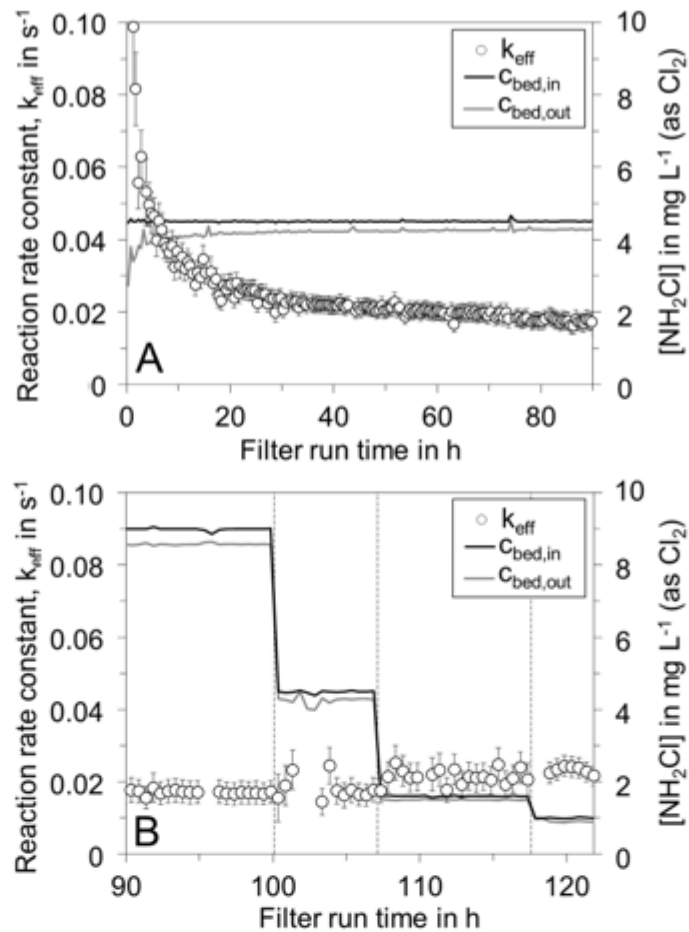
825 <sup>g</sup>... Bashkova et al., 2007.

826

827 **Figures**



829 Figure 1: Scheme of the laboratory-scale FBR system.



830

831 Figure 2:

Reaction rate constant  $k_{\text{app}}$  for MCA removal at the K835 GAC at

832

$C_{\text{bed,in}}$  of  $4.5 \text{ mg L}^{-1}$  (as  $\text{Cl}_2$ ) until stationary conditions were reached

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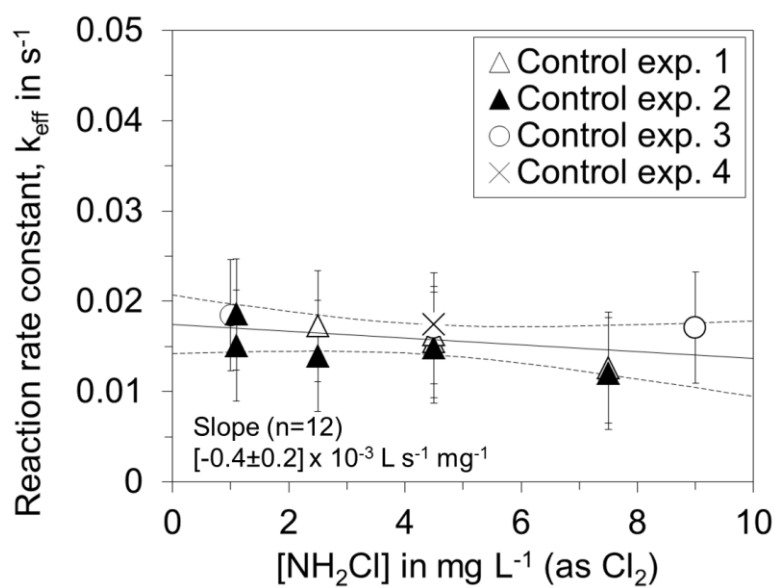
(A) and at three different inflow concentrations ( $9.0$ ,  $4.5$ ,  $1.6$  and  $0.9$

834

$\text{mg L}^{-1}$  (as  $\text{Cl}_2$ )) (B). Error bars represent the standard deviation of

835

$k_{\text{app}}$  as determined by Gaussian error propagation ( $n = 50$ ).



836

837 Figure 3:

Effect of MCA inflow concentration on the apparent reaction rate

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constant for MCA conversion at the altered unfractionated K835

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GAC. The solid line represents the linear least-squares best fit and

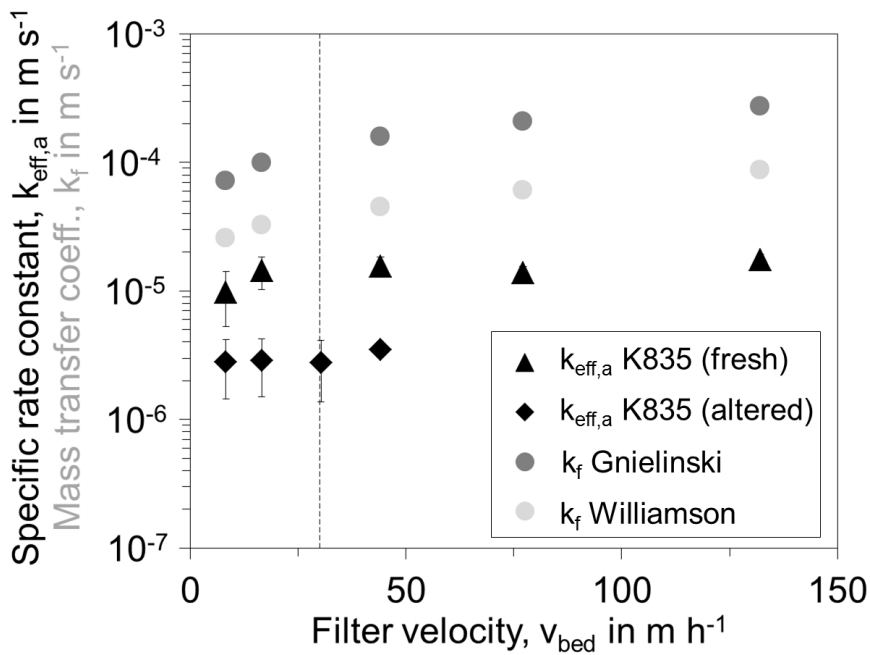
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dashed lines represent the 95% confidence band of the regression.

