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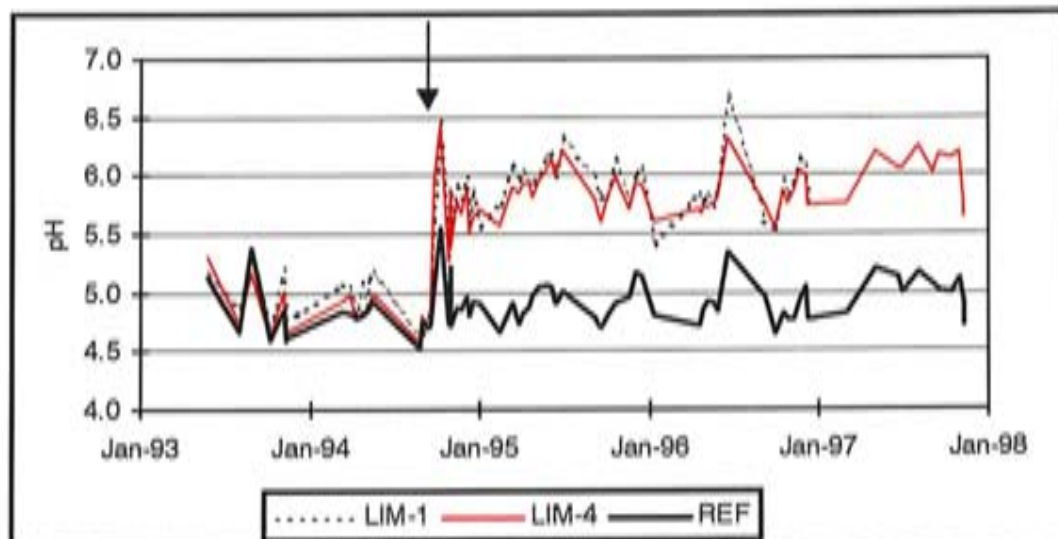


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**Whole-catchment**  
Application of Dolomite  
to an Acidified Forest  
Ecosystem in Gjerstad,  
Southern Norway

**Acid  
Rain  
Research**

REPORT 50/99



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**Abstract**

As part of the research programme "Countermeasures Against Acidification of Forest Ecosystems" (Miljøtiltak i skog), a coniferous forested catchment was limed with dolomite in September 1996. 3 t km<sup>-2</sup> of coarse dolomite powder were spread on 0.8 km<sup>2</sup> by helicopter. The liming resulted in an immediate improvement in runoff water quality relative to an adjacent reference catchment. pH, Ca<sup>2+</sup>, Mg and ANC<sup>+</sup> (acid neutralising capacity) increased and the pathic Al decreased. Favourable water quality was maintained for 3 years. NO<sub>3</sub><sup>-</sup> concentrations increased the second year after liming, whereas concentrations of total N and DOC were not significantly changed. Liming did not affect concentrations of 7 trace metals (As, Cd, Cu, Fe, Ni, Pb, Zn) whereas concentrations of Mn, Co, Cr, and Zn decreased. Only minor changes in soil solution, and only in some of the lysimeters were detected. Steep slopes, thin soils, high amounts of precipitation and thus dominance of surface and subsurface flow in this catchment may explain the rapid response in runoff. During the first three years after liming there have been no significant effects on tree growth and vitality (crown density and crown colour). This experiment shows that liming of forested catchments is a viable method to obtain long term improvement in water quality and positive effects for acid-sensitive aquatic organisms.

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Counteractions Against Acidification  
in Forest Ecosystems

**Whole-catchment Application of Dolomite to an  
Acidified Forested Ecosystem in Gjerstad,  
Southern Norway**

## Preface

The research program "Counteractions against acidification in forest ecosystems" (Miljøtiltak i skog) was initiated in 1991 as a Norwegian Forest Research Institute (NISK) program with funding from the Ministry of Agriculture. In 1993 the program was reorganised and a five-year research program led by NISK was started. Three co-operating institutes, NISK, the Norwegian Institute for Water Research (NIVA) and the Norwegian Institute for Nature Research (NINA), were main actors. Later on the Agricultural University participated.

The program was divided into an initial literature study and five separate research activities thereafter. This report gives background information and summarises the results from sub-project IV, the whole catchment application of dolomite to an acidified forested ecosystem in Gjerstad, southern Norway.

We would like to thank the forest owner Olav Ejerbu for careful sampling and maintenance of monitoring stations during the project period and for offering his forest as a research site in an early phase of the program.

This work has been financed by the Ministry of Agriculture, the Ministry of Environment, the four counties Rogaland, Vest-Agder, Aust-Agder and Telemark and NIVA.

Grimstad, September 16, 1999

*Alle Hinder*

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## Summary

As part of the research programme "Counteractions Against Acidification in Forest Ecosystems", a forested catchment dominated by pine (*Pinus sylvestris*) and Norway spruce (*Picea abies*) was limed with dolomite in September 1994. 3 t ha<sup>-1</sup> of coarse dolomite powder were spread on 80 ha by helicopter. Mean pre-liming stream water qualities (n = 16 to 18) at four independent stations were in the range: pH 4.6 - 4.9; 0.85 - 1.28 mg L<sup>-1</sup> Ca; 0.30 - 0.41 mg L<sup>-1</sup> Mg; 194 - 275 µg L<sup>-1</sup> reactive Al (Al<sub>r</sub>) and 34 - 105 µg L<sup>-1</sup> inorganic monomeric Al (Al<sub>i</sub>). Dolomite application resulted in an immediate increase in runoff water quality relative to an adjacent reference catchment. For the period September 1994-December 1996 mean values at the main outlet of the limed catchment were: pH 5.86; 1.51 mg L<sup>-1</sup> Ca; 0.81 mg L<sup>-1</sup> Mg; 15 µg L<sup>-1</sup> Al<sub>r</sub>. A small increase in NO<sub>3</sub><sup>-</sup> was found at the main outlet of the limed catchment the second winter after liming. Total N and TOC were not changed. Mn, Co and Zn showed significantly decreasing trends when compared to the reference stream, whereas Fe, Pb, Cd, Cu, As and Ni remained relatively unaffected by the dolomite application. Cr was mostly below detection limits. Only minor changes in soil solution, most clearly for Mg, could be detected after liming. Steep slopes, thin soils, high amounts of precipitation and thus dominance of surface and subsurface flow in this catchment may explain the rapid response in runoff. During the first three years after liming there have been no significant effects on tree growth and vitality (crown density and crown colour). This experiment shows that liming of forested catchments is a viable method to obtain long-term improvement in water quality and positive effects for acid-sensitive aquatic organisms.

# 1. Introduction

## 1.1 Background

In areas in southern Norway characterised by widespread fish-death due to acidification of surface waters, so far no negative effects of acid rain on forest health can be documented (Aamdal *et al.* 1998). Forests tend to remove base cations, scavenge air pollutants and concentrate pollutants due to increased evapotranspiration (Jenkins *et al.* 1990). In the long run, acidification of forest soils may therefore be expected in vulnerable areas if the soil has a limited base cation supply. The supply of nitrogen from atmospheric deposition may stimulate forest growth and thereby also the loss of base cations from the soil. Although atmospheric deposition of sulphur in Norway has decreased during the last decade, the situation for the forest ecosystems is still uncertain, and the need to examine various countermeasures against acidification of the soils has been recognised.

Episodic acidification, enhanced by sea salts, may be particularly important for sensitive aquatic organisms in forested areas (Hindin *et al.* 1994, 1995a). Afforestation may have undesired impacts on aquatic ecosystems, and the question of whether afforestation should be augmented with measures to prevent this extra acidification has been raised.

Several countermeasures against forest soil acidification are available for forest management practices, including both silvicultural measures and application of chemicals. The first group involves use of broadleaved trees, both in pure stands or mixed within conifer stands (Frank 1996).

Changes in forestry management may reduce the removal of base cations but will not prevent strong acid anions from entering soil solution and runoff. Strong acid anions, especially  $\text{SO}_4$ , are the driving force in soil acidification when balanced by base cations ( $\text{BC}^+$ , mainly Ca and Mg). Poorly buffered soils have a very limited capacity of base cation production because of the low weathering rate. In such soils strong acid anions may also be balanced by acid cations, especially  $\text{H}^+$  and  $\text{Al}(\text{OH})_n^{3-n}$ .

A second group of countermeasures includes application of carbonates (liming), ash and various commercial fertilisers. The addition of carbonates as calcic and dolomitic limestone increase the base saturation of the soils and thereby prevent leaching of  $\text{H}^+$  and toxic Al species. Liming may thus be regarded as desirable for both forest soils and aquatic systems in acidified areas, and has been practised for a long time as a compensatory or ameliorating measure in forest soils (Hind and Zöfel 1995; Krentzer 1995) and in acidified lakes and rivers (Coley 1991; Henriksen *et al.* 1995). Fertilisation with the goal of revitalisation of forest stands suffering from nutrient deficiencies has also proved successful (Hind 1991).

The question of whether liming should be launched on a practical scale in Norwegian forestry has been raised several times during the last decade (Nilsen *et al.* 1994). The inter-institutional research program "Counteractions against acidification in forest ecosystems" (Miljøtiltak i skog) was initiated in 1991. Main goals for the program were to:

- Give a status and evaluate potential and practical measures, such as liming, vitality fertilisation, use of broad-leaved trees and other silvicultural methods, that may counteract negative effects of acid deposition in forests.
- Perform research on the effects of selected measures on trees, ground vegetation, animal life, forest soils, soil solution and the runoff water quality.
- Give necessary basis for making decisions about future activities on a practical scale.



## 1.2 Definitions and targets

Liming is the addition of carbonates, oxides and hydroxides of calcium and magnesium to terrestrial and aquatic ecosystems to neutralise soil or surface water. This is consistent with the definition of Olem (1991), although he included all bases. Operational liming of forest soil has gradually changed towards dolomitic materials in Germany (Feger *et al.* 1995) due to the anticipated positive effect of Mg supply to prevent Mg deficiency. Dolomite and calcite mixture was also recommended for operational liming in Sweden (Lindström *et al.* 1993). However, some of the Swedish experiments were carried out using calcitic limestone, i.e. products composed of CaCO<sub>3</sub> with only a few percent Mg.

The targets for terrestrial liming operations may be many, e.g. the soil pH, the soil (Ca+Mg)/Al ratio, tree growth, forest health, stream water pH, stream water inorganic Al, a trout population and so on. Various application strategies may be used to achieve one or more of these goals (Henriksson *et al.* 1995), and lime quality, lime dosages, and application strategies may vary. If the goals are clarified, reasonable liming strategies may be recommended, if the potential effects are adequate and acceptable.

The main targets for forest soil liming should be to avoid the negative effects on forest soils of strong acid inputs and to avoid unwanted effects on the terrestrial and aquatic ecosystems. Another target may be improved living conditions for fish in the stream water draining the limed forest. If all the potential positive effects of forest liming are achieved all the negative effects may also occur, see section 1.3. If no undesired effects can be accepted, no positive effects may occur either. The real challenge is to maximise the positive and minimise the negative effects.

## 1.3 Potential effects of application of dolomite

Moderate addition of dolomitic limestone to the topsoil of forest ecosystems may be intermediate in both positive and negative effects. In this report the main focus is on reduced toxicity towards aquatic organisms.

### 1.3.1 Soil solution and stream water

The added dolomite gradually dissolves and releases Ca and Mg that can be adsorbed in the organic layer of the top soil. Vertical movement of these base cations to deeper soils then takes place, but this is a slow process, and may occur in the range of 1 cm year<sup>-1</sup> (Brahmer 1994). Fast changes in the chemical composition of soil solution are therefore not likely. The first change may be increased concentration of Mg in the soil solution due to the more mobile nature of this element relative to Ca (Brahmer 1994). On the other hand dolomite often contains almost four times as much Ca on a weight basis, thereby counteracting the dominance of Mg-increase over Ca-increase in the soil solution.

Although the change in soil solution may be slow, increased amounts of base cations in the topsoil may be important for runoff water quality when hydrological events are characterised by overland or subsurface flow. Overland and subsurface flow is promoted by thin soils and steep slopes, typical of acidified areas in Norway. High amounts of precipitation increase the likelihood for a high water table and saturated soils. Dominance of a more lateral movement of the runoff and reduced residence time of rainwater in the soil may be the result. Under such circumstances runoff acidity may be exchanged with the added base cations in the top soils, thus increasing the concentration of Ca and Mg and also increasing pH.

Dolomite addition thus may result in several positive effects, increased base saturation of the topsoil, eventually also at greater depths in the soil, and increased pH will reduce Al mobilisation. This in turn decreases the leaching of toxic Al to soil solution and surface water. Also, increased concentrations of base cations should be expected. If no significant changes in SO<sub>4</sub> and NO<sub>3</sub> occur, liming with dolomite will therefore result in increased ANC (acid neutralising capacity).

In acidified regions low pH and high Al concentrations in runoff may occur during episodes caused by sea salt inputs (Lindar *et al.* 1994; 1995a), snowmelt and other hydrological events. Increased base saturation in the forest soil and increased pH of the runoff after liming may reduce this Al transport, both by decreasing the preferential ion exchange with  $H^+$  and Al in acid soils and by polymerising and precipitating ion exchanged Al in the overland flow. If forest liming in such areas reduces the Al transport and promotes terrestrial polymerisation of Al, the magnitude and frequency of episodes with toxic water will be reduced and mixing zones in streams and rivers characterised by unstable Al chemistry (Roseland *et al.* 1992) may be avoided.

Undesirable chemical effects of terrestrial liming include increased  $NO_3^-$ -leaching and mobilisation of trace metals (Mätzner *et al.* 1985; Petsson *et al.* 1989; Hartl and Zintl 1993; Kreuzer 1995). This may partly be due to increased mineralisation of the surface organic (humus) matter, and may effect soil solution and surface water.

The liming-mediated change in the top soil pH may stimulate the bacterial degradation of organically bound N in the humic material to  $NH_4^+$  due to a more favourable environment for these organisms (Kreutzer 1994). Nitrification, the bacterial mediated conversion of  $NH_4^+$  to  $NO_3^-$ , may be stimulated and lead to decreased pH. If the supply of  $NO_3^-$  is already in excess due to atmospheric deposition, the result may be increased leaching of  $NO_3^-$  to the soil solution (Kreutzer 1995) and probably to runoff, with subsequent increased acidification and possibilities for increased supply of N to coastal areas. N leaching may thus reduce the demineralisation effect of the dolomite in runoff. In less N polluted areas,  $NO_3^-$  released will probably be stored in the soil and consumed by microorganisms and terrestrial vegetation.

Mineralisation may result in increased mobilisation of organically-complexed trace metals (especially Cu, Pb and Fe) and result in elevated concentrations in the soil solution. This may be especially so if the deposition rate of the metals is high, and they have accumulated in the surface soils, as in the Hapswald experimental area in Germany (Kreutzer 1995). Solubility of Cd and Zn (and Cu) is pH dependent and these metals will probably not be transported out after liming. On the contrary, increased pH may result in increased retention. Dissolved inorganic Mn and Cd may be oxidised and precipitated at increased pH and decrease in the soil solution after liming (Kreutzer 1995).

The deposition of Cu in Norway is mainly due to local sources and is generally low (Berg *et al.* 1996). Deposition of Zn and Pb in 1996 was an order of magnitude higher than other long-range transported metals but has decreased by 60-80 % in the period 1978-1996 (Torseth and Maur 1997). This reduction is also true for Cd. Compared to the situation in Germany deposition of trace metals in Norway is low.

### 1.3.2 Aquatic organisms

Forest liming, according to the Swedish concept (Lindström *et al.* 1993) tends to result in a relatively moderate water quality change in the runoff. Exceptions may be found in areas heavily polluted with nitrogen or trace metals, as referred to above. An important question is: will these changes in any way affect aquatic organisms?

Several investigations have documented that even small water quality changes or differences may be of significance for the survival of fish (Staurnes *et al.* 1993; Dalziel *et al.* 1995; Kroglund and Staurnes 1997) and invertebrates (Bakken and Fjellheim 1984). Decreased concentration of the toxic inorganic Al-species is probably most important for sensitive aquatic organisms and increased concentration of Ca may ameliorate Al toxicity (reviewed by Wood and McDonald 1987; Roseland *et al.* 1990). Increased concentration of dissolved organic matter will detoxify a larger fraction of Al at low pH due to complex binding. Forest soil liming may therefore represent an effective liming strategy for aquatic systems if significant changes in these parameters occur.

## 1.4 The whole-catchment liming project in Gjerstad

The question whether liming should be launched on a practical scale in Norwegian forestry may be answered if adequate results exist. Due to lack of integrated studies on soil, water and vegetation under Norwegian conditions, a catchment study with dolomite application was initiated in 1993. The project in Gjerstad, which is presented in this report, is one of five projects under the research program.

The purpose of the experiment was to study the effects of dolomite application to a forested area on soil solution chemistry, runoff water chemistry, soil chemistry, vegetation, tree growth and vitality. Few research programmes have focused on the effects of forest liming on both soil solution, stream water and the significance of water quality changes for aquatic organisms. One of the main purposes of the activities in the Gjerstad-project is to link these topics.

Results from this experiment have been published as part of the Acid Rain'95 Conference in Gothenburg (Hindar *et al.* 1995b) and in annual and final reports in Norwegian from the research programme. This report presents the results in more detail and covers the pre-liming and post-liming periods (until end of 1997).

## 2. Site description and liming

### 2.1 Site selection

Several possible areas were visited prior to the establishment of the project sites. The catchment was chosen to satisfy the following criteria:

- acidified due to long-range transported air-pollutants; pH in runoff in the range 4.5-5.5 and markedly elevated Al-concentrations
- mixed forest stands, typical for the moderately productive forests of the area
- mixed, but not too poor, growing conditions (nutrient and water availability) for ground vegetation
- not too close to the sea, due to possible speed effects related to sea-salts
- two defined catchments with brooks suited for sampling and monitoring of runoff, one for liming operation and one as reference.

A mixed forest in Gjerstad, Aust-Agder County, southern Norway (58° 53' N, 9° 00' E) was chosen for investigations of the effects of forest soil liming. Tugleliåsen (denoted LIM and REF; 84 ha) was selected as liming area and Spjøtåsen (denoted REF and SP; 41 ha) as reference. The limed catchment was further divided in 4 by the sampling stations LIM-1 (10 ha), LIM-2 and LIM-3, whereas LIM-4 is the main sampling station (Figure 1).

### 2.2 Forest cover

The catchments are forested with a mixture of mainly Norway spruce (*Picea abies*) and Scots pine (*Pinus sylvestris*) stands. A relatively small amount of broad-leaved trees are mixed within the conifer stands. The forest consists of old mature stands and with a certain amount of newly regenerated areas (less than 15 years).

**Table 1.** Catchment areas (ha), standing volume (m<sup>3</sup>), mean age (yr) and area distribution according to tree age classes and impediment.