841

Errors bars represent the repeatability standard deviation (n = 63).

842



843

844

Figure 4:

Impact of filter velocity on the experimentally determined apparent surface related reaction rate constant  $k_{app,a}$  for MCA conversion and the theoretical mass transfer coefficient  $k_f$  of the unfractonated K835 GAC. Error bars represent the 95% confidence interval. Mass transfer coefficients  $k_f$  were calculated using a mean grain diameter  $d_{hy,K835}$  of 1.39 mm. The dashed vertical line represents the recommended filter velocity for GAC filters in swimming pool water treatment in Germany ( $\leq 30$  m h<sup>-1</sup>) (DIN 19643-2, 2012).

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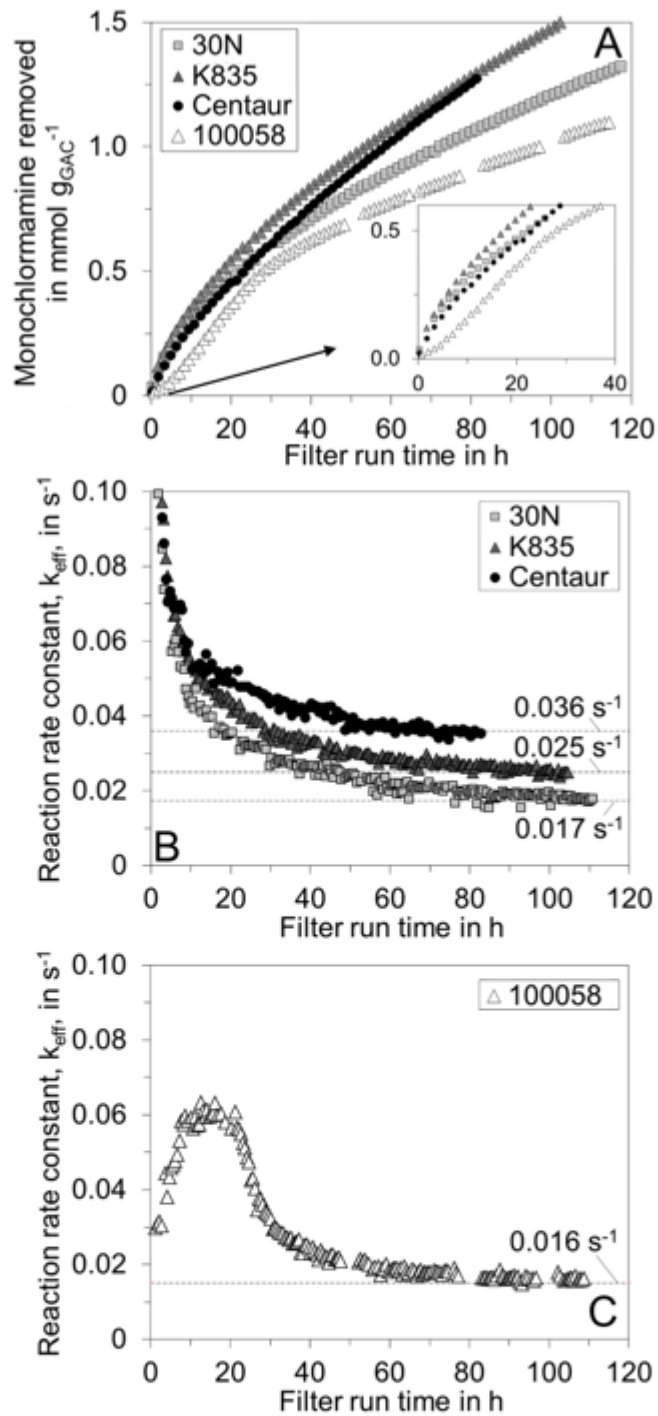
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853 Figure 5:

Removed MCA (A) and bed volume based apparent reaction rate

854

constant  $k_{\text{app}}$  for MCA removal (B, C) over the filter run time. All

855

GACs had the same grain size of  $\sim 0.55 \text{ mm}$  ( $0.57 \text{ mm}$  for the 100058