	Size (ha)	Standing volume (m <sup>3</sup> )			Mean age (yr)	Middle aged & old forest	Regenerated areas	Imped.
		Total	Spruce	Pine				
LIM	84,4	9900	4400	4500	67	70 %	17 %	13 %
REF	40,8	6000	1600	3600	77	84 %	9 %	7 %

## 2.3 Precipitation and deposition

Mean annual precipitation is about 1200 mm and mean annual discharge about 900 mm ( $29 \text{ L s}^{-1} \text{ km}^{-1}$ ). Mean total wet + dry deposition of S and N in the project period 1994-1996 at the Birkenes monitoring station 70 km SW of Gjerstad was  $0.90 \text{ g m}^{-2}$  and  $1.52 \text{ g m}^{-2}$ , respectively. At the Solheimfjell monitoring station 20 km NW of Gjerstad the mean wet deposition of S and N was  $0.61 \text{ g m}^{-2}$  and  $0.91 \text{ g m}^{-2}$ , respectively, in the same period. Deposition of trace metals at Solheimfjell in 1996 was as follows (numbers in  $\text{mg m}^{-2} \text{ yr}^{-1}$ ): 4.95 Zn; 2.01 Pb; 0.82 Cu; 0.26 Ni; 0.18 As; 0.16 Cr; 0.042 Cd; 0.02 Co (all deposition data from Fossesh and Munis 1997).

## 2.4 Liming

In September 1994 a total of 240 tonnes of coarse grained dolomite was spread over the Engleliassen catchment, except for two small (about  $1000 \text{ m}^2$ ) ponds, by helicopter. This gives a mean dose of approximately  $2.9 \text{ t ha}^{-1}$ . The grain-size distribution of the dolomite was 10 %  $> 1.7 \text{ mm}$ ; 90 %  $> 0.18 \text{ mm}$ . The Ca, Mg and water content were 23, 12 and 1 % by weight, respectively.

# 3. Methods

## 3.1 Hydrology and water chemistry

### 3.1.1 Discharge

The limed catchment was divided into three sub-catchments (Figure 1), of which two (sampling stations LIM-1 of one of the sub-catchments and LIM-4 at the outlet of the limed catchment) have been included in the water chemistry monitoring in the whole project period. Monitoring at LIM-2 and LIM-3 of the two other sub-catchments were terminated in the summer of 1995.

Water flow was recorded beginning 26 May 1993 at 15-minute intervals at calibrated  $90^\circ$  (at LIM-1) and  $120^\circ$  (at LIM-4 and RIF) V-notch weirs.

### 3.1.2 Runoff water sampling

Volume weighted stream water samples were collected biweekly beginning 26 May 1993 at LIM-1, LIM-4 and RIF. Point samples were collected biweekly at the other two stations in the limed catchment. Water samples were sent to NIVA by post and were analysed 2-3 days after sampling.

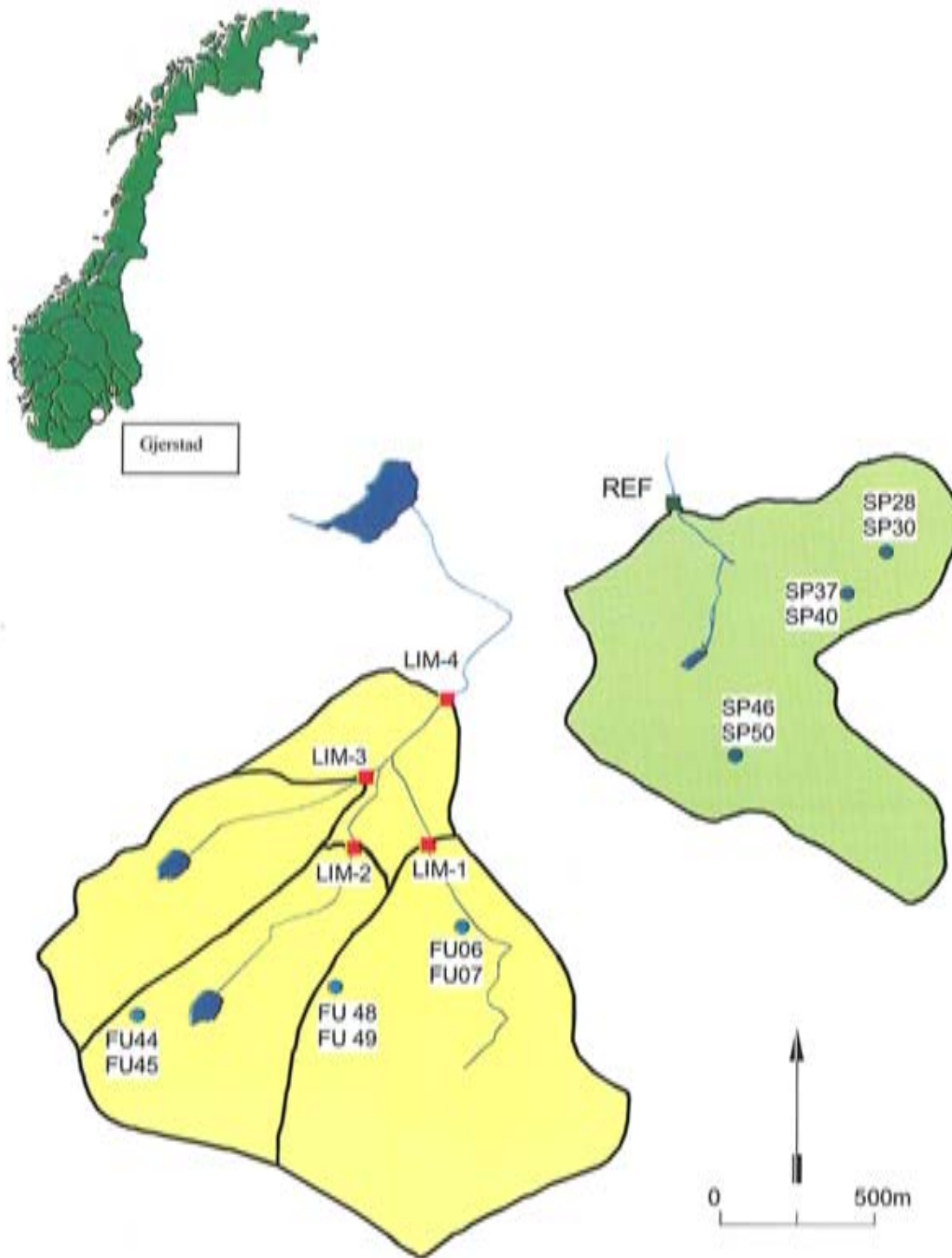
Samples for analysis of trace metals were taken directly in 100-ml polyethylene bottles. Before sampling the bottles were washed in 3%  $\text{HNO}_3$  for at least 24 hours and thereafter rinsed with demineralised water. After the washing procedure the PE-bottles were refilled with demineralised water until sampling. Before sampling the PE-bottles were emptied for demineralised water and immediately filled with the sample.

### 3.1.3 Chemical analyses - water and soil solution

Chemical analyses of all major ions and Al-fractions were carried out at NIVA according to standard procedures (Table 2). Trace metal analyses were performed with ICP-MS at the Norwegian Institute for Air Research (NILU). In the event that concentrations were reported below the detection limit, these were set at  $\frac{1}{2}$  the detection limit.

### 3.1.4 Calculations and statistical analyses

Al (inorganic monomeric Al) is defined as the difference between Al<sub>r</sub> (reactive Al) and Al<sub>o</sub> (organic monomeric Al). Organic N is defined as the difference: total N - ( $\text{NO}_3\text{-N} + \text{NH}_4\text{-N}$ ).



**Figure 1.** Situation of the limed and reference catchments in Gjerstad. Stations for surface water sampling (LIM-1, LIM-2, LIM-3, LIM-4 and REF), soil chemistry and soil solution sampling (FU and SP stations) are shown. Runoff is registered continuously at LIM-1, LIM-4 and REF.

**Table 2.** Chemical analysis methods used at NIVA (major components) and NHEI (trace metals). ICP= Inductive coupled plasma emission spectrometry, ICP-MS= Inductive coupled plasma emission spectrometry combined with mass spectrometry; FIA=Flow injection analysis.

Parameter	Parameter name	Unit	Analytical method
pH	pH	$-\log [H^+]$	Potentiometric
Cond	Conductivity	$mS\ m^{-1}$ at 25 °C	Electrometric
Ca	Calcium	$mg\ L^{-1}$	ICP
Mg	Magnesium	$mg\ L^{-1}$	ICP
Na	Sodium	$mg\ L^{-1}$	ICP
K	Potassium	$mg\ L^{-1}$	ICP
Cl	Chloride	$mg\ L^{-1}$	Ion chromatography
SO <sub>4</sub>	Sulphate	$mg\ L^{-1}$	Ion chromatography
NO <sub>3</sub>	Nitrate	$\mu g\ N\ L^{-1}$	Automatic colorimetry
NH <sub>4</sub>	Ammonium	$\mu g\ N\ L^{-1}$	Automatic colorimetry
Alk	Alkalinity	$\mu eq\ L^{-1}$	Potentiometric titration to pH 4.5
TOC	Total organic carbon	$mg\ L^{-1}$	Oxidation to CO <sub>2</sub> and then IR detector
Alr	Reactive aluminium	$\mu g\ L^{-1}$	Automatic colorimetry
Alo	Non-labile aluminium	$\mu g\ L^{-1}$	Automatic colorimetry after ion exchange
SiO <sub>2</sub>	Silica	$mg\ L^{-1}$	Photometry (PIA)
Tot N	Total nitrogen	$\mu g\ L^{-1}$	Automatic colorimetry
Tot P	Total phosphorus	$\mu g\ L^{-1}$	Automatic colorimetry
As	Arsenic	$\mu g\ L^{-1}$	ICP-MS
Cd	Cadmium	$\mu g\ L^{-1}$	ICP-MS
Co	Cobalt	$\mu g\ L^{-1}$	ICP-MS
Cu	Copper	$\mu g\ L^{-1}$	ICP-MS
Fe	Iron	$\mu g\ L^{-1}$	ICP-MS
Mn	Manganese	$\mu g\ L^{-1}$	ICP-MS
Ni	Nickel	$\mu g\ L^{-1}$	ICP-MS
Pb	Lead	$\mu g\ L^{-1}$	ICP-MS
Zn	Zinc	$\mu g\ L^{-1}$	ICP-MS

ANC<sup>+</sup> (Acid neutralising capacity), in  $\mu eq\ L^{-1}$ , is defined as:

$$ANC^+ = (Ca^{2+} + Mg^{2+} + Na^+ + K^+) - (Cl^- + SO_4^{2-} + NO_3^-) \text{ if } NH_4^+ = 0. \text{ All concentrations are in } \mu eq\ L^{-1}.$$

Non-marine Na (Na\*) in  $\mu eq\ L^{-1}$  is calculated as:

$Na^* = Na^+ - 0.859 * [Cl^-]$ , assuming that Cl<sup>-</sup> is entirely derived from sea salts. If the Na/Cl ratio departs from the ratio in the sea (0.859), this indicates sources or sinks of Na in the catchment.

Net change in transport of Ca and Mg from the lined catchment is calculated from water flow and the difference between simultaneously taken samples from LIM 4 and REP. This difference was then corrected for the mean pre-lining difference between the two catchments. This correction was +0.14  $mg\ L^{-1}$  for Ca and -0.01  $mg\ L^{-1}$  for Mg.

The data material was tested statistically by different methods:

#### Runoff water samples;

1. Systematic pre liming variability between sub catchments was tested with simple ANOVA. The tests were performed on the means of calculated differences of simultaneously taken samples at independent (no autocorrelation) LIM stations and the REF station. This difference, which reduces the contribution of variance from seasonal variability, is referred to as e.g. Ca diff.
2. Changes due to liming were also tested by use of this dataset and ANOVA.
3. Trends after, and due to, liming were tested by use of simple regression on the differences. General trends, as reflected in the reference catchment, were thereby eliminated.
4. Changes in water chemistry in the limed catchment (LIM4) relative to the reference catchment (REF) at the point of time of liming (September 1994) were tested by Random Intervention Analysis (RIA) (Carpenter *et al.* 1989). RIA compares differences in concentrations in paired, chronologically-ordered samples for the pre- and post-liming periods with the change obtained from 1000 randomly-ordered sets of differences.

## 3.2 Soil and soil solution chemistry

### 3.2.1 Soil sampling

Soil were sampled in summer 1994 before liming at the 6 plots (3 each in the limed and reference catchments) used for analysis of ground vegetation (Figure 1). At each plot 1 sample was collected from the humus horizon and 2-3 samples from the mineral soil horizon.

### 3.2.2 Chemical analyses – soils

Soils were extracted in water for measurement of pH and in 1 M  $\text{NH}_4\text{NO}_3$  for determination of exchangeable cations. Bulk density, loss on ignition, and Kjeldahl N were measured on bulk soil. Contents of 31 elements was determined by ICP (Inductive coupled plasma emission spectrometry) on acid digests of the bulk soil. Analyses were performed at NISK using standard procedures.

Cation exchange capacity (CEC;  $\text{NH}_4\text{NO}_3$  extract) was calculated as the equivalent sum of  $\text{H}^+$ , Al, Ca, Mg, Na, and K. Base saturation (%) is defined as the sum of  $(\text{Ca} + \text{Mg} + \text{Na} + \text{K})/\text{CEC}$ .

### 3.2.3 Soil solution sampling

60 Prenat lysimeters were installed prior to liming (August 1994) in the 6 locations selected for vegetation analyses (Figure 1), at 3 sites in Engleklåsen (EE) and 3 in Spjotåsen (SP), two lysimeters for each sampling depth. The samples were collected following a 24-hour evacuation period. The soil solution was collected from 5 and 15 cm depth and stored in 2 litre Prenat bottles, which were covered to minimize reaction catalysed by sun light or alga growth. The soil solution samples were filtered through 2 µm membranes and stored for less than 48 hours prior to analysis at NIVA. Sampling began in September 1994. There are no pre-liming samples of soil solution.

## 3.3 Forest inventory

Forest inventories were carried out during summer 1994 in both limed and reference areas. The inventory was made as a systematic sample plot inventory. Lines were laid out with individual distances of 100 m in the forest and along each line sample plots were placed with 50-m distances (Figure 2). The plots were circular with an area of 200  $\text{m}^2$ . Plots on mineral soil were permanently marked with a wooden pole in the centre. Plots that fell on peat-land, roads, water or barren rocks were classified as non-productive land. Plots that fell on border areas between stands were systematically moved 10 meters in one direction. A total of 122 (limed area) and 56 (reference area) plots were measured.

### 3.3.1 Tree measurement

In each sample plot all trees more than 5 cm in diameter at breast height were measured by diameter and permanently marked with tree number. Tree species were noted.

Sample trees were picked out with relascope. Relascope factor 2 was used and each third of the trees falling within the relascope were picked out for further measurements. These sample trees were measured for height above ground, height to living crown and different kinds of damages were noted according to a manual. Sample cores were taken from all sample trees and annual ring width was measured for the latest 40 years. From each sample plot, where possible, the two largest trees (by diameter) without damages were picked out for site index estimation. In addition to the variables already mentioned, bark thickness was measured and annual height growth visually estimated on these trees. All annual rings were measured for breast height age determination on these trees. The annual rings were measured in laboratory.

### 3.3.2 Site index, volume- and cutting class calculations

Site index for each sampling plot was estimated from age at breast height and height of the two largest trees (Tveite 1977). Curves for diameter to height relationship for each tree species were constructed for the two areas together. Tree height was then estimated for all trees. Volume was calculated using functions for spruce, pine and birch (all broad leaved trees are treated as birch in this connection) (Braastad 1966, Vesjordnet 1967, Brandsegg 1967) with diameter and height as independent variables. Cutting class (for definition see text in Figure 24) is dependent on site index, species and age and the area distribution of cutting classes have been calculated from the sample plot data for the two areas.

## 3.4 Intensive monitoring plots

In each of the two areas, 10 sample plots (25 x 30 m) were subjectively laid out (Figure 3). The criteria for placing the different plots were to cover the vegetation gradient in the limited catchment, and a comparable gradient in the reference catchment. The gradient varied from dry to wet and from fertile to infertile soils. Within each sample plot an area of 5 x 10 m was used for intensive regeneration of vegetation (Eilertsen *et al.* 1994). On each plot steepness and exposition were measured.

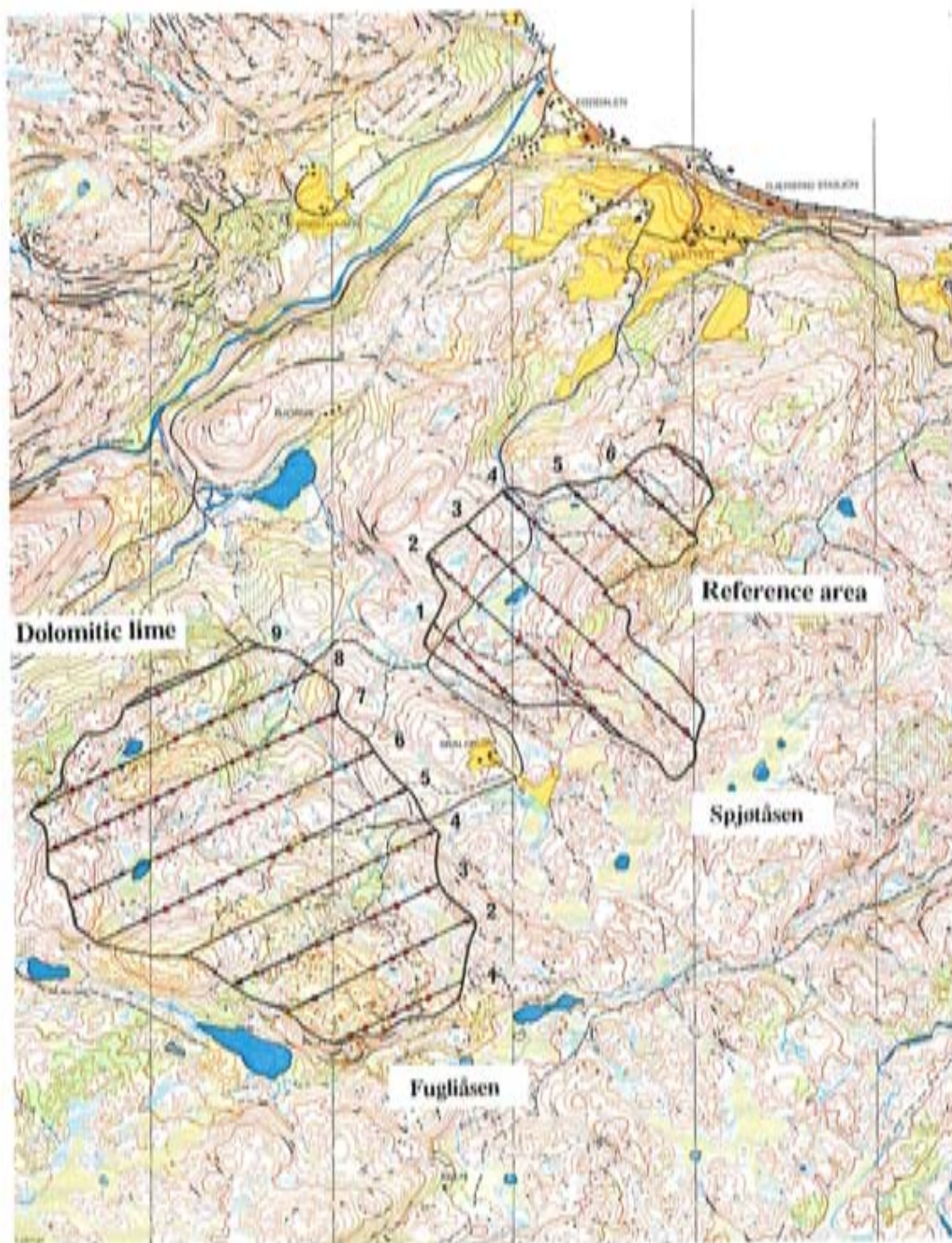
### 3.4.1 Registration in the plots

All trees more than 5 cm in breast height were measured by diameter and permanently marked with a number. Height and crown height was measured on about half of the trees in 1994. Height increment was visually estimated on the trees and bark thickness was measured. In 1998 height was just measured on 1/3 of the trees. Damages were also noted for these trees.

The 3 largest trees by diameter (spruce or pine) were cored to pith, and age and annual ring width were measured in the laboratory.

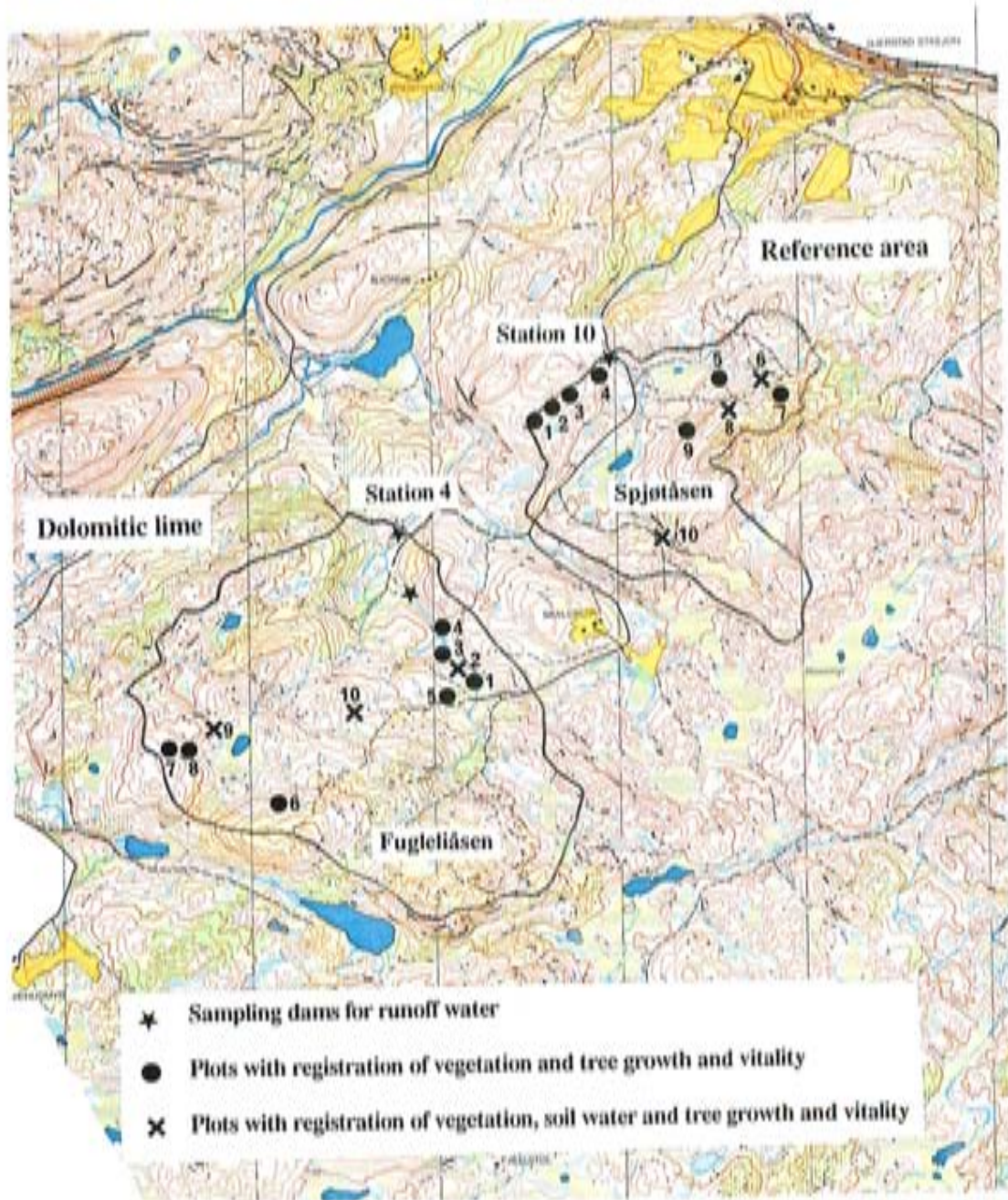
Each autumn, diameter was measured on all trees and tree vitality evaluated. The vitality evaluation is based on tree crown density and colour. The crown density (%) is defined as the amount of needles in tree crowns compared to what could be expected on a sound tree in that particular area. The crown colour is an overall impression of the colour based on a scale from 1 to 4: 1 is normal green, 2 slightly yellow, 3 average yellow and 4 strongly yellow. The result from the evaluation is based on the same trees each year and the work has been done by the same person and follows the same criteria as for the Norwegian forest monitoring programme (Groggen 1997).





**Figure 2.** Map showing the two catchments and the sampling lines in each area. Red dots indicate plots where trees were measured. Scale approx. 1:20000.





**Figure 3.** The location of the intensive monitoring plots. All plots were used for vegetation investigations and some for soil solution sampling. Scale approx. 1:20000.

### 3.4.2 Statistical analysis

Based upon height and diameter the volume of the sample trees was estimated using functions for different tree species (Hraastad 1966, Vestjordet 1967, Brandsegg 1967). A linear regression between diameter and volume on these trees was used to estimate volume for all trees. The function developed from the 1994 registration was also used for the 1998 data. Volume increment in the 4 year period was calculated as the difference in standing volume between the two years.

Site index (Tveite 1977) was estimated on the basis of age and height of the 3 trees per plot. Site index curves for the dominant tree species was used to determine the site index of the plots.

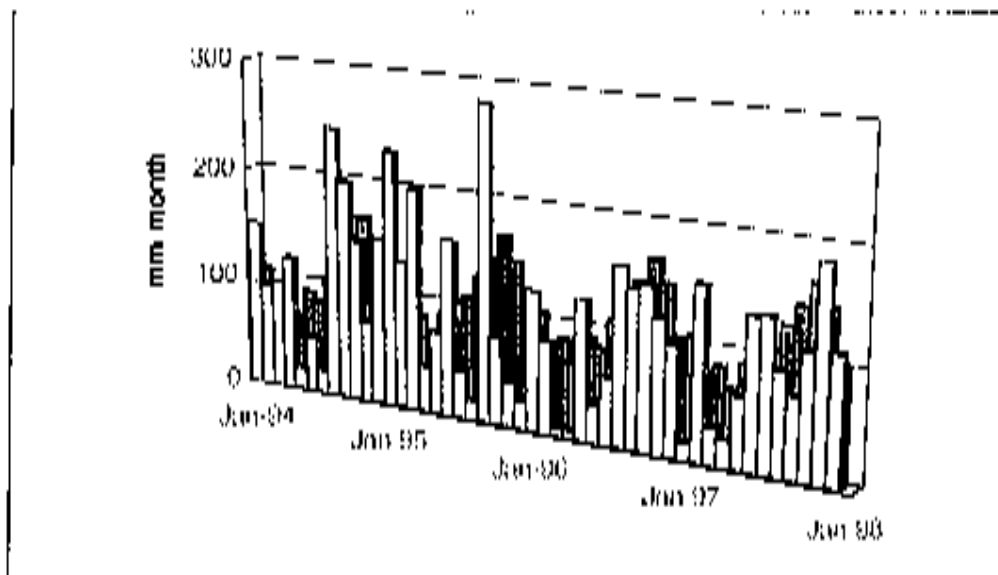
The increment in the 4-year period from 1994 to 1998 was compared to increment functions developed for Norwegian conditions based on increment investigations on permanent sample plots (Blingsmo 1988). The increment function has standing volume, stand age and site index as explanatory variables. Functions for pine, spruce and birch (used for all broad-leaves) were used and the increment was weighted with the tree species share of total volume.

## 4. Results

### 4.1 Precipitation and hydrology

Monthly precipitation amount at the nearby meteorological station Åsbu in Gjerstad and water discharge at the three monitoring stations are shown in Figure 4 and Figure 5.

According to the runoff map of Norway average discharge in the area is  $900 \text{ mm yr}^{-1}$  ( $29 \text{ L s}^{-1} \text{ km}^{-2}$ ). Based on the measured runoff and the subcatchment areas of stations LIM 1, LIM 4 and RIF specific discharge for the four years 1994-1997 were calculated (Table 3).



**Figure 4.** Monthly precipitation amount at the meteorological station (3520 Åsbu) in the monitoring period (in the front) and the 30-year monthly normal precipitation (in the back). Data from the Norwegian Meteorological Institute.

1994 and the hydrological year from summer 1994 to summer 1995 was very wet compared to the following years. Only the year 1995 was close to normal. 1996, 1997 and especially the hydrological year 1995-1996 were dry. Lower discharge in 1996 and 1997 compared with the initial period after liming is also evident when accumulated discharge is calculated (Figure 6).

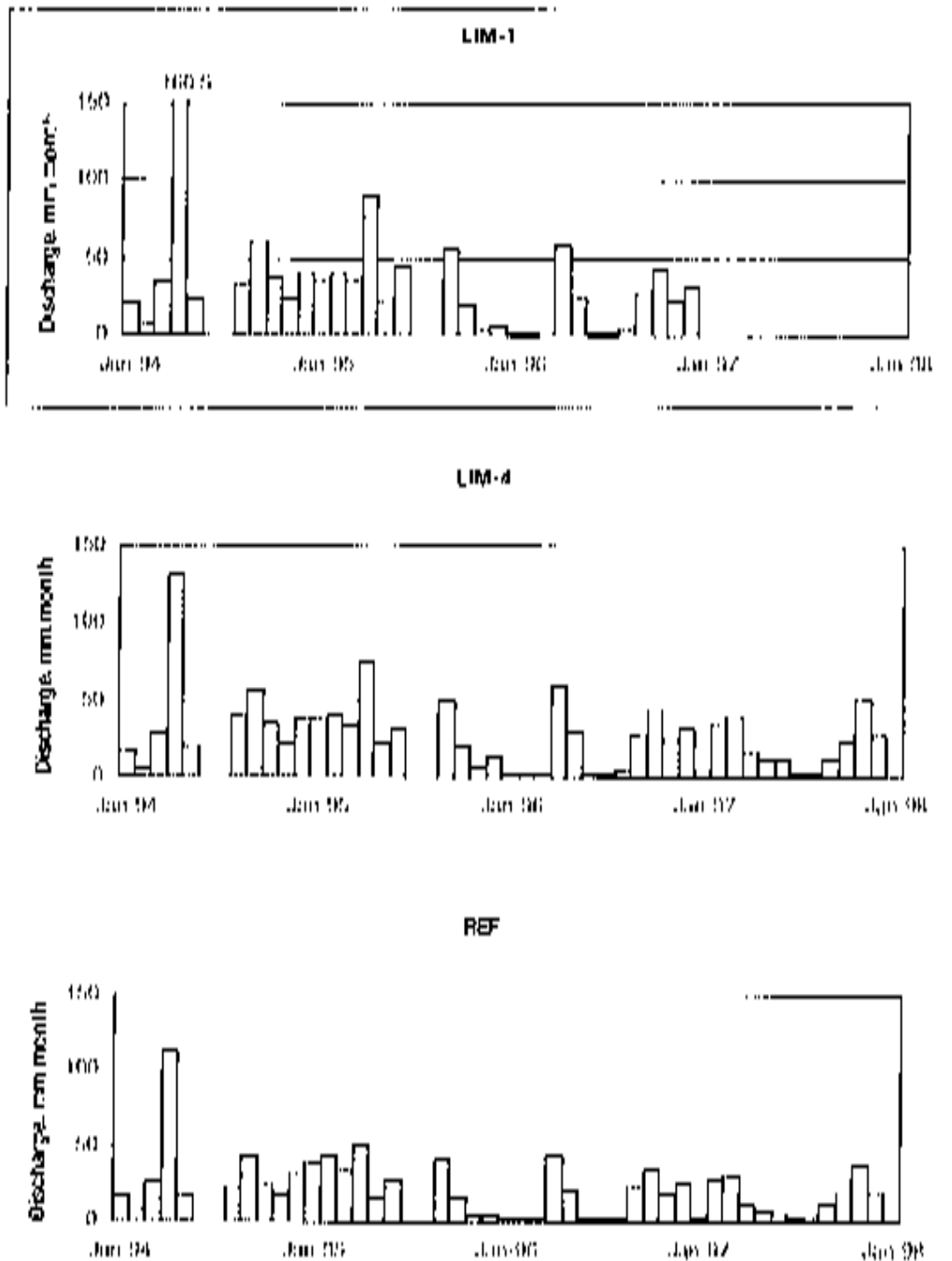
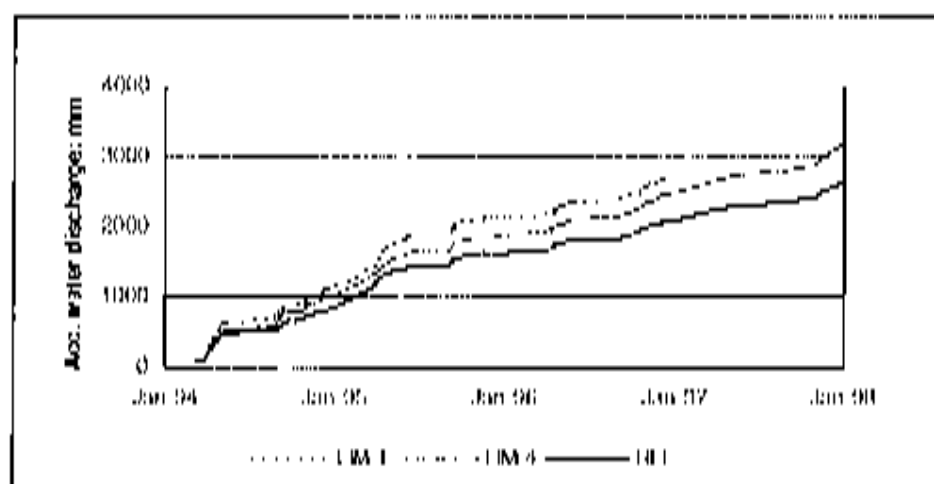


Figure 5. Monthly discharge at LIM-1, LIM-4 and REF. Discharge in 1997 was not measured at LIM-1.



**Figure 6.** Accumulated water discharge (in mm) at monitoring stations LIM-1, LIM-4 and REF for the period 1994-1997.

**Table 3.** Calculated specific discharge ( $\text{mm yr}^{-1}$ ) for catchments LIM-1, LIM-4 and REF. Results are given for calendar years and for hydrologic years (1 July to 30 June).

Calendar year	1994	1995	1996	1997
LIM-1	1057	839	517	612
LIM-4	1034	855	602	609
REF	1082	918	599	
Hydrologic year	1994-1995	1995-1996	1996-1997	
LIM-1	1104	410	653	
LIM-4	1126	486	640	
REF	1170	464		

## 4.2 Runoff water quality

### 4.2.1 Pre-liming characteristics

Significant pre-liming variability between subcatchments is reflected in pH, Ca, Mg, ANC, Al-fractions and TOC but not in  $\text{SO}_4$  and  $\text{NO}_3$  (Table 4). This reflects the relatively uniform response of S and N deposition on  $\text{SO}_4$  and  $\text{NO}_3$  in runoff within the catchment and the more non uniform geochemical characteristics. The areas draining to LIM-2 and LIM-3 are more acid sensitive than the LIM-1 catchment. These two areas also had the most toxic water before liming, as seen in lower pH and ANC and higher Al.

Prior to liming the stream water in both catchments was chronically acidified (Table 4 and Table 5). pH was generally between 4.5 and 5.5 (Figure 7), and mean concentrations of reactive aluminium (Alr) were  $248 \mu\text{g L}^{-1}$  and  $191 \mu\text{g L}^{-1}$  in the two catchments (Table 5). About 70% of the Al was organically bound. The relatively high organic Al fraction can be related to the generally high contents of total organic matter (TOC) in the streams ( $6-8 \text{ mg TOC L}^{-1}$ ).

### 4.2.2 Post-liming characteristics

After liming significant increases in pH, Ca, Mg and ANC were found for all LIM stations relative to the REF-station. Table 5 summarises this effect for LIM-4. The immediate response in pH, which increased to 5.5-6.0, is clearly seen in Figure 7. At the main outlet of the limed catchment (at LIM-4) the mean concentration of Ca and Mg increased by 0.35 and  $0.44 \text{ mg L}^{-1}$ , respectively. Acid

Table 4. Predicting mean values and standard deviations of 4 independent sampling stations (n=1) for LHM-2, n=19 for LHM-3, LHM-3 and REF and LHM-4 (n=19) stations and detectable significant differences ( $p < 0.05$ ) between columns based on multiple range tests.

	LHM-1	LHM-2	LHM-3	REF	LHM-4
SO <sub>4</sub>	4.5 (0.6)	4.2 (0.5)	4.4 (0.7)	4.3 (0.7)	4.4 (0.6)
NO <sub>3</sub> -N	74 (8)	67 (7)	50 (5)	50 (5)	70 (7)
H <sup>+</sup>	13 (6)	24 (3)	17 (6)	17 (6)	15 (6)
Fe <sup>2+</sup>	4.94 (0.23)c	4.62 (0.10)a	4.82 (0.20)bc	4.79 (0.20)bc	4.52 (0.19)bc
Ca <sup>2+</sup>	1.22 (0.17)b	0.85 (0.19)a	1.09 (0.23)bc	1.26 (0.33)bc	1.13 (0.29)bc
Mg <sup>2+</sup>	0.21 (0.11)c	0.39 (0.06)c	0.32 (0.05)ab	0.36 (0.05)ab	0.36 (0.05)ab
ANC	17 (4)cd	6 (9)bc	2 (10)a	27 (7)cd	10 (11)bc
Al <sup>3+</sup>	233 (72)ab	277 (65)b	265 (74)b	192 (56)a	248 (63)bc
Al <sup>2+</sup>	175 (63)	154 (51)	164 (51)	157 (49)	176 (52)
Al <sup>1+</sup>	55 (20)b	89 (7)c	93 (35)c	34 (12)a	32 (13)bc
TDC	3 (15)ab	6.5 (2.4)ab	5.1 (2.1)a	5.5 (2.9)bc	7 (2)2.6)ab

neutralising capacity (ANC) increased from 11 to 49  $\mu\text{eq L}^{-1}$ , whereas a decrease was seen in the unfined catchment (Table 5). Liming did not result in significant changes in Na, K,  $\text{NH}_4$ ,  $\text{SO}_4$ , Cl, Al<sub>o</sub>, total N, organic N,  $\text{CO}_3$  or organic C/N.