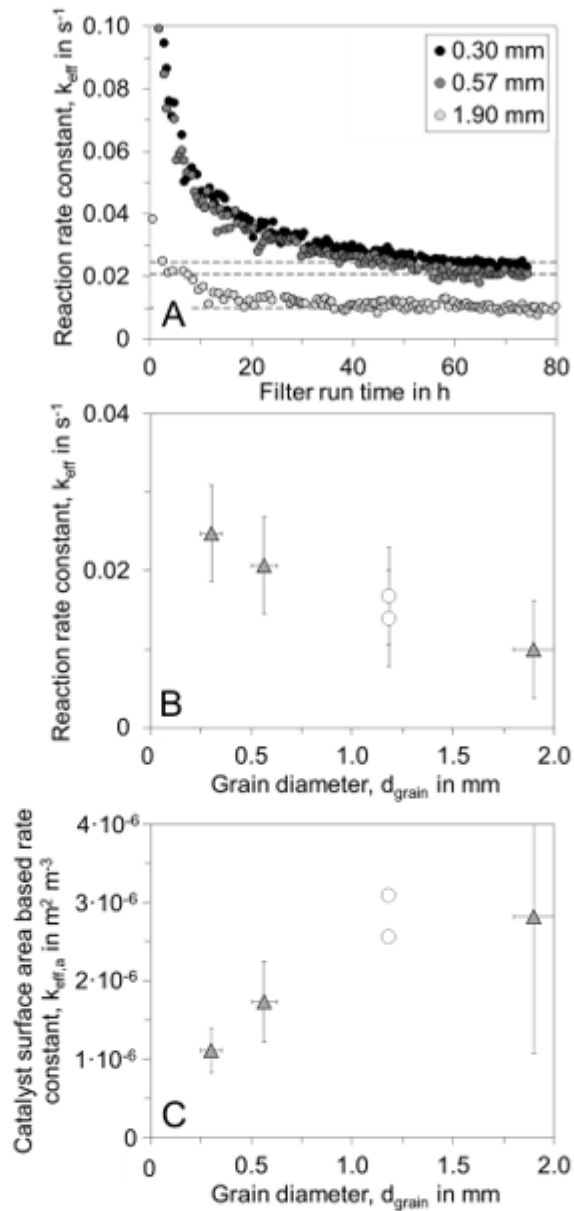
856

GAC). Dashed horizontal lines in (B) represent the GAC reactivity

857

when stationary conditions were reached.





858

859 Figure 6:

(A) Reaction rate constant  $k_{app}$  over the filter runtime for different

860

grain size fractions of the 30N GAC. Dashed horizontal lines

861

represent the level of GAC reactivity at stationary conditions. (B)  $k_{app}$

862

and (C)  $k_{app,a}$  at stationary conditions as function of the grain

863

diameter. Open circle symbols in (B) and (C) represent  $k_{app}$  for the

864

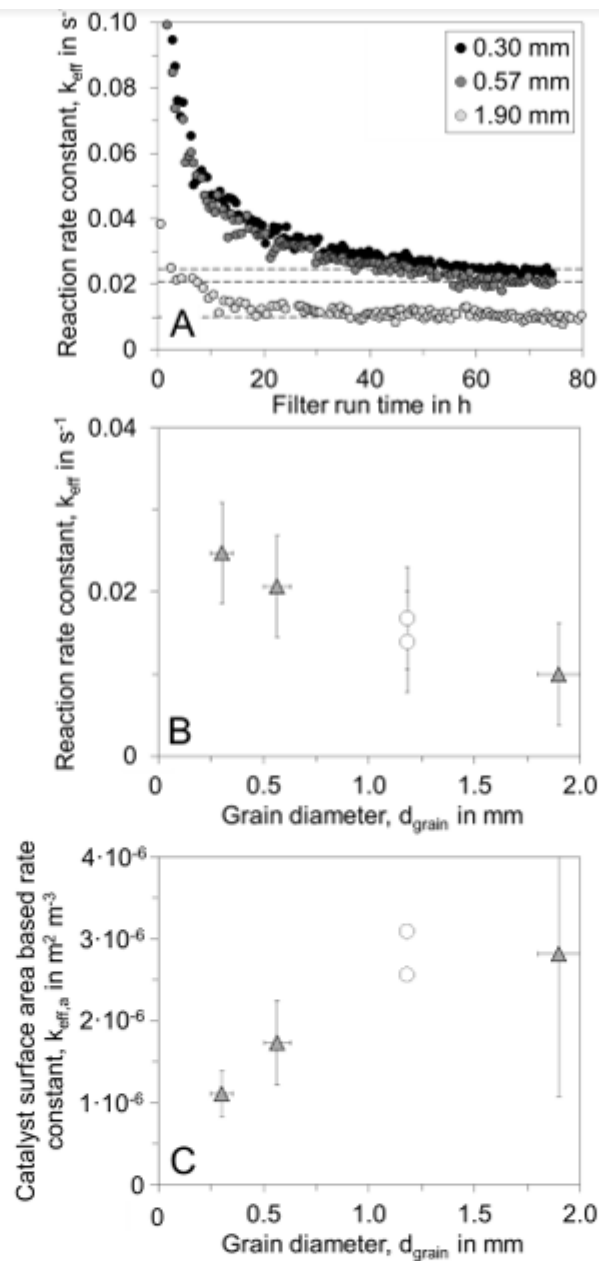
unfractionated 30N GAC using  $d_{hy,30N} = 1.18$  mm as representative

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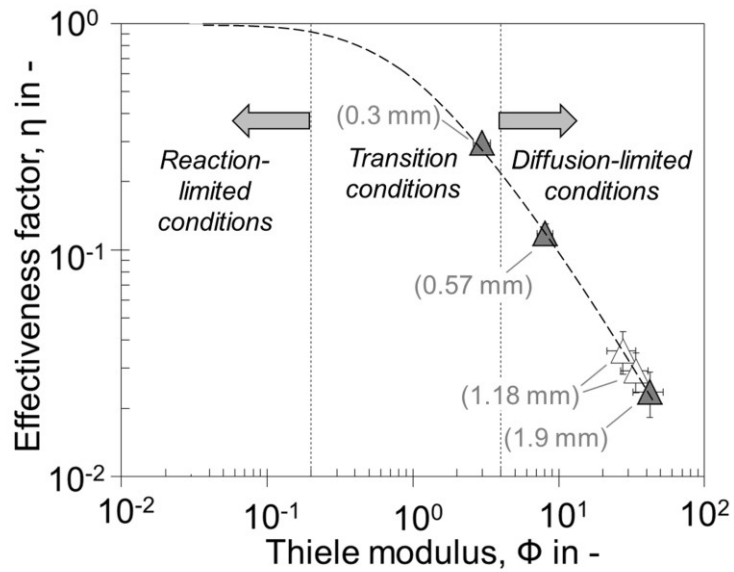
grain size (data not shown in A). Error bars in (B) and (C) represent

866

the 95% confidence intervals.