After the initial increase during the first 10 months after liming, no significant trends were detected for base cations, pH, ANC or aluminium species at LIM-4. There were thus no signs of re-acidification.

A significant decrease in Al<sub>t</sub> ( $p < 0.001$ ) due to liming was found for LIM-1 and LIM-4 (Table 6 and Figure 9). The decrease in Al<sub>t</sub>-concentration was rather modest, however, from 47 and 47  $\mu\text{g L}^{-1}$  pre-liming to 30 and 39  $\mu\text{g L}^{-1}$  post-liming for LIM-1 and LIM-4, respectively, relative to REF. This means that the Al<sub>t</sub>-concentration in the limed catchment was only slightly lower than in the reference catchment after liming. A corresponding decrease from 24 and 38 to 22 and 21  $\mu\text{g L}^{-1}$  was found for the difference in the inorganic monomeric fraction (Al<sub>o</sub>-diff). Al<sub>t</sub> is supposed to include the toxic Al-species, and the mean concentration of this fraction was 17  $\mu\text{g L}^{-1}$  after liming (Table 5). No significant liming related change in Al<sub>o</sub> or  $\text{CO}_3$  was found.

The Random Intervention Analysis also shows that pH, Ca, Mg, and ANC concentrations increased and Al<sub>t</sub> and Al<sub>o</sub> decreased significantly in LIM-4 relative to REF in September 1994, the point of time of liming (Table 6). RIA indicates no significant changes in  $\text{CO}_3$  or  $\text{NO}_3$ .

A significant ( $p < 0.05$ ) increasing trend was found for  $\text{CO}_3$  at LIM-1, but not at LIM-4. For sulphate and the organic C/N-ratio this was reversed, a significant increasing trend at LIM-4, but not at LIM-1.

**Table 5.** Mean stream water values and std. deviation (in brackets) for different variables in the limed (LIM-4) and reference (REF) catchments before and after treatment in September 1994: n = 17-19 pre-liming and n = 55-57 post-liming for all variables; \* denotes significant change ( $p < 0.05$ ) of mean or median at LIM-4 due to liming and at REF after the time of liming. Significance of change at LIM-4 is based on differences of simultaneously taken samples according to ANOVA or the non-parametric Kruskal-Wallis test, see also Table 6

		LIM-4		REF	
		Before	After	Before	After
H <sup>+</sup>	µeq L <sup>-1</sup>	15.9 (6.2)	1.6 (0.88)*	17.3 (6.3)	12.7 (4.5)*
pH	units	4.80 (0.19)	5.86 (0.23)*	4.79 (0.20)	4.93 (0.18)*
Ca	mg L <sup>-1</sup>	1.13 (0.29)	1.51 (0.25)*	1.26 (0.33)	1.29 (0.28)
Mg	mg L <sup>-1</sup>	0.36 (0.08)	0.81 (0.17)*	0.36 (0.08)	0.38 (0.08)
Na	mg L <sup>-1</sup>	1.81 (0.35)	1.73 (0.20)	1.81 (0.39)	1.74 (0.25)
K	mg L <sup>-1</sup>	0.21 (0.11)	0.21 (0.09)	0.17 (0.08)	0.20 (0.12)
NH <sub>4</sub> -N	µg L <sup>-1</sup>	23 (31)	40 (56)	19 (14)	52 (58)
SO <sub>4</sub>	mg L <sup>-1</sup>	4.4 (0.6)	4.3 (0.9)	4.3 (0.7)	4.0 (0.9)
Cl	mg L <sup>-1</sup>	2.2 (0.8)	2.6 (0.7)	2.2 (0.8)	2.5 (0.7)*
NO <sub>3</sub> -N	µg L <sup>-1</sup>	70 (71)	98 (98)**	50 (57)	59 (58)
ANOC	µeq L <sup>-1</sup>	11 (11)	54 (19)*	23 (17)	17 (14)
Alr	µg L <sup>-1</sup>	248 (69)	167 (40)*	191 (56)	176 (38)
Alb	µg L <sup>-1</sup>	175 (59)	152 (40)	157 (49)	130 (30)
Alf	µg L <sup>-1</sup>	72 (19)	15 (12)*	54 (14)	37 (18)
Total N	µg L <sup>-1</sup>	314 (163)	355 (115)	328 (115)	358 (124)
Org N	µg L <sup>-1</sup>	223 (116)	217 (80)	261 (106)	247 (100)
TOC	mg L <sup>-1</sup>	7.0 (2.6)	7.4 (2.1)	8.5 (2.9)	8.4 (2.3)
Org C/N		34 (6)	35 (4)	34 (4)	35 (6)

**Table 6.** Results of random intervention analysis comparing difference in concentrations in paired samples collected from LIM-4 (limed) and REF (reference) for the periods pre- and post-liming.

			H <sup>+</sup>	pH	Ca	Mg	Alr	Alf	TOC	ANOC	NO <sub>3</sub> -N
			µeq L <sup>-1</sup>		mg L <sup>-1</sup>	mg L <sup>-1</sup>	µg L <sup>-1</sup>	µg L <sup>-1</sup>	mg L <sup>-1</sup>	µeq L <sup>-1</sup>	µg L <sup>-1</sup>
pre-liming n = 18	average	LIM4	15.9	4.80	1.13	0.36	250	73	7.2	11	70
	average	REF	17.3	4.80	1.28	0.37	191	54	8.7	24	50
	avg. diff-pre	LIM4-REF	-1.4	0.00	-0.14	-0.00	56	38	-1.5	-13	20
post-liming n = 57	average	LIM4	1.6	5.86	1.51	0.81	167	15	7.4	54	98
	average	REF	12.7	4.93	1.29	0.38	176	36	8.4	17	59
	avg. diff-post	LIM4-REF	-11.1	0.93	-0.22	-0.42	-9	21	-1.0	36	40
change in difference between LIM4 and REF at point of time of liming, Sept. 94											
diff post - diff pre			-9.7	0.89	-0.36	-0.43	65	60	0.5	-49	19
RIA level of significance, p-			0.001	0.001	0.001	0.001	0.001	0.001	n.s.	0.001	n.s.



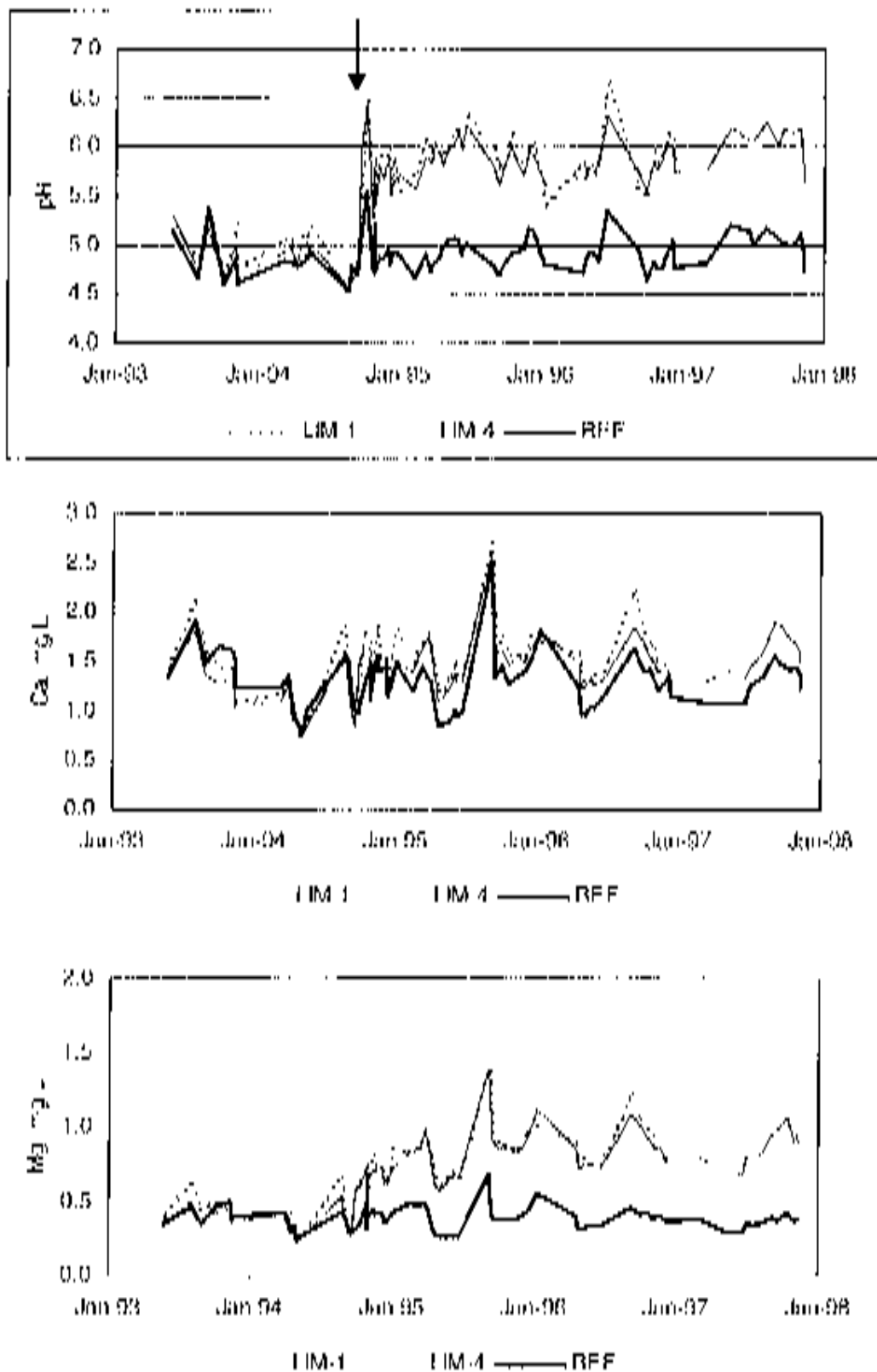
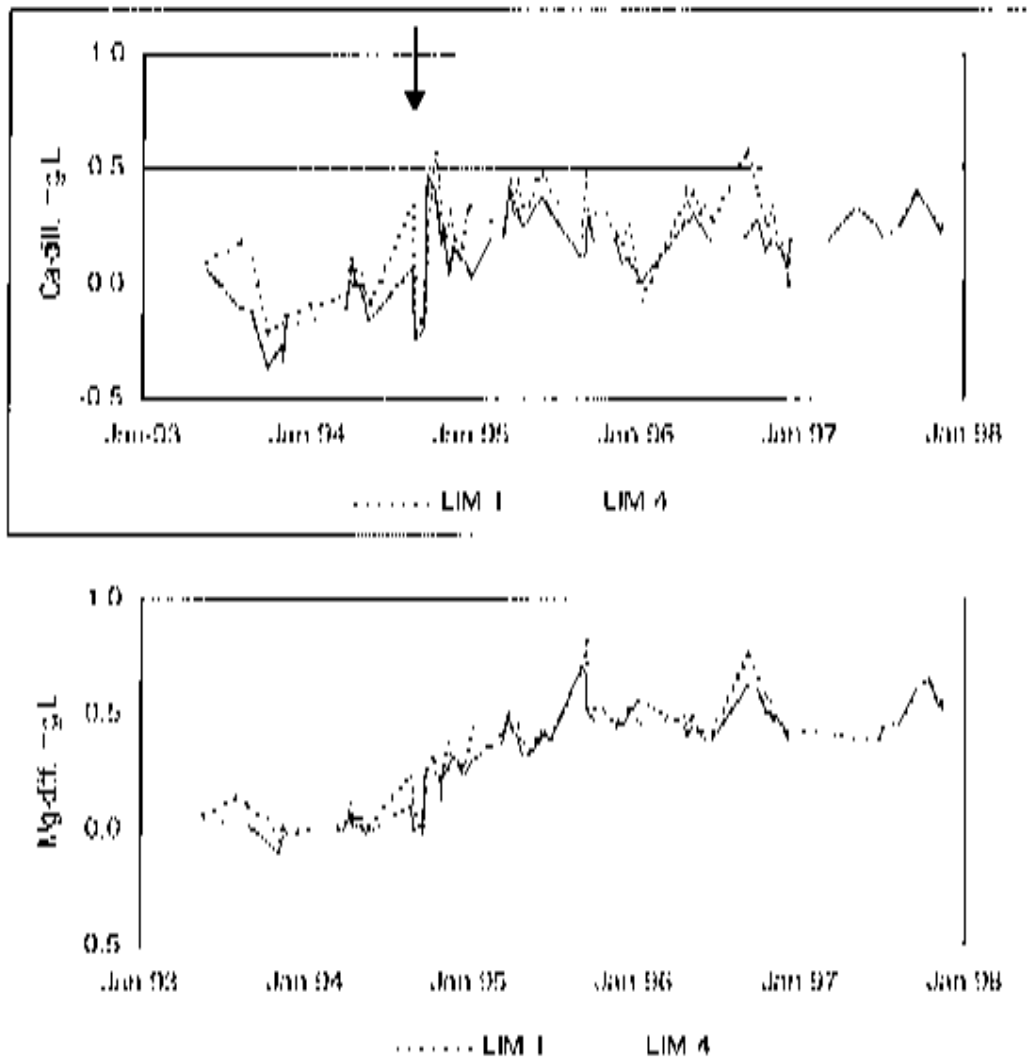
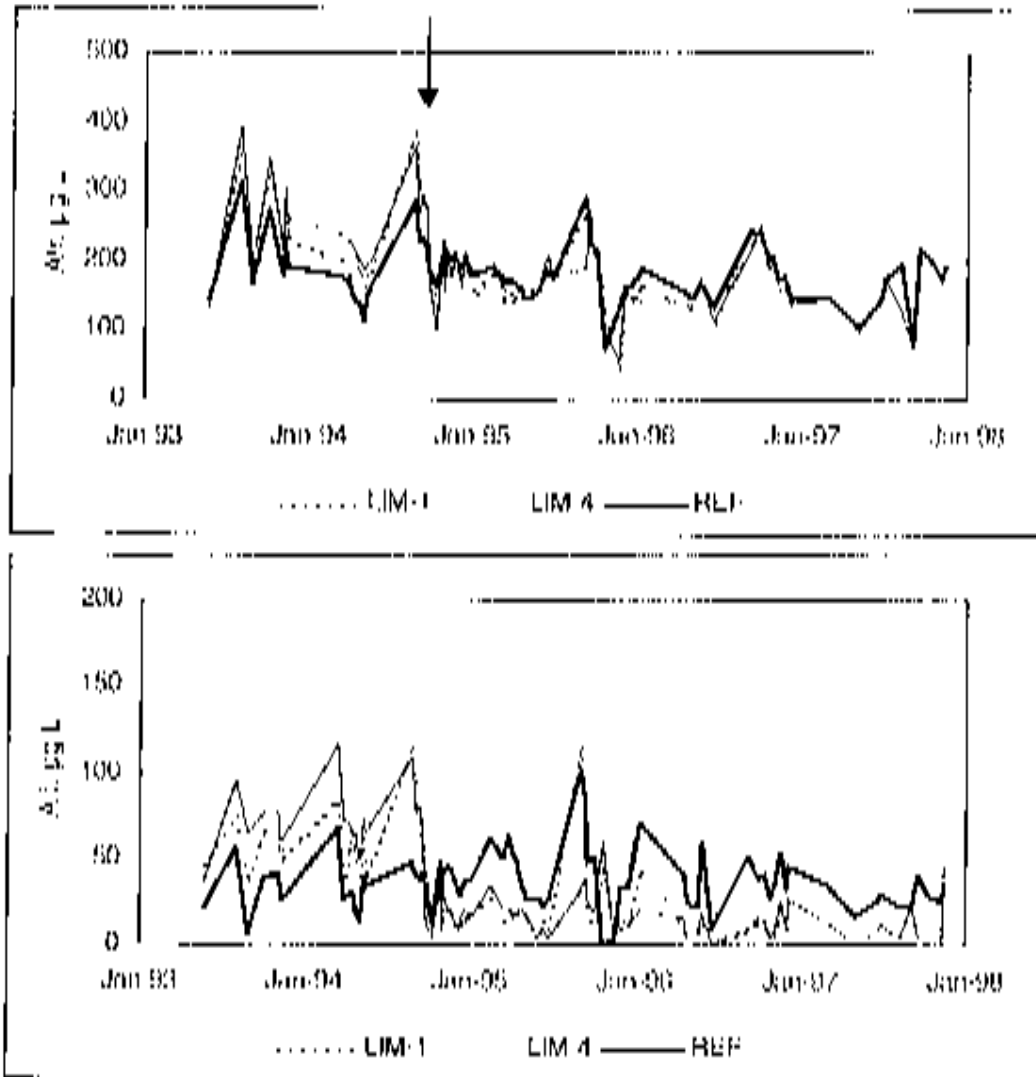


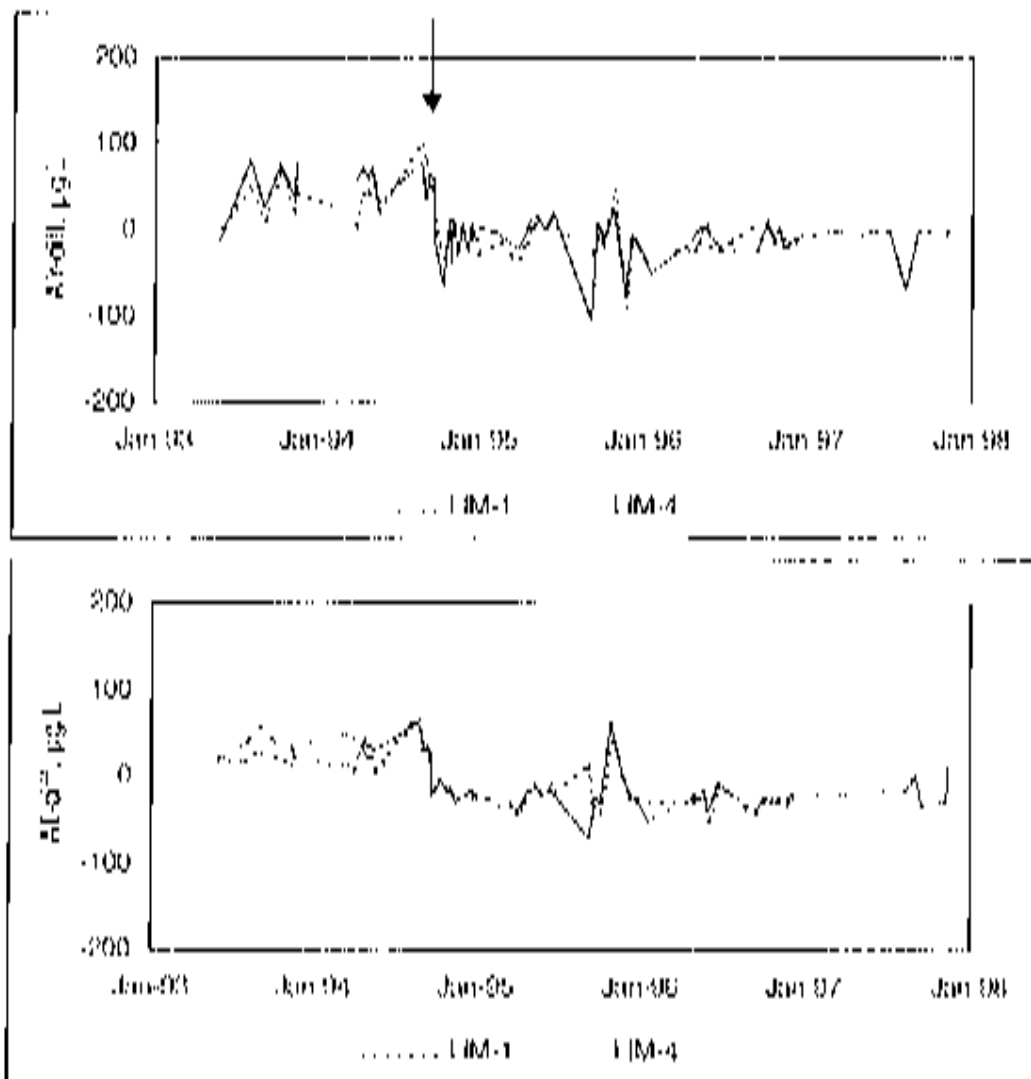
Figure 7. pH, Ca and Mg in streams of limed (LIM 1 and LIM 4) and reference (REF) catchment in the Gjerstad forest. Dolomite was spread in September 1994 (arrow).