867  
 868 Figure 6: (A) Reaction rate constant  $k_{\text{app}}$  over the filter runtime for different  
 869 grain size fractions of the 30N GAC. Dashed horizontal lines  
 870 represent the level of GAC reactivity at stationary conditions. (B)  $k_{\text{app}}$   
 871 and (C)  $k_{\text{app},a}$  at stationary conditions as function of the grain  
 872 diameter. Open circle symbols in (B) and (C) represent  $k_{\text{app}}$  for the  
 873 unfractionated 30N GAC using  $d_{\text{hy},30\text{N}} = 1.18$  mm as representative  
 874 grain size (data not shown in A). Error bars in (B) and (C) represent  
 875 the 95% confidence intervals.



876

877 Figure 7:

Effectiveness factor  $\eta$  of the MCA-GAC reaction at stationary conditions for different grain size fractions of the 30N GAC as

878

function of the Thiele modulus  $\Phi$ . Open triangle symbols represent

879

880

$k_{app}$  for the unfractionated 30N GAC using  $d_{hy,30N} = 1.18$  mm. Filled

881

triangle symbols represent  $k_{app}$  for the fractionated 30N GAC. The

882

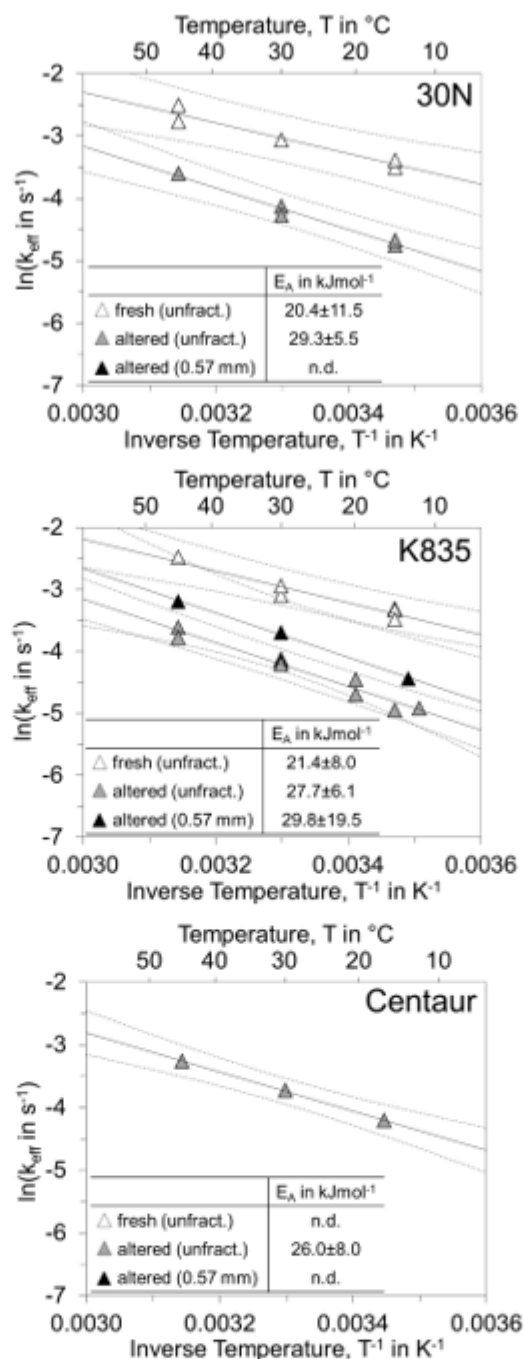
dashed black line is for orientation only and represents the expected

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relationship between  $\Phi$  and  $\eta$  as discussed elsewhere (Ertl et al.,

884

1997). Error bars represent the 95% confidence intervals.



885  
 886 Figure 8: Impact of water temperature on the apparent reaction rate constant  
 887 for MCA conversion at the GACs 30N, K835 and Centaur. Solid lines  
 888 represent linear least-squares best fit and dashed lines represent the  
 889 95% confidence band of the regression. Errors for  $E_A$  represent the  
 890 standard error of the slope of the linear regression ( $n = 3 - 6$ ,  
 891 depending on the carbon). (n.d. = not determined)

1 **PORE DIFFUSION LIMITS REMOVAL OF MONOCHLORAMINE IN**  
2 **TREATMENT OF SWIMMING POOL WATER USING GRANULAR ACTIVATED**  
3 **CARBON**

4 **- Supplementary Material -**

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11

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14

15 **This PDF file includes:**

16 3 Sections, 2 Figures, 1 Table and 5 Equations for further information.

17 **A Additional information on the methods used for physical**  
18 **GAC characterisation**

19 **A.1 Specific outer surface area**

20 The specific outer surface area  $a_0$  of the GAC bed was calculated based on the grain  
21 size distribution of the unfractionated GACs as follows:

$$a_0 = \frac{A_{\text{bed}}}{V_{\text{bed}}} = 6 \cdot \frac{\rho_{\text{bed}}}{\rho_{\text{grain}}} \cdot \sum_i \frac{q_i}{d_{\text{grain},i}} \cdot \frac{1}{\psi} \quad (1)$$

22 Here  $\rho_{\text{bed}}$  and  $\rho_{\text{grain}}$  are the dry bed density and the dry density of the GAC grains,  $q_i$   
23 is the mass fraction of the corresponding mean grain size  $d_{\text{grain},i}$  as derived from the  
24 grain size distribution, and  $\psi$  is the dimensionless non-sphericity correction factor,  
25 introduced in order to respect the deviation in shape of the GAC grains from a  
26 spherical form ( $\psi = 0,75$  for GACs (Hawksley, 1951)). Grain size distributions of all  
27 GACs were analysed according to a German standard (DIN 66165-1, 1987) using a  
28 sieve tower (AS 200, Retsch, Germany). Results are given in section B.1 of the  
29 Supplementary information.