**Figure 8.** Differences in Ca and Mg concentrations between limed and reference streams in the Gjerstad forest. Dolomite was spread in September 1994 (arrow)



**Figure 9.** Reactive (Al<sub>r</sub>) and monomeric (Al<sub>m</sub>) aluminium in streams of limed (LIM 1 and LIM 4) and reference (REF) catchment in the Gjerstad forest, Dolomite was spread in September 1994 (arrow).

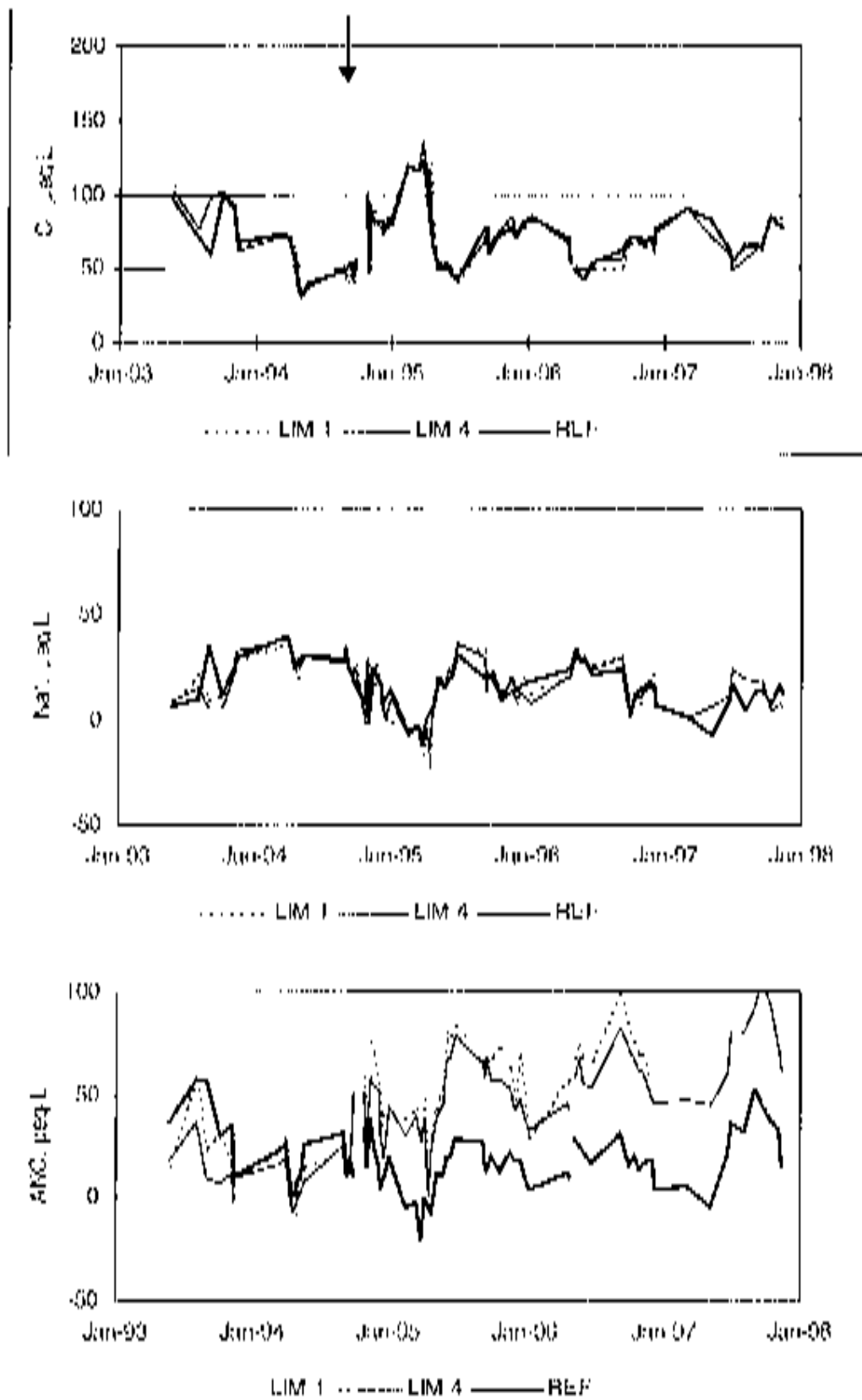


**Figure 10.** Differences in concentrations of reactive ( $Al_r$ ) and labile ( $Al_L$ ) aluminum between limed and reference streams in the Gjerstad forest. Dolomite was spread in September 1994 (arrow).

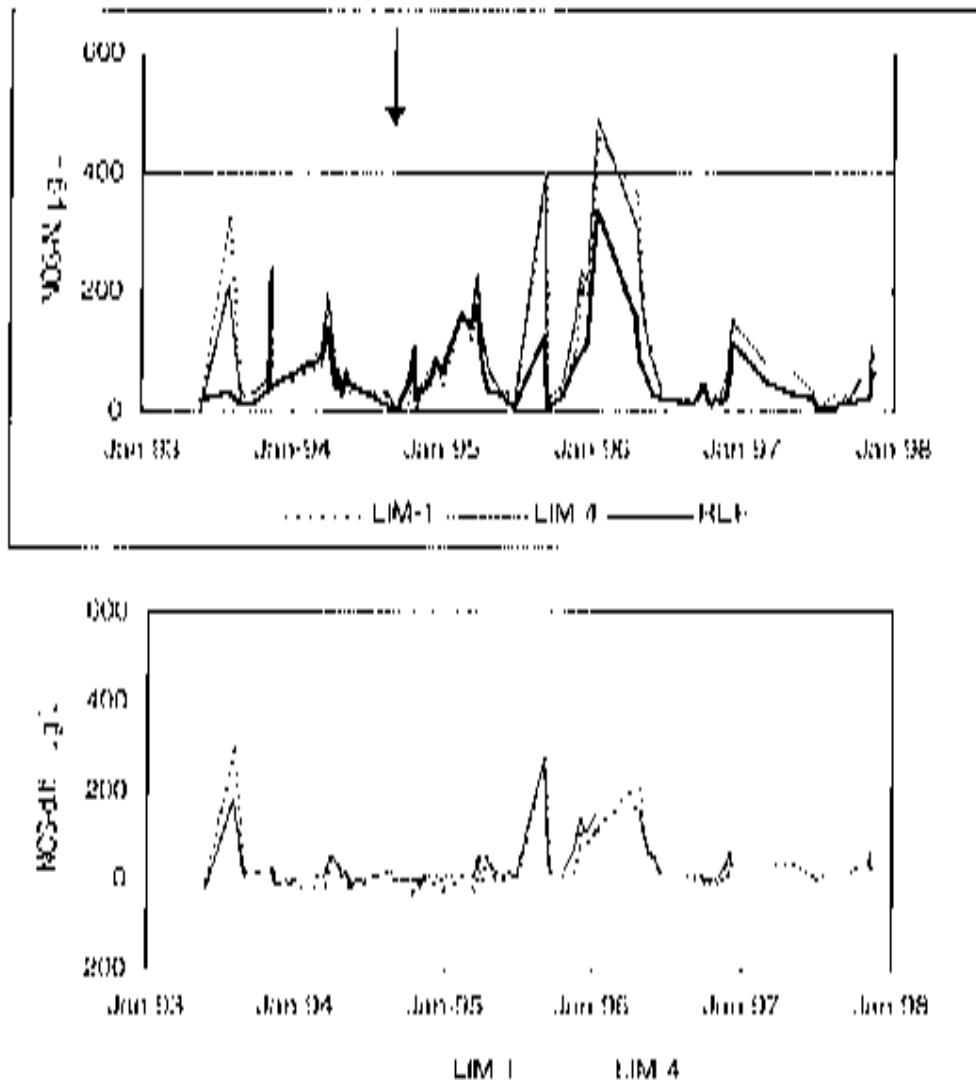
Sea salts were not supposed to have significant impacts on the water chemistry, in terms of episodic acidification and  $Mg$  supply to the forest soil, due to the distance from the coast. However, in January 1995 (four months after liming) a strong increase in stream  $Cl$  concentrations indicated relatively large inputs of salts, and this resulted in negative values for non marine  $Na$  and the lowest estimated  $ANC$  in the unlimed catchment in the project period **Figure 11**. The limed catchment also showed a drop in  $ANC$ , but not to negative values.

The seasonal variability was relatively large for concentrations of  $NO_3^-$  (**Figure 12**). Some of the problems this might have on the statistical tests were eliminated by use of the differences between data-pairs of limed and reference samples. But also the differences were characterised by seasonal variability (**Figure 12**). This was found for both the preliming and post liming period. The streams do not appear to differ until the second winter following liming. The result of the ANOVA test of the pre-liming and post-liming  $NO_3^-$ -diff for LIM-1 and LIM-4 was a small but significant increase in the medians at the  $p < 0.05$  level in LIM-1, but not in LIM-4. Random intervention analysis showed no significant difference (**Table 6**).

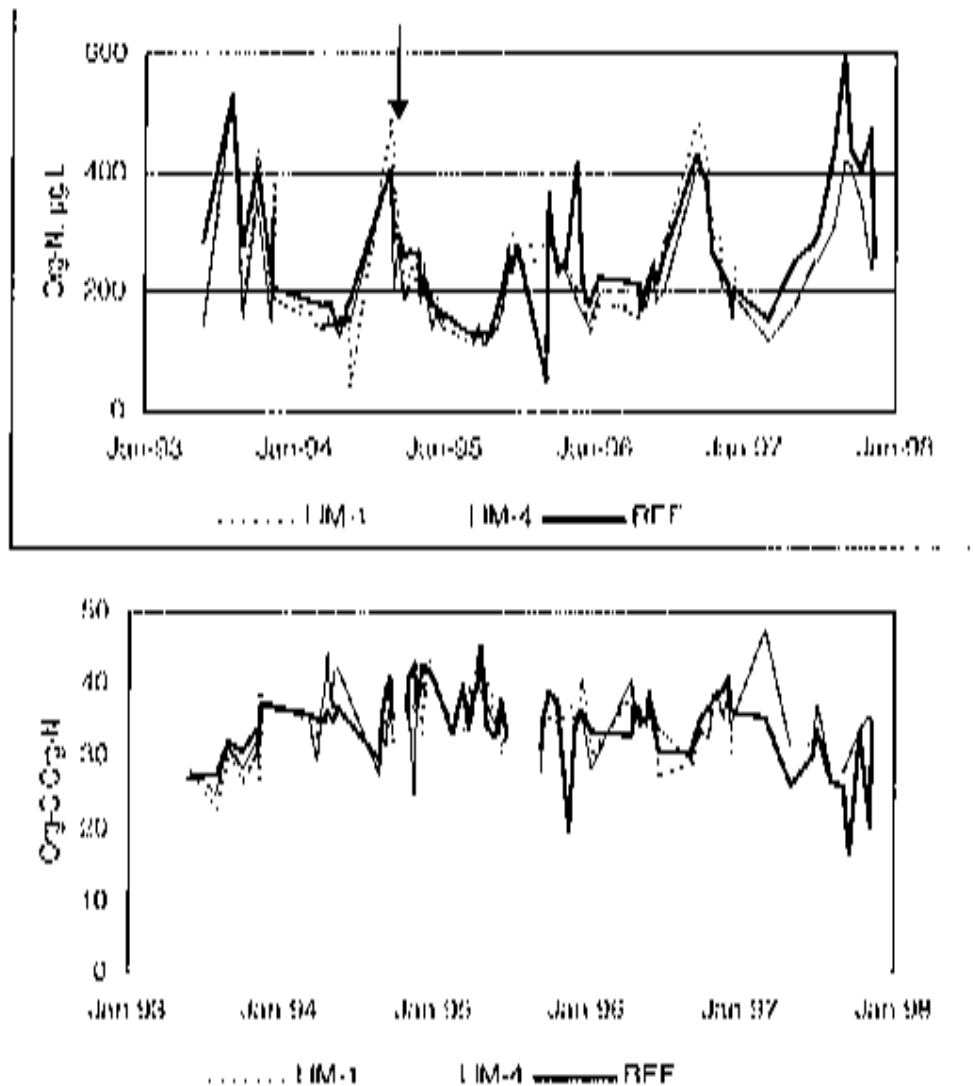
The organic N fraction, the C/N-relationship of the dissolved organic matter and total phosphorus concentration did not change significantly after liming (**Table 5** and **Figure 13**).



**Figure 11.** Chloride, non-marine Na (Na<sup>+</sup>) and ANC<sup>-</sup> in streams of limed (LIM-1 and LIM-4) and reference (REF) catchment in the Cijerstad Forest. Dolomite was spread in September 1994 (arrow).



**Figure 12.** NO<sub>3</sub>-N in streams of limed (LIM-1 and LIM-4) and reference (REF) catchments in the Gjerstad forest. The NO<sub>3</sub>-N-diff. is shown in the lower panel. Dolomite was spread in September 1994 (arrow).

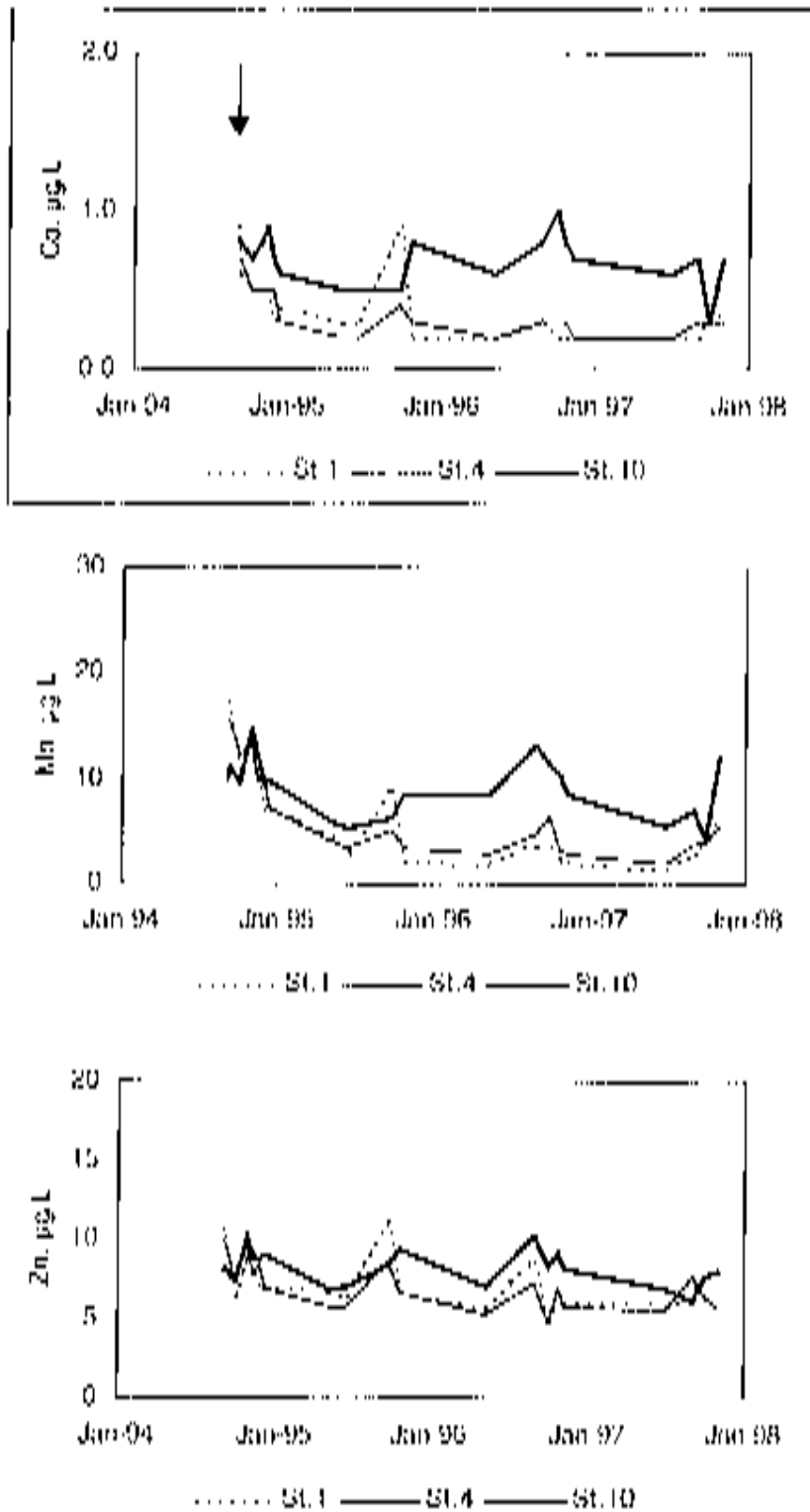


**Figure 13.** Dissolved organic N and the Org C/Org N ratio in streams of limed (LIM-1 and LIM-4) and reference (REF) catchment in the Cještad forest. Dolomite was spread in September 1994 (arrow).