30 **A.2 Bed and grain densities**

31  $\rho_{\text{bed}}$  and  $\rho_{\text{grain}}$  of the four GACs investigated were determined as follows: A volume of  
32 25–35 mL of the respective GAC was soaked in a beaker with roughly 50 mL of  
33 deionised water. Vacuum was applied to remove air from the pores and to fill them  
34 with water. The grains were removed from the water and spread on a metal sieve  
35 gently shaken. Thus the GAC grain pores remained filled with water.

36 25 mL of deionised water ( $V_w$ ) were placed in a 50 mL measuring cylinder and the  
37 mass noted. Then GAC grains with water-filled pores were slowly transferred from  
38 the metal sieve to the cylinder such that a bed of 15 mL ( $V_{bed}$ ) was formed, after  
39 carefully consolidating the bed by manually tamping the measuring cylinder 15 times  
40 on a plate. The total volume of water with GAC grains in the measuring cylinder was  
41 denoted as  $V_{w+GAC}$ . The mass of the water and the grains was noted as  $m_{w+GAC}$ .

42 The volume of the wet GAC grains was:

$$V_{GAC(wet)} = V_{w+GAC} - V_w \quad (2)$$

43 Subsequently, the GAC suspension was taken from the cylinder and spread over a  
44 filter paper of mass  $m_{fil}$ . The filter paper with GACs was dried at 110 °C for ~24 h until  
45 the weight of the filter paper reached constancy and no water remained in the pores  
46 of the GAC grains. The mass of the dry filter paper with GAC grains was denoted as  
47  $m_{fil+GAC}$ . The mass of the dry GAC grains  $m_{grain}$  was calculated as the difference  
48 between  $m_{fil+GAC}$  and  $m_{fil}$ .

49 According to the procedure, then bed density (dry)  $\rho_{bed}$  of the GAC bed is given by

$$\rho_{bed} = \frac{m_{grain}}{V_{bed}} \quad (3)$$

50 The density of the GAC grains (dry)  $\rho_{grain}$  is given by

$$\rho_{grain} = \frac{m_{grain}}{V_{GAC(wet)}} \quad (4)$$

51  $\rho_{bed}$  and  $\rho_{grain}$  are given as average obtained from three determinations.

### 52 **A.3 Hydraulic diameter**

53 The representative hydraulic grain size  $d_{hy}$  of the unfractionated GACs was  
54 calculated from the grain size distribution as follows (Kozeny, 1927):

$$d_{hy} = \frac{1}{\sum \left( \frac{q_i}{d_{grain,i}} \right)} \quad (5)$$

55

#### 56 **A.4 Internal surface area and pore volumes**

57 The internal surface area, the pore volume of micropores and mesopores, the pore  
58 size distribution (PSD) and the tortuosity of the fresh unfractionated GACs were  
59 determined from nitrogen adsorption/desorption isotherms. Nitrogen  
60 adsorption/desorption isotherms were determined in a relative pressure range ( $p/p_0$ )  
61 of  $10^{-6}$  to 1 at 77 K using an automated gas-sorption apparatus (Autosorb-1C,  
62 Quantachrome, Germany). The internal surface area of the GACs was determined  
63 using the BET equations (Brunauer et al., 1938). The total pore volume was  
64 calculated from the nitrogen sorption data at  $p/p_0$  of  $\sim 0.98$  while the micropore  
65 volume was determined using the Dubinin-Radushkevich equation at  $p/p_0$  of  
66  $10^{-6}$  -  $10^{-1}$ . For calculations, ASiQWin Software (Version 3.0, Quantachrome  
67 Instruments) was used. The sum of mesopore (2 – 50 nm) and macropore (>50 nm)  
68 volume was calculated by subtracting the micropore (<2 nm) volume from the total  
69 pore volume.  $N_2$  adsorption-desorption isotherms were further used to estimate the  
70 tortuosity of the GACs using the CSTM-model as described previously (Salmas and  
71 Androutsopoulos, 2001) (see Table 2).

72 The surface area and pore size distribution (PSD) of the GACs were determined  
73 using the quenched solid density functional theory (QSDFT) assuming graphite



74 material with pores of slit-like shape (Neimark et al., 2009). The validity of this  
75 assumption is discussed in Section 3.2. The proportion of macropores (>50 nm) was  
76 determined by mercury intrusion porosimetry using a Porosimeter 2000 apparatus  
77 (Carlo Erba Instruments, Milan, Italy). The intrusion experiments were performed in a  
78 pressure range of 0.4 to 200 MPa.

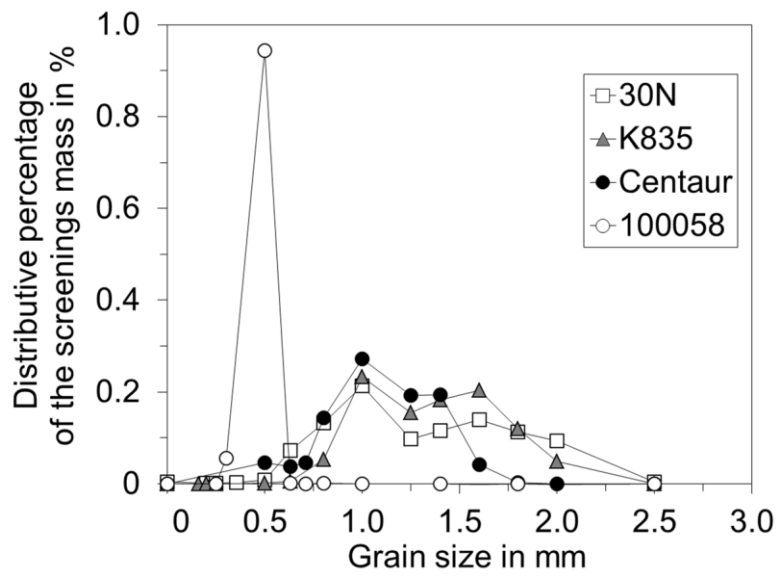
79 Images of the 100058 GAC were made with a High Resolution Scanning Electron  
80 Microscope (HRSEM) (FEI Nova NanoSEM, 5 kV) and were taken at four different  
81 positions across the grain radius. In total, six close-up HRSEM images of each  
82 position across the diameter were taken. The pore size distribution of the 100058  
83 GAC across the grain diameter was analysed by image processing of the close-up  
84 HRSEM-images as described previously by the authors (Skibinski et al., 2016). Due  
85 to the limited resolution of the HRSEM-images, only pores >10 nm could be  
86 analysed.