#### 4.2.3 Trace metals

No pre-liming data of trace metals are available and thus changes from pre liming to post liming period cannot be analysed. However, concentrations in the limed stream relative to the reference stream were analysed for differences and post-liming trends.

Relative to REF none of 10 trace metals in monthly point samples were significantly higher after liming (Table 7). In fact, Cd, Co, Fe, Mn, Ni, Pb and Zn were significantly lower in the stream of the limed catchment, which may be due to inherent differences between the streams or may be an effect of the liming. Only Mn, Co and Zn showed significantly decreasing trends in runoff from the limed catchment during the 3 year monitoring period (Figure 14). Fe, Pb, Cd, Cu, As and Ni remained relatively unaffected by the dolomite application. Cr was mostly below detection limits in both limed and unlimed catchments.



**Figure 14.** Cobalt (Co), manganese (Mn) and zinc (Zn) in streams of limed (LIM 1 and LIM 4) and reference (REF) catchment in the Ciperstul forest. Dolomite was spread in September 1994.

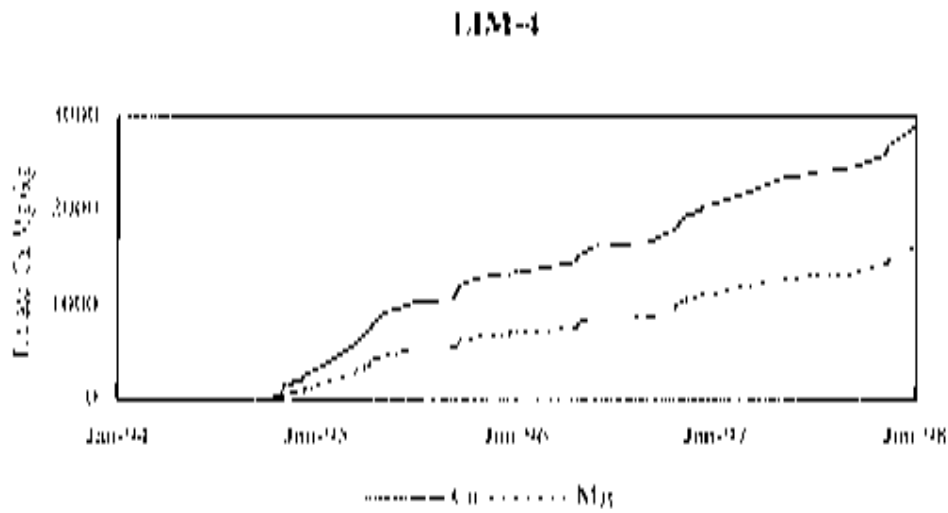
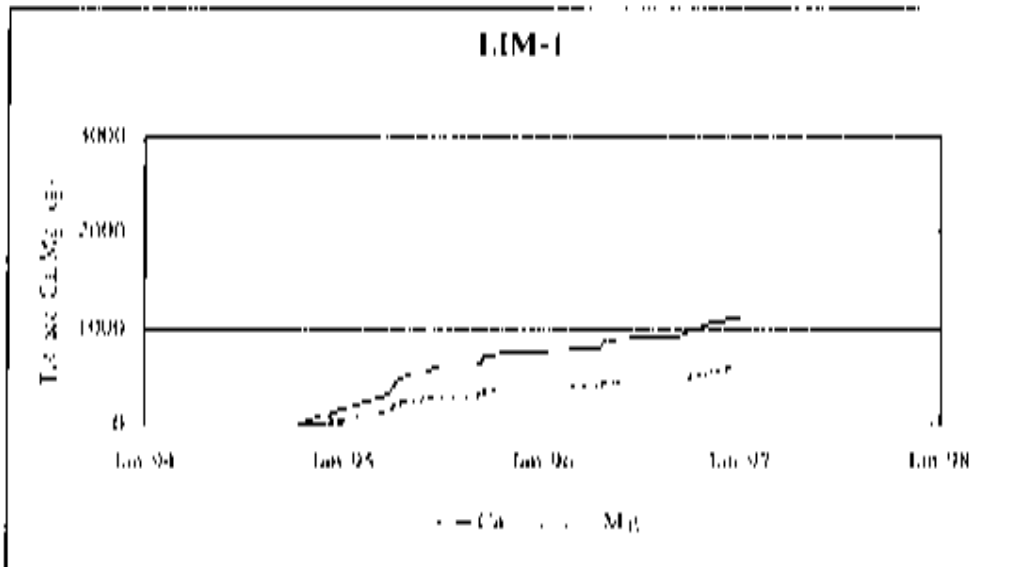


**Table 7.** Mean and st. dev. or median (in  $\mu\text{g L}^{-1}$ ) of nine trace metals from point samples of streams in limed (LIM-4) and unlimed (REF) catchments. All samples (n = 18) are from the post liming period. \* denotes significantly ( $p < 0.05$ ) lower mean or median values for LIM-4 than for REF. Other metals are not significantly different. Concentrations in the samples reported under the detection limit were set at  $\frac{1}{2}$  the detection limit.

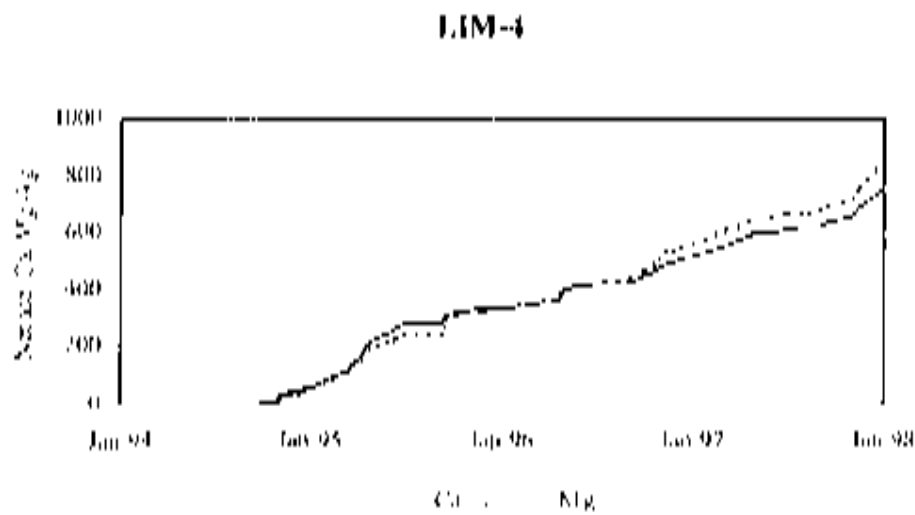
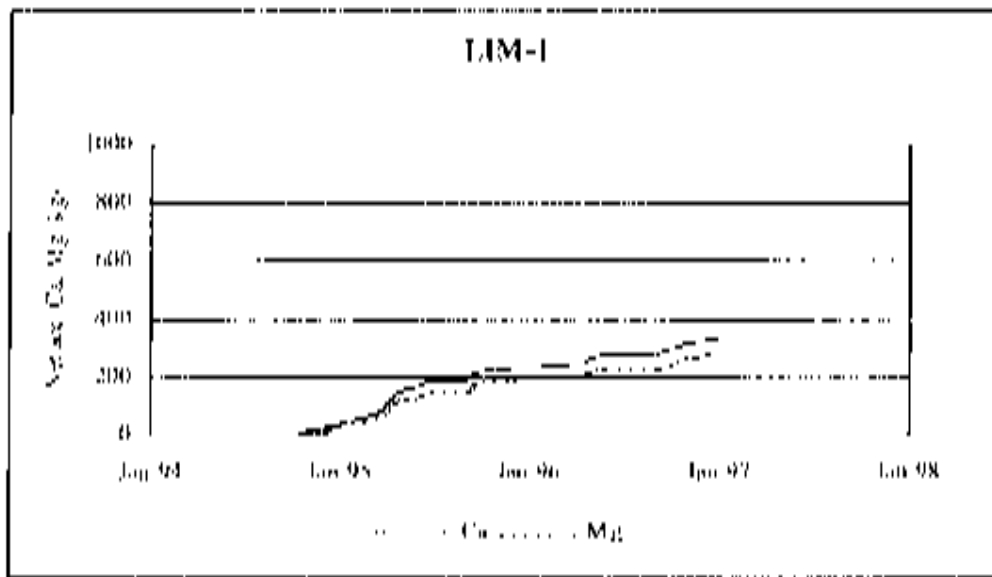
	LIM-4		REF	
	Mean (st.dev.)	Median	Mean (st.dev.)	Median
As	0.27 (0.15)		0.36 (0.20)	
Cd*		0.03		0.05
Co*	0.32 (0.11)		0.68 (0.17)	
Cu	0.52 (0.17)		0.58 (0.21)	
Fe*		122		201
Mn*	5.7 (3.4)		9.0 (3.0)	
Ni*		0.8		1.05
Pb	0.64 (0.19)		0.72 (0.16)	
Zn*	6.8 (1.3)		8.3 (1.7)	

#### 4.2.4 Ca and Mg transport

The total transport of Ca and Mg at LIM-4 during the first 3 years after liming was 2090 and 1150 kg, respectively (**Figure 15**; transport at LIM-1 is also shown). Based on the pre-liming and post-liming differences between samples from LIM-4 and REF a net transport of 518 kg Ca and 576 kg Mg from the applied dolomite was calculated for the first 3 years after liming (**Figure 16**). This corresponds to 0.9 % and 2.0 % of the added amounts of these elements. An annual transport of less than 1 % of the dose for both Ca and Mg was thus found in this experiment.



**Figure 15.** Accumulated total amount of Ca and Mg transported out at stations LIM-1 and LIM 4 of the lined catchment.



**Figure 16.** Accumulated net (due to liming) amount of Ca and Mg transported out at stations LIM-1 and LIM-4 of the limed catchment, relative to that expected based on transport from the reference catchment (RDC).

### 4.3 Soil chemistry

The soil data at the lysimeter stations represent the composition of the organic (5 cm) and mineral (15 cm) soil layer before liming. Data for six locations, three located in the limed catchment and three located in the control catchment are given in **Table 8** and **Table 9**.

The results show that the soils vary widely in chemical composition. This reflects the wide range in site characteristics. The 3 sites in each catchment were intentionally chosen to cover a range in vegetation, soil and moisture conditions.

#### 4.3.1 Soils of the control catchment

Stations SP 38 and SP 39

This location is covered with ferns and deciduous trees. The ground water table is high, and there were never problems in filling the lysimeter bottles. The pH values in the water samples extracted from the organic and mineral soil were relatively high (4.82 and 4.9, respectively).

Stations SP 37 and SP 40

Station SP 37 is located in a hollow in the ridge hill, while SP 40 is located in a rather dry and poor slope just above SP 37. The dominating forest is coniferous. The soil solution samples normally had a vigorous yellow colour and the volume at SP 40 was usually low, especially at 15 cm.

Stations SP 46 and SP 50

These stations are located in a shady area beneath large spruce trees. The ground vegetation was poorly developed. The site is moderately dry.

#### 4.3.2 Soils of the limed catchment

Stations FU 06 and FU 07

Stations FU 06 and FU 07 were located in poor soil in a pine forest. The soil solution samples from this location normally were low in volume and of yellow colour. The soil quality can be compared to soils at stations SP 30 and SP 40.

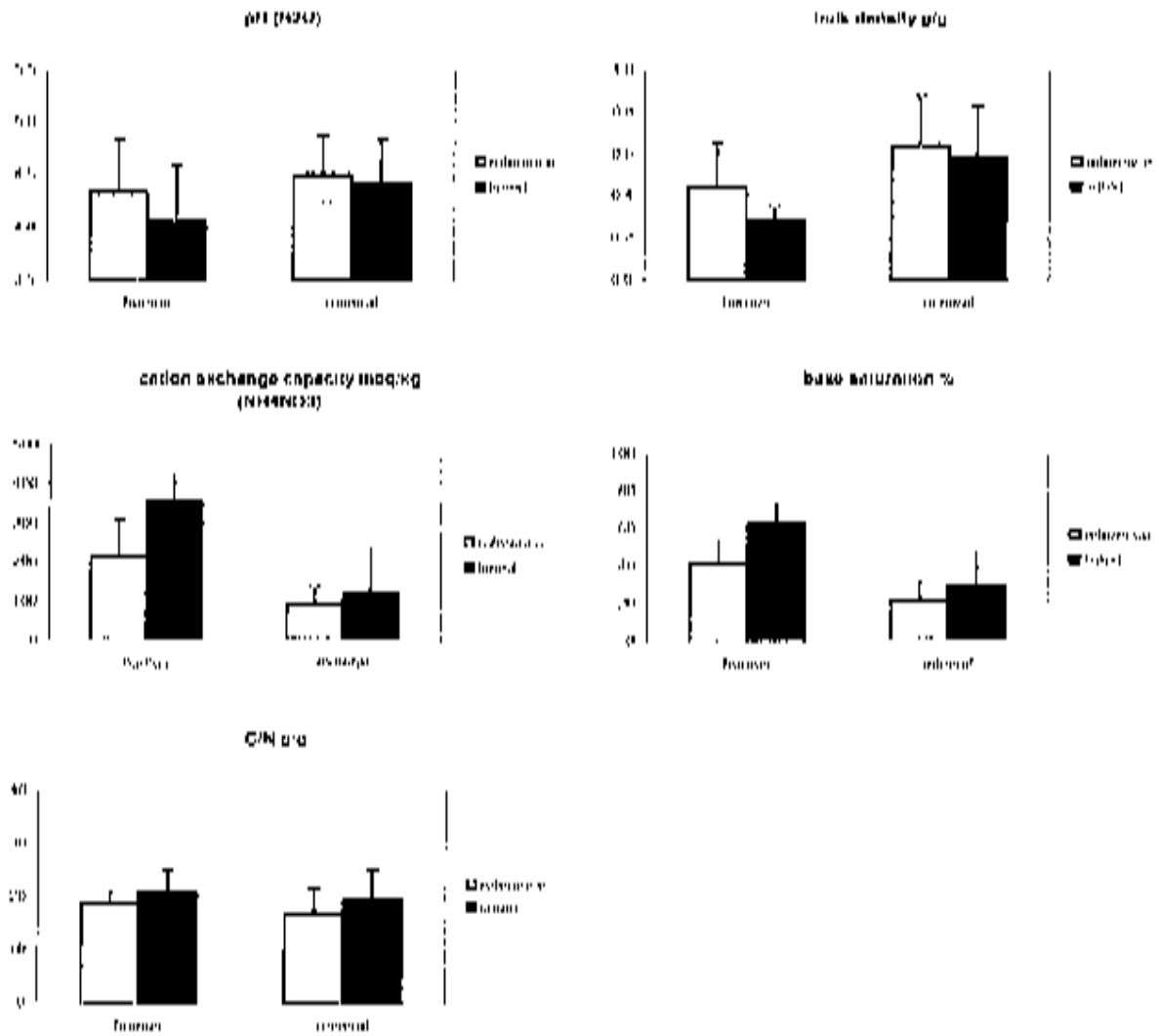
Stations FU 44 and FU 45

These stations are located in the bottom of a small valley with a mixed forest of coniferous and deciduous trees, among them a number of birches. The ground water level is high and the lysimeter bottles were filled up with water within a few hours.

Stations FU 48 and FU 49

Station FU 48 was placed in a boggy soil close to a stream, while FU 49 was located higher up on the ground in between the roots of a big alder. The lysimeter FU 49 seldom gave water samples at all.

Taken together the 6 soils at the 3 sites in the reference catchment are similar chemically to the 6 soils of the 3 sites in the limed catchment (**Figure 17**). The heterogeneity of the sites is reflected in the large standard deviations about the means. The soils are typical for coniferous forests in Norway, although they have rather low C/N ratios. This may reflect the long-term deposition and storage of N in the soil.



**Figure 17.** Soil chemistry at the 6 sites in the reference and limed catchments at Gjerstad, sampled in summer 1994 prior to liming. Mean and standard deviation of the samples (n = 6 for humus; n = 12-15 for mineral soil) are shown.

Table 8. a. Preliminary composition of organic soil in three locations in the reference catchment (SP) and three locations in the limed catchment (FU) B.D. = bulk density; L.O.I. = loss on ignition; KjeldN = KjeldahlN.

HUMLES GJERSTAD, 1984

OBS	SITE	pH	B.D. g/g	Dry wt. %	L.O.I. %	NjeldN mg/kg	CN g/g	total analysis ICP						
								Al mg/kg	Ca mg/kg	K mg/kg	Mg mg/kg	Na mg/kg	S mg/kg	
35	SP25	4.56	1.576	60.4	42.76	1157	16.18	17.55	23.6	19.2	104.3	6.79	55.62	
36	SP21	4.82	0.73	59.3	91.53	1111	6.73	9.77	23.5	18.4	17.3	6.28	31.52	
44	SP27	3.57	3.236	60.1	85.95	1454	21.65	160.2	33.5	24.9	16.9	8.25	7.26	
45	SP46	3.6	3.263	62.7	74.38	1069	26.26	114.2	52.5	23.8	23.6	4.55	50.28	
55	SP46	4.44	3.578	64.7	73.59	636	17.26	42.77	62.6	22.4	153.5	6.24	30.51	
60	SP56	4.42	3.757	67.6	15.25	386	14.21	152.9	42.1	12.8	40.6	4.56	7.51	
67	FU26	3.55	1.12	60.1	83.49	1051	29.57	50.66	93.3	25.7	21.1	8.62	68.34	
68	FU27	3.56	3.231	60.6	86.33	1141	27.02	61.59	75.5	22.3	21.4	7.73	62.16	
113	FU44	4.51	3.253	53.9	91.72	2051	15.69	403.2	134.2	25.5	23.5	12.51	111.6	
114	FU45	4.55	3.255	55.1	84.54	1359	19.39	513.1	73.4	22.7	15.6	9.22	33.99	
118	FU45	3.79	3.354	62.7	68.47	1430	17.34	112.2	42.2	19.7	44.5	6.72	3.34	
119	FU49	4.12	3.549	62.5	61.35	1323	26.57	51.24	111.8	15.2	62.7	7.4	55.34	
control	SP	b	b	b	b	b	b	b	b	b	b	b	b	
average	SP	4.24	0.446	62.70	67.25	1063.77	19.65	376.77	39.63	19.92	38.62	6.24	53.44	
min	SP	3.60	0.24	63.30	52.5	386.09	15.11	114.23	23.52	10.63	15.30	4.66	16.53	
max	SP	4.56	0.77	57.60	69.55	1777.00	26.26	1255.00	62.60	26.90	196.90	8.25	61.42	
S.D.	SP	0.51	0.22	3.42	29.52	535.34	4.66	453.31	13.29	5.76	36.10	1.24	26.05	
control	FU	b	b	b	b	b	b	b	b	b	b	b	b	
average	FU	4.03	0.29	56.55	90.26	1429.59	20.57	256.53	69.62	21.32	29.65	8.91	75.07	
min	FU	3.55	0.22	53.10	51.48	1091.00	13.39	39.66	39.22	13.22	16.60	6.72	33.34	
max	FU	4.53	0.35	65.90	61.72	2061.00	29.57	520.10	124.22	26.70	49.70	12.51	121.60	
S.D.	FU	0.54	0.06	2.11	17.27	353.40	3.66	325.27	29.31	3.84	13.69	2.34	23.77	

**Table 8. b.** Preliminary composition of organic soil leivatus  $\pm$  NaNO<sub>3</sub> (M) in three locations in the reference catchment SP and three locations in the limed catchment FU. CEC = cation exchange capacity; BS = base saturation.

OBS site	CEC meq/kg	BS %	exchangeable ions							
			Al mmol/kg	Ca mmol/kg	K mmol/kg	Mg mmol/kg	Na mmol/kg	Zn mmol/kg	S mmol/kg	
34 SP24	190.5	31.6	34.25	12.79	3.39	4.25	3.48	6.1	4.1	
35 SP20	230.0	24.7	62.94	35.64	14.77	13.44	3.47	6.1	5.02	
44 SP27	356.4	32.3	55.5	29.45	24.31	12.22	7.45	2.85	6.29	
45 SP40	283.5	48.7	14.12	46.21	17.53	13.94	7	2.85	3.72	
53 SP45	189.5	51.4	31.12	35.05	7.41	5.85	1.56	3.15	3.12	
64 SP50	366.2	57.1	9.33	20.93	3.55	4.07	0.92	3.15	1.24	
67 SP35	427.6	54.7	7.15	73.6	11.32	17.32	8.54	2.85	3.31	
68 SP37	333.5	52	16.7	22.99	17.51	17.33	7.26	3.12	4.07	
112 SP41	446.4	71.4	35.74	12.73	22.53	13.22	12.57	3.45	6.64	
114 SP43	322.5	58	47.54	47.84	14.07	3.47	3.76	0.1	3.22	
115 SP45	265	57.6	25.4	40.62	15.19	2.52	3.4	2.95	5.14	
115 SP49	326.5	79.4	4.15	50.25	12.25	25.74	3.2	2.35	3.55	
total	5	6	5	6	6	5	6	5	6	
average	215.32	46.57	34.67	30.23	12.53	5.65	3.62	3.77	4.09	
min	95.22	24.0	9.33	13.79	3.55	4.07	0.92	2.15	1.24	
max	355.42	57.10	63.92	46.21	24.31	12.94	7.45	2.85	6.29	
S.D.	95.14	13.29	24.25	11.32	7.39	4.13	2.42	1.05	1.57	
total	5	6	5	6	6	5	6	5	6	
average	332.65	52.52	23.23	29.63	12.26	12.01	5.12	2.15	4.74	
min	263.02	52.29	4.15	40.62	12.25	3.47	3.20	2.12	3.55	
max	445.45	79.40	47.54	127.59	22.53	25.73	12.57	3.12	6.52	
S.D.	73.02	11.3	17.65	26.67	4.23	5.57	2.55	0.75	1.24	

**Table 9. 2.** Preliminary composition of mineral soils at three locations in the reference treatment (SP) and three locations in the limed treatment: (FCU, B.D. = bulk density; L.O.I. = loss on ignition; Kj-N = Kjeldahl-N)

OBS	sex	anal. analyses, 200												
		H/C	B.D. g/g	Density %	L.O.I. %	Kj-N mg/kg	calc. C/N	Al	Ca	K	Mg	Na	P	S
1	FCU	3.9	3.05	93.2	13.73	33	24.586	49.24	19.5	7.6	9	2.93	4	11.75
2	FCU	4.32	0.995	98.5	5.94	84	24.285	193.2	12	8.5	21.2	2.72	3.9	4.97
3	FCU	3.75	0.515	97.2	22.95	434	23.255	363.22	28.5	11.2	15.8	2.95	6.1	22.1
4	FCU	3.52	0.333	96.5	25.02	438	24.457	97.25	7	9.6	15.4	2.91	5.5	21.52
5	FCU	4.44	0.645	97	11	193	24.933	391.2	18.5	12.1	22.2	2.82	7.6	10.27
6	FCU	4.05	0.235	90.1	96.35	364	17.17	397	286.4	14.8	25.2	4.22	26.9	30.5
7	FCU	4.87	0.715	97.5	14.74	375	14.15	273.1	19.9	4.4	19.6	2.51	8.5	29.51
8	FCU	4.81	0.225	93	94.24	245	16.521	1120	45	5.6	12.5	2.55	47.5	37.52
9	FCU	5.26	0.305	99.1	29.7	182	18.055	197.6	52.1	7.8	21.9	2.53	5.2	39.77
10	FCU	4.25	0.334	93.8	95.96	175	15.699	197.9	49.6	5.6	21.5	2.5	54.5	35.28
11	FCU	4.545	0.545	96.9	23.1	182	14.555	295.7	29	11.5	33.9	2.45	12.6	25.48
12	FCU	4.54	0.51	97.6	8.55	175	17.51	395.9	24	12.5	12.8	2.56	7.6	11.82
13	FCU	4.26	0.67	97.6	15.44	425	13.151	293.1	53.6	8.5	16.4	2.59	13.4	22.25
14	FCU	4.55	0.735	97.7	8.77	186	15.692	511.6	52.5	10.2	20.1	6.44	8.9	25.92
15	SP1	4.99	0.545	99.5	57.21	232	14.268	1411	52.1	12.2	30.5	2.7	23.1	26.39
16	SP1	4.75	0.75	99	66.25	236	14.97	142	222	13.2	30.5	2.72	44	31.15
17	SP1	4.51	0.519	97.5	31.51	225	15.205	136	22.2	13.2	22.9	2.75	28.1	24.66
18	SP1	4.55	0.512	97.5	22.25	1890	15.821	197.6	54.1	5.5	33.4	2.59	52.5	37.21
19	SP1	4.55	0.625	94.5	26.25	479	15.414	1285	26.2	12.9	34.2	2.61	22	28.55
20	SP1	5.11	0.495	97.1	25.29	244	15.2	193.5	10.2	6.4	12.9	2.45	7.9	23.24
21	SP1	4.25	0.625	98.8	5.71	166	15.969	186.9	7.4	4.2	5.2	1.86	4	4.5
22	SP1	4.25	0.661	95.8	4.74	302	16.426	185.4	5.7	9.2	5.5	1.52	3.7	4.85
23	SP1	4.29	0.531	98.1	22.25	222	22.294	189.2	12.4	22.2	6.2	2.74	3.1	29.64
24	SP1	3.9	0.922	94.9	5.4	69	25.461	148.5	3.5	1.5	2.1	2.15	3.8	3.39
25	SP1	4.35	0.62	93.5	21.95	294	28.561	347	50	15.4	15.5	2.79	22.9	26.95
26	SP1	4.73	0.665	96	13.13	249	15.561	645.5	22.9	10.1	14.2	2.99	5	15.13
27	SP1	4.47	0.61	95.1	20.85	252	15.501	346.4	28.1	7.4	22.1	2.68	16.2	14.19
28	SP1	4.77	0.77	95.2	26.65	246	15.435	191.1	22.9	8	22.4	2.75	11	11.85
29	SP1	4.93	0.55	94.2	6.15	241	15.945	222.1	25.1	6.5	21.9	4.1	14.7	12.29



MINERAL Soils: GJERSTAD.  
1994

OBS	SITE	PH H <sub>2</sub> O	B.D. g/g	Dry wt. %	L.O.I. %	K <sub>2</sub> O %	N %	C %	total analysis ICP					S				
									Al	Ce	K	Mg	Na		P			
	total	12	14	14	14	14	14	14	14	14	14	14	14	14	14	14	14	14
av	av	4.63	0.95	92.7	31.7	6.62	19.4	363.6	37.4	3.6	52.5	3.6	18.0	35.7				
min	min	3.73	0.77	89.0	2.9	3.46	12.4	49.2	11.1	4.4	9.0	1.1	2.9	4.6				
max	max	5.16	1.2	93.6	90.4	133.4	23.3	1378.0	136.4	14.8	109.1	6.4	34.3	103.8				
S.D.	S.D.	0.46	0.23	4.03	23.01	573.36	4.04	650.38	30.77	2.49	59.23	1.27	13.90	37.23				
	HF	12	15	15	15	15	15	15	15	15	15	15	15	15				
av	av	4.51	0.84	93.4	23.3	52.9	17.1	539.5	24.9	11.6	65.6	3.5	23.3	25.0				
min	min	3.89	0.71	87.3	4.7	99.6	14.3	253.4	7.4	4.7	9.2	1.8	3.7	3.9				
max	max	4.91	0.95	94.9	73.1	1650.0	23.6	1506.0	50.0	23.1	146.3	6.6	32.3	57.2				
S.D.	S.D.	0.27	0.06	3.06	21.89	533.79	2.93	533.37	13.46	4.59	31.44	1.23	11.19	29.13				

**Table 9. b.** Preliminary composition of mineral soils (means in  $\text{NH}_4\text{NO}_3$  (M) at three locations in the reference catchment (SP) and three locations in the timed catchment (FC). CEC = cation exchange capacity; BS = base saturation.

OBS site	sample	CEC cmol/kg	BS %	exchangeable ions												
				Ca mmol/kg	K mmol/kg	Mg mmol/kg	Na mmol/kg	P mmol/kg	S mmol/kg	AC mmol/kg	Cl mmol/kg	AN mmol/kg	AS mmol/kg			
1 FC06	1	62.1	37.5	5.15	4.24	2.3	0.57	0.15	0.49							
2 FC06	2	46.7	5.3	15.52	0.65	0.29	0.22	0.1	0.37							
3 FC07	1	126.4	34.8	15.95	3.55	3.9	6.75	0.1	0.77							
4 FC07	2	126.8	33.3	15.57	2.53	3.59	6.77	0.1	0.55							
5 FC07	3	59.1	7.3	19.31	1.11	0.37	0.22	0.1	0.9							
6 FC44	1	467.3	69	43.72	25.73	21.18	3.56	0.97	3.94							
7 FC44	2	55.6	23.4	15.92	2.31	0.54	0.42	0.1	0.66							
8 FC45	1	242.6	33.8	35	9.37	4.47	1.9	0.1	2.95							
9 FC45	2	133.5	50	24.88	5.5	2.69	1.66	0.1	2.96							
10 FC45	3	132.9	45.5	25.52	4.97	2.71	1.72	0.1	2.21							
11 FC48	1	59.8	28.7	19.35	3.44	4.01	0.55	0.1	1.7							
12 FC48	2	45.5	16.1	13.2	2.09	0.55	0.35	0.1	0.95							
13 FC49	1	73.3	30.3	15.73	3.11	3.1	0.77	0.1	1.5							
14 FC49	2	44.5	8.5	14.95	1.31	0.63	0.32	0.1	1.5							
15 SP25	1	159.4	32.3	36.32	9.15	3.06	1.58	0.1	2.25							
16 SP25	2	158.6	32.1	33.62	8.59	3.1	1.48	0.1	2.05							
17 SP25	3	33.5	23.8	22.39	2.56	1.65	0.87	0.1	2.51							
18 SP30	1	150.5	33.2	45.15	7.74	4.24	1.71	0.1	2.62							
19 SP30	2	60.6	23.1	17.62	2.36	0.97	0.72	0.1	1.55							
20 SP37	1	129.5	35.7	24.55	4.1	2.28	0.65	0.1	0.9							
21 SP37	2	46.7	6.5	14.95	0.77	0.29	1.25	0.1	0.22							
22 SP37	3	43.6	6.3	15.75	0.68	0.29	1.26	0.1	0.25							
23 SP40	1	93.9	24.5	16.57	3.19	2.42	0.58	0.1	0.84							
24 SP40	2	33	9.5	14.63	0.54	0.64	0.28	0.1	0.26							
25 SP46	1	34.6	3.1	19.02	3.53	2.58	0.51	0.1	1.95							
26 SP46	2	54.5	4.6	13.51	1.67	0.59	0.41	0.1	1.45							
27 SP50	1	86.7	24.7	19.59	2.94	2.05	0.41	0.1	1.11							
28 SP50	2	84.5	22.9	19.1	2.53	1.57	0.37	0.1	1.12							
29 SP50	3	46.9	12.6	12.47	1.31	0.81	0.37	0.1	0.69							

OBS	SITE	CFC µg/kg	BS %	Al mmol/kg	Ca mmol/kg	exchangeable ions					S mmol/kg	
						K mmol/kg	Mg mmol/kg	Na mmol/kg	P mmol/kg	S mmol/kg		
count	FL	14	12	14	14	14	14	14	14	14	14	14
average	FL	222.18	30.2	21.9	18.7	4.5	3.7	1.9	0.2	0.2	0.2	0.2
min	FL	44.56	5.3	5.2	0.4	0.7	0.3	0.2	0.1	0.1	0.1	0.1
max	FL	467.36	63.0	55.1	115.0	16.7	21.1	3.6	1.0	1.0	1.0	1.0
S.D.	FL	113.18	17.06	12.54	33.07	4.19	5.25	0.93	0.23	0.23	0.23	0.23
count	SP	15	15	15	15	15	15	15	15	15	15	15
average	SP	90.28	22.1	21.9	6.8	3.5	2.1	0.7	0.1	0.1	0.1	0.1
min	SP	46.70	6.2	10.7	0.8	0.6	0.3	0.1	0.1	0.1	0.1	0.1
max	SP	180.90	34.7	48.2	14.4	6.2	5.1	1.7	0.2	0.2	0.2	0.2
S.D.	SP	45.55	9.96	11.90	5.02	2.88	1.60	0.59	0.03	0.03	0.03	0.03

#### 4.4 Soil solution

The lysimeters were installed in September 1994. Samples collected through 31 October 1994 were disregarded here due to possible effects of disturbance caused by the installation.

The chemical composition of soil solution varies widely between the sites, in part reflecting the heterogeneity of soil and site properties. For example, pH in the humus layer was about 4 at site SI00, FU 06 and FU 07, but 5.5–6.0 at site SI28, FU 44 and FU 45 (**Figure 18** and **Figure 19**). There were no trends over time in pH at any of the lysimeters in either the limed or reference catchments.

Mg concentrations in soil solution also varied from site to site (**Figure 20** and **Figure 21**). At 5 of the lysimeters in the limed catchment the Mg concentrations show a statistically significant ( $p < 0.05$ ) increase during the 3 years following liming. There were no significant trends in Ca concentrations (**Figure 22** and **Figure 23**).

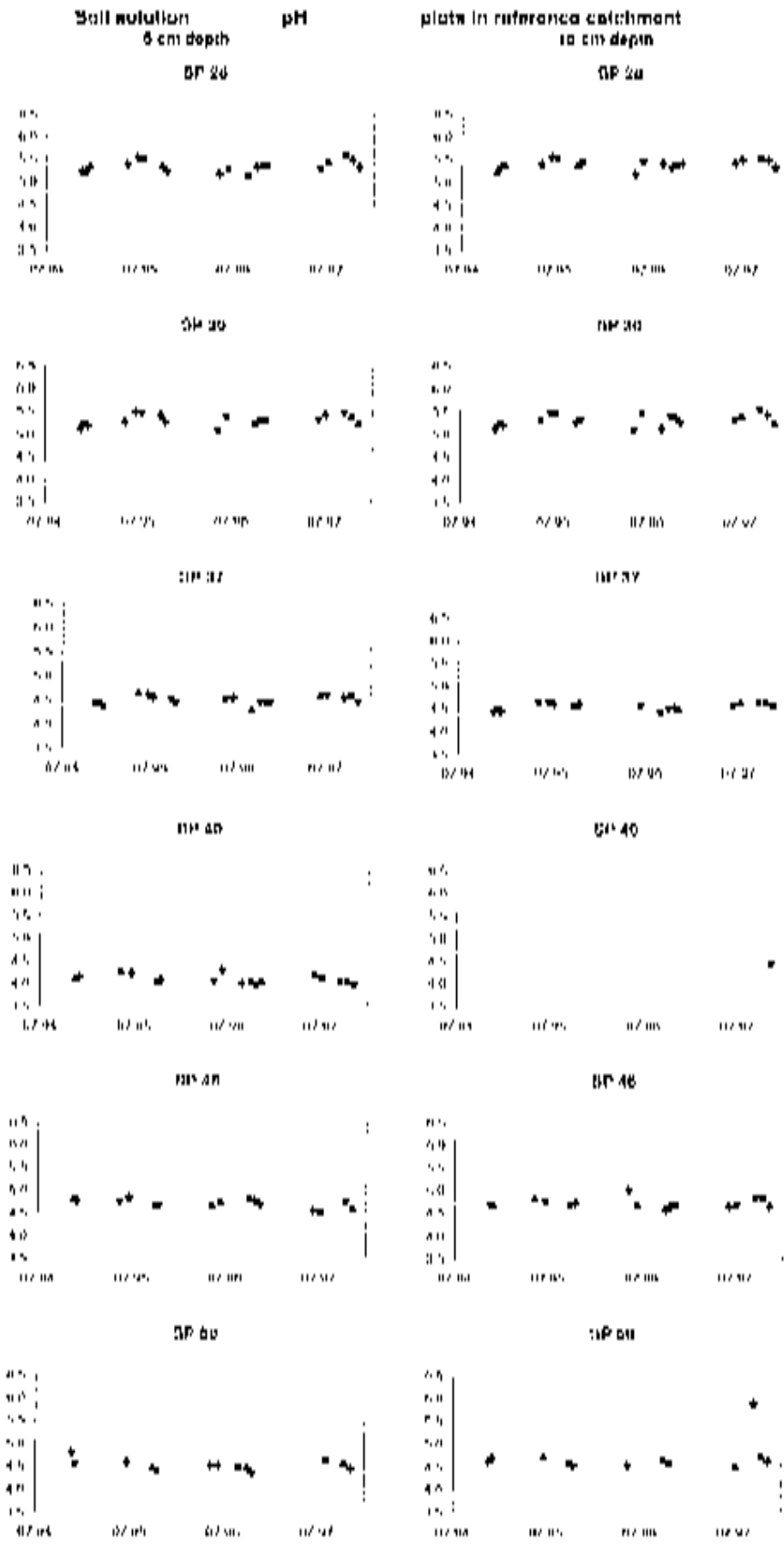


Figure 18. pH in soil solution in plots in the reference catchment.

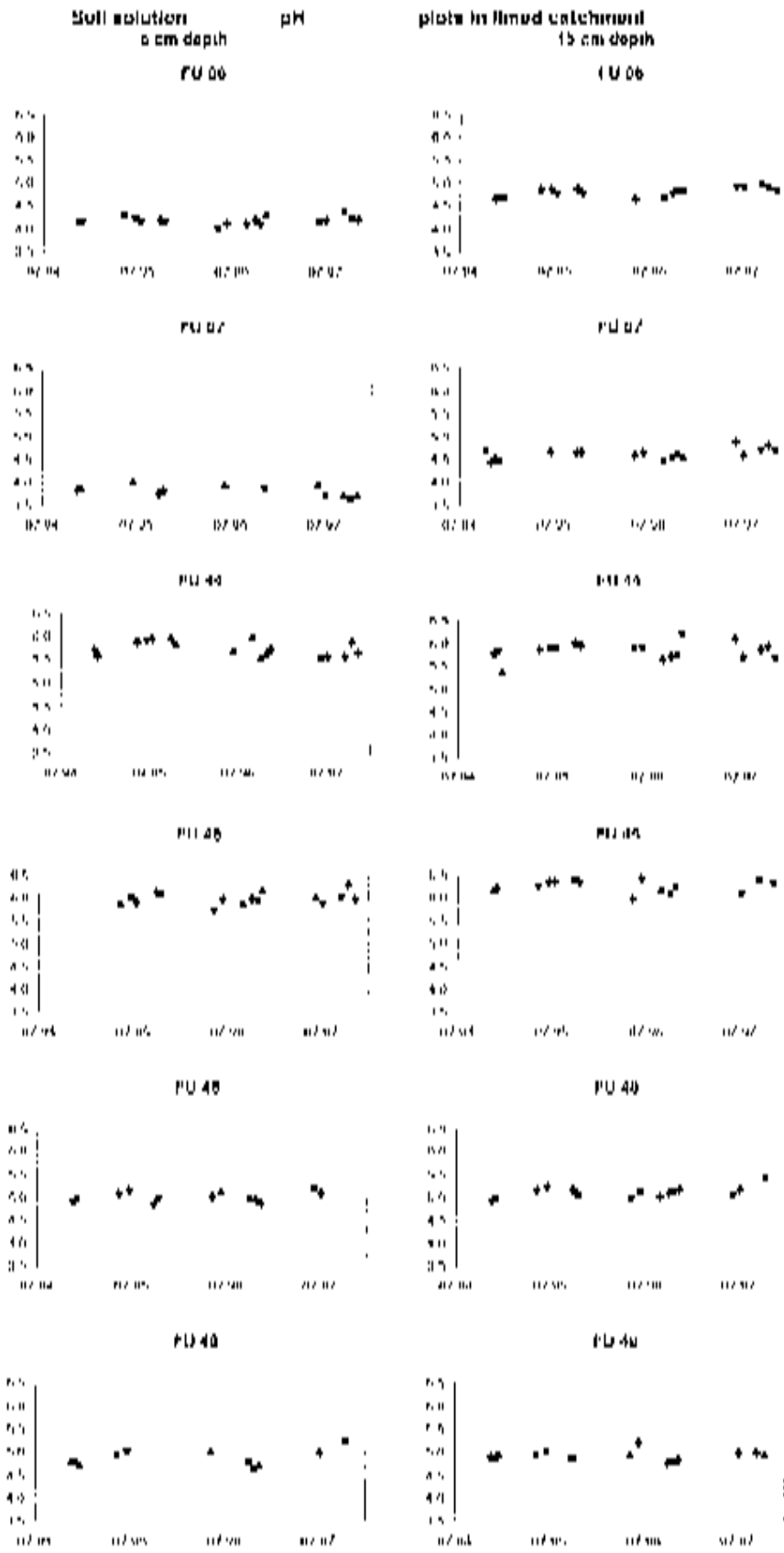


Figure 19, pH in soil solution in plots in the limed catchment.

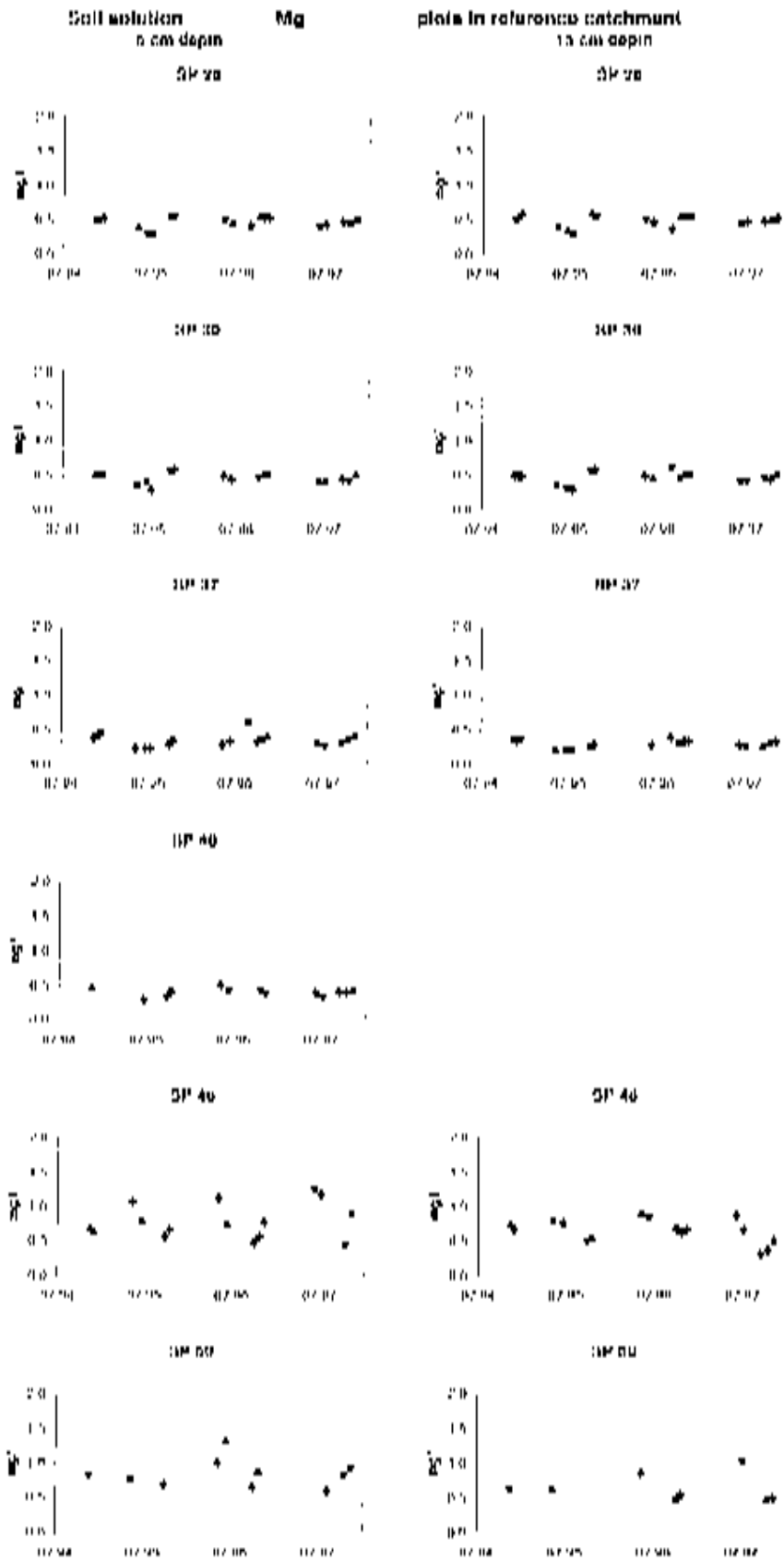


Figure 20. Mg concentrations in soil solution in plots in the reference catchment

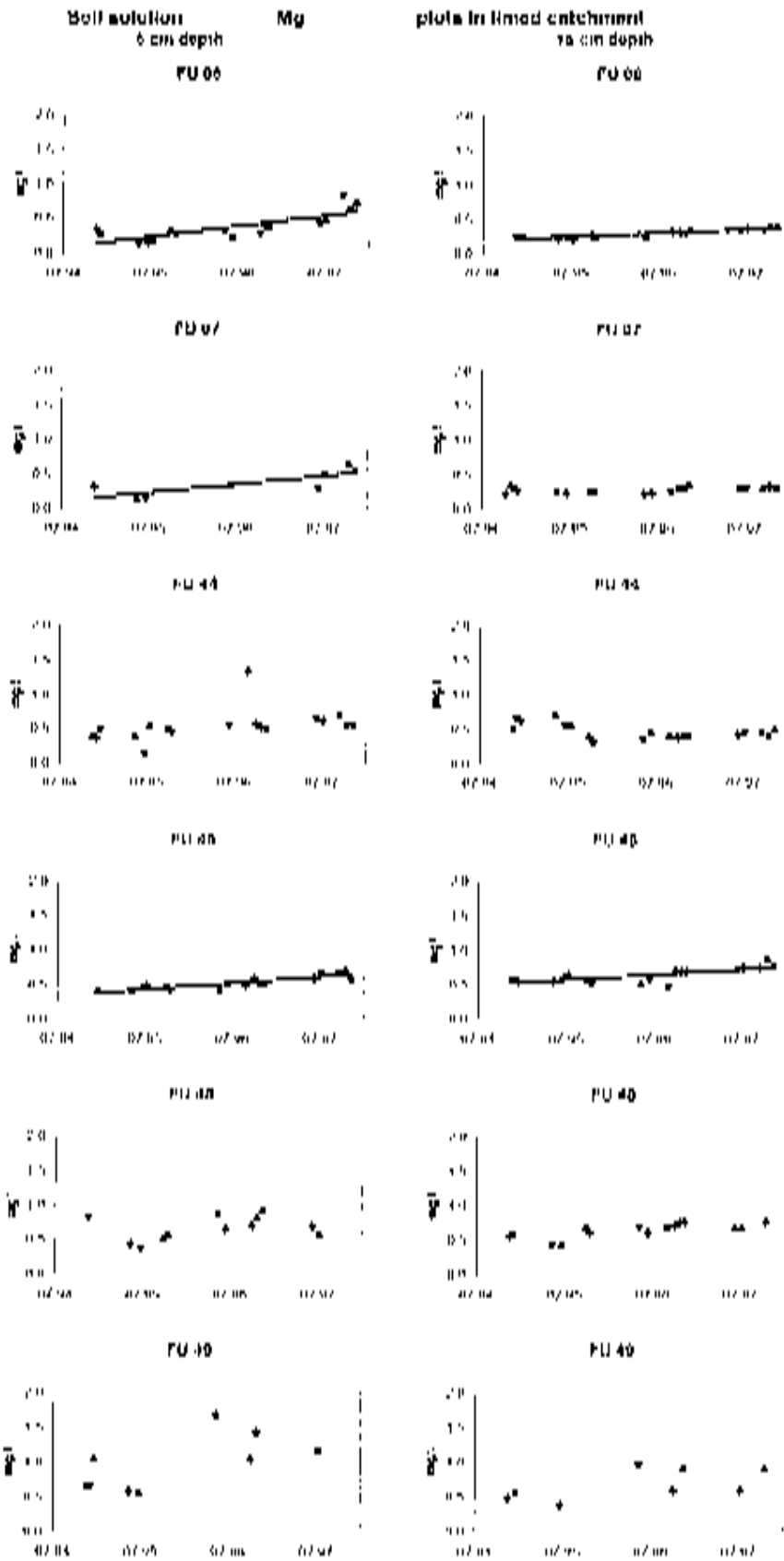


Figure 21. Mg concentrations in soil solution in plots in the lined catchment. Significant linear trends indicated by regression lines.



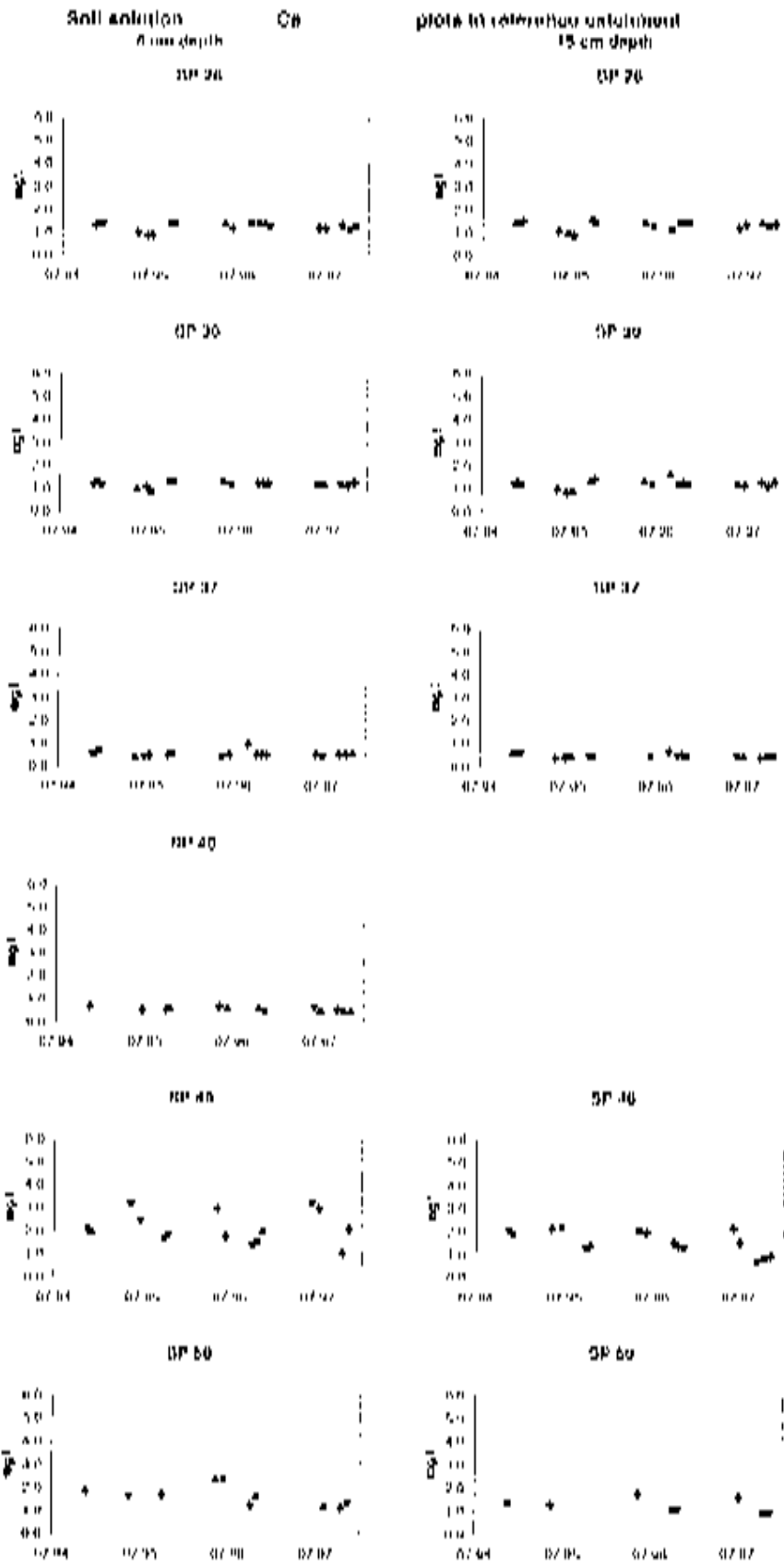


Figure 22. Ca concentrations in soil solution to plots in the reference catchment.

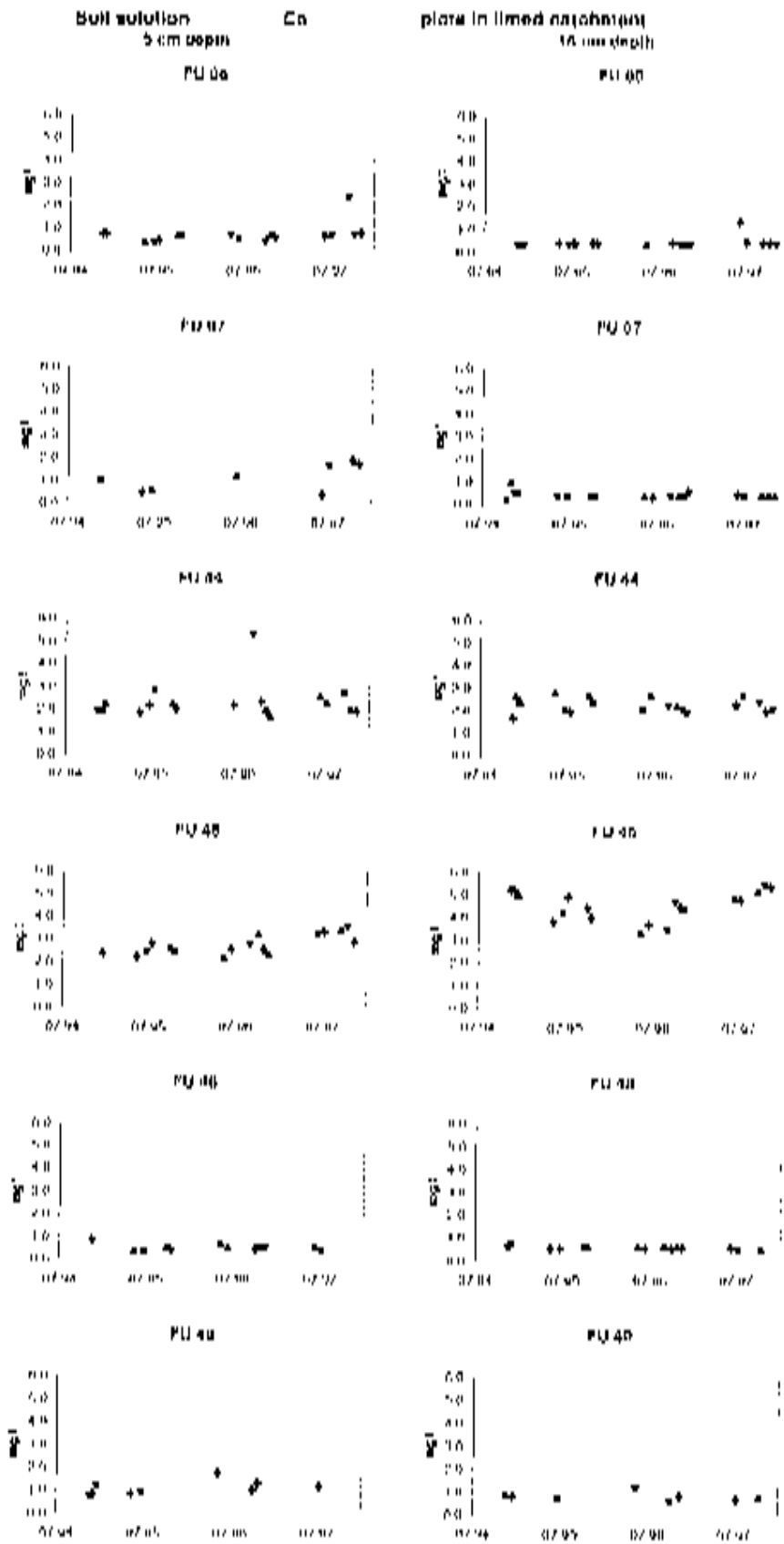