## 87 **B Additional results of the physical GAC characterisation**

### 88 **B.1 Grain Size Distributions**

89 The distributive grain size distribution is shown in Figure B.1 and cumulative grain  
90 size distribution of the GACs is shown in Figure B.2.

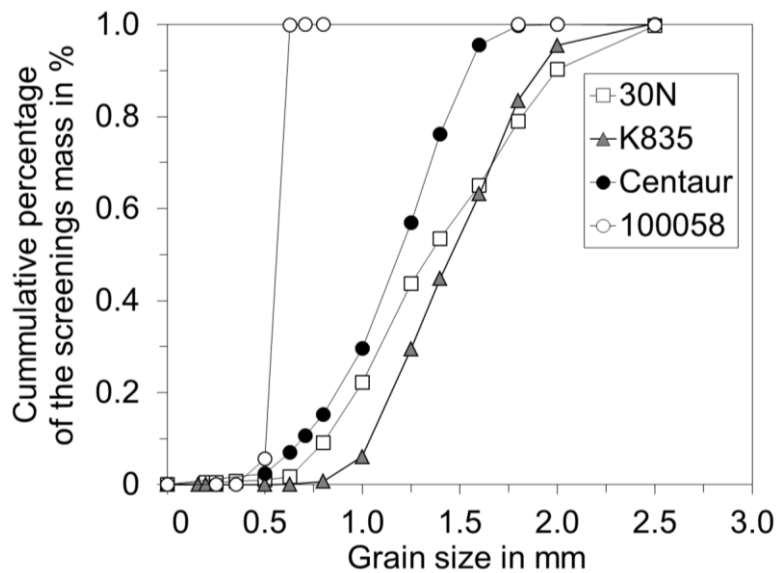
91 The specific outer surface areas, densities and hydraulic diameters are summarised  
92 in Table 1.



93

94 Figure B.1: Distributive grain size distribution of the fresh unfractionated GACs  
 95 30N, K835, Centaur and 100058.

96



97

98 Figure B.2: Cumulative grain size distribution of the fresh unfractionated GACs  
 99 30N, K835, Centaur and 100058.

100 Table B.1: Hydraulic diameter  $d_{hy}$ , bulk density (dry)  $\rho_{bed}$ , density of the GAC  
 101 grains  $\rho_{grain}$  and specific outer surface area  $a_o$  of the fresh  
 102 unfractionated GACs 30N, K835, Centaur and 100058.

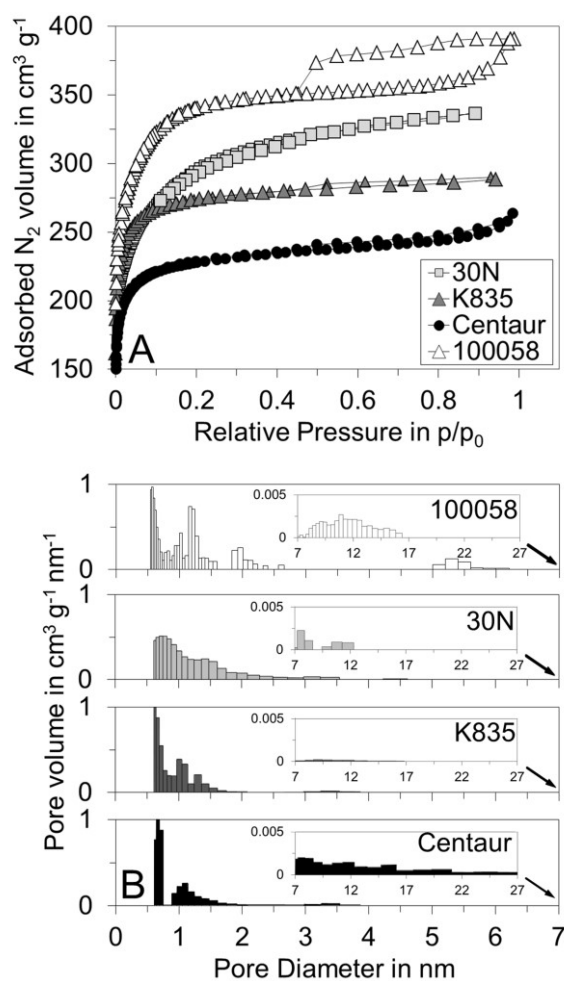
<b>GAC type</b>	$d_{hy}$ in mm	$\rho_{bed}$ in g mL <sup>-1</sup>	$\rho_{grain}$ in g mL <sup>-1</sup>	$a_o$ in m <sup>2</sup> m <sup>-3</sup>
30N	1.18	0.47	0.56	5444
K835	1.39	0.59	0.76	4480
Centaur	1.00	0.50	0.77	5532
100058	0.55	0.72	0.60	17418

103

## 104 B.2 Pore Size Distributions

105 Figure B.3 shows the N<sub>2</sub> adsorption-desorption isotherms of the fresh unfractionated  
 106 GACs and the corresponding pore size distribution (PSD). The N<sub>2</sub> adsorption-  
 107 desorption isotherms were type I according to the Brunauer-Deming-Deming-Teller  
 108 (BDDT) classification for all GACs considered. The total internal surface area (BET  
 109 surface) for all GACs ranged between 895 and 1291 m<sup>2</sup> g<sub>GAC</sub><sup>-1</sup> (see Table 2) and was  
 110 in accordance with data found in the literature. The sharp increase of N<sub>2</sub> adsorption in  
 111 the low pressure region indicates the presence of plentiful micropores (Brunauer et  
 112 al., 1940). The percentage of micropores with regard to the total pore volume is  
 113 shown in Table 2. The relative proportion of mesopores was the highest for 100058  
 114 (18 %) and Centaur (17 %) GACs. The hysteresis of nitrogen physisorption (the  
 115 difference between adsorption and desorption isotherms) for the 30N and K835  
 116 GACs was almost zero. The fact that adsorption and desorption isotherms branch in  
 117 parallel to each other and almost horizontally is associated with the presence of  
 118 narrow pores of slit-like shape, which are commonly found for activated GACs (Hu et  
 119 al., 2001).

120 In the micropore region (<2 nm), the Centaur, K835 and 100058 GAC showed PSD  
121 peaks at the smallest detectable pore size of 0.6 nm and at ~1.1 nm. Both Centaur  
122 and K835 GACs showed a PSD peak at 1.1 nm, while the 30N GAC showed peaks  
123 at 0.72 and 1.5 nm. The characteristic pore size distribution of the Centaur GAC is in  
124 agreement with data reported previously (Bashkova et al., 2007). Both the Centaur  
125 and the 100058 GACs contained a significant proportion of large mesopores of  
126 >7 nm (0.015 cm<sup>3</sup> g<sup>-1</sup> (Centaur) and 0.013 cm<sup>3</sup> g<sup>-1</sup> (100058)).  
127 Mercury intrusion porosimetry measurements showed that the proportion of  
128 macropores (macropore volume related to the total pore volume) was very low and  
129 ranged between 1.6 and 3.6 %.



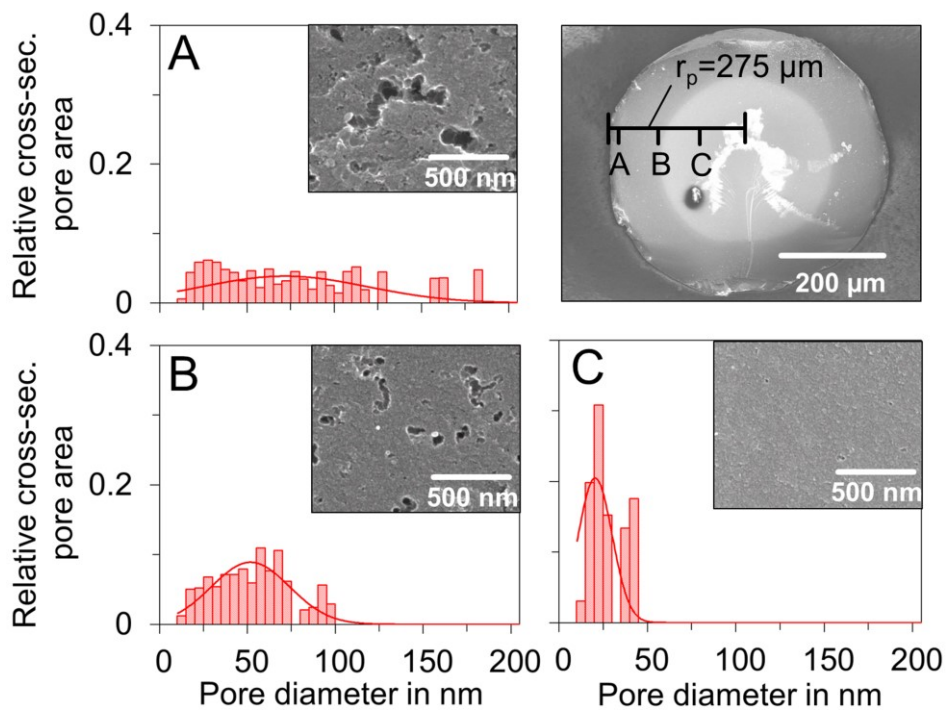
130  
 131 Figure B.3: N<sub>2</sub> adsorption and desorption isotherms at 77 K (A) and pore size  
 132 distribution of the unfractionated fresh GACs (B) (inset represents the  
 133 distribution of large mesopores of >7 nm).

134 **B.3 Characterisation of the ordered pore size distribution of the 100058 GAC**

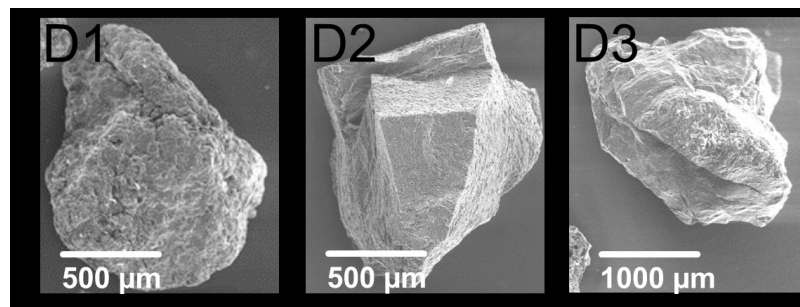
135 Examples of close-up HRSEM images of the 100058 GAC taken at different positions  
 136 across the radius of a cross-section are presented in Figure B.4. It revealed that the  
 137 obvious textural change across the diameter of the grain (Figure B.4(A)) is  
 138 accompanied by a change in the pore size distribution (Figure B.4(B-D)). From the  
 139 relative pore size distribution it becomes obvious that plenty of mesopores are  
 140 present at the outer region of the grain, while their relative proportion decreases

141 towards the centre of the grain. Near of the centre of a grain (position D), no  
142 macropores were present and mesopores revealed a mean diameter of ~25 nm.

143 No structural changes were revealed from the HRSEM images of the K835 (D1), 30N  
144 (D2) and Centaur (D3) carbon, supporting the assumption of an equal pore size  
145 distribution proposed for the three GACs by earlier studies (Radian et al., 2011;  
146 Aleghafouri et al., 2012; Bashkova et al., 2007).



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Figure B.4: HRSEM image of a cross-section of the 100058 GAC at a zoom of 150 (top right) and at different positions across the diameter with the corresponding relative cross-sectional pore area distribution at a zoom of 200,000 (A-C). Further, HRSEM images of the K835 (D1), 30N (D2) and Centaur (D3) GACs are shown at a zoom of 50 (D1, D2) and 21 (D3). Solid lines in (A-C) represent the log-normal distribution.

157 **C Additional information on the GAC's conversion–time**  
158 **behaviour**

159 The difference in conversion–time behaviour between the microporous GACs (K835,  
160 30N and Centaur) and the 100058 GAC can be explained by the unusual pore size  
161 distribution of the latter, as follows:

162 (i) Due to the high proportion of mesopores near to the external surface of the  
163 grains of the 100058 GAC (Figure 3(B)), the effective diffusion coefficient  
164 of monochloramine ( $D_{E,MCA}$ ) in this area is high and the overall reaction is  
165 controlled by the intrinsic chemical reaction at the beginning.

166 (ii) The observed decrease in mean pore size towards the centre of the grains  
167 of the GAC 100058 will lead to an increase of the accessible pore area  
168 reached by the reaction front when moving towards the centre. As a result,  
169 the reactivity of the 100058 GAC increases during the first ~20 min of the  
170 reaction.

171 (iii) With on-going reaction time, the reaction front moves further towards the  
172 grain's centre and diffusion paths of monochloramine molecules increase  
173 until a shift in the reaction controlling mechanism from chemical control to  
174 diffusional control occurs. This shift results in a decrease in reactivity,  
175 which in turn leads to a sigmoid conversion time curve (Levenspiel, 1999),  
176 as was found for the 100058 GAC.

177 In contrast to the 100058 GAC, the pore structure of the conventional GACs (30N,  
178 K835, Centaur) is assumed to be microporous throughout the entire particle. Thus,



179 for those GACs the overall process is controlled by diffusion right from the start of the  
180 reaction, leading to hyperbolic conversion–time curves (Levenspiel, 1999).

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## 182 **References**

183 Aleghafouri, A., Mohsen-Nia, M., Mohajeri, A., Mahdyarfar, M., Asghari, M., 2012.  
184 Micropore size analysis of activated carbons using nitrogen, carbon dioxide and  
185 methane adsorption isotherms: Experimental and theoretical studies.  
186 Adsorption Science and Technology 30, 307-316.

187 Bashkova, S., Baker, F.S., Wu, X.X., Armstrong, T.R., Schwartz, V., 2007. Activated  
188 carbon catalyst for selective oxidation of hydrogen sulphide: On the influence of  
189 pore structure, surface characteristics, and catalytically-active nitrogen. Carbon  
190 45, 1354-1363.

191 Brunauer, S., Emmett, P.H., Teller, E., 1938. Adsorption of gases in multimolecular  
192 layers. Journal of the American Chemical Society 60, 309-319.

193 Brunauer, S., Deming, L.S., Deming, W.E., Teller, E., 1940. On a theory of the van  
194 der Waals adsorption of gases. Journal of the American Chemical Society 62,  
195 1723-1732.

196 DIN 66165-1, 1987. Particle size analysis - Sieve analysis, Part1: Fundamentals (in  
197 German: Partikelgrößenanalyse – Siebanalyse, Teil 1: Grundlagen).  
198 Normenausschuß Siebböden und Kornmessung (NASK) im DIN Deutsche  
199 Institut für Normung. Beuth Verlag, Berlin.

200 Hawkslei, P.G.W., 1951. The physics of particle size measurement: Part I. Fluid  
201 dynamics and the Stokes diameter. British Coal Utilisation Research  
202 Association 15, 105-146.

203 Hu, Z.H., Srinivasan, M.P., Ni, Y.M., 2001. Novel activation process for preparing  
204 highly microporous and mesoporous activated carbons. Carbon 39, 877-886.

205 Kozeny, J., 1927. Über kapillare Leitung des Wassers im Boden (in German).  
206 Akademie der Wissenschaften Wien, 271–306.

207 Levenspiel, O., 1996. The Chemical Reactor Omnibook. OSU Book Stores Inc.,  
208 Corvallis.

209 Neimark, A.V., Lin, Y.Z., Ravikovitch, P.I., Thommes, M., 2009. Quenched solid  
210 density functional theory and pore size analysis of micro-mesoporous carbons.  
211 Carbon 47, 1617-1628.

212 Radian, A., Carmeli, M., Zadaka-Amir, D., Nir, S., Wakshal, E., Mishael, Y.G., 2011.  
213 Enhanced removal of humic acid from water by micelle-montmorillonite  
214 composites: Comparison to granulated activated carbon. Applied Clay Science  
215 54, 258–263.

- 216 Salmas, C.E., Androutsopoulos, G.P., 2001. A novel pore structure tortuosity concept  
217 based on nitrogen sorption hysteresis data. Industrial and Engineering  
218 Chemistry Research 40, 721-730.
- 219 Skibinski, B., Mueller, P., Uhl, W., 2016. Rejection of submicron sized particles from  
220 swimming pool water by a monolithic SiC microfiltration membrane: Relevance  
221 of steric and electrostatic interactions. Journal of Membrane Science 499, 92-  
222 104.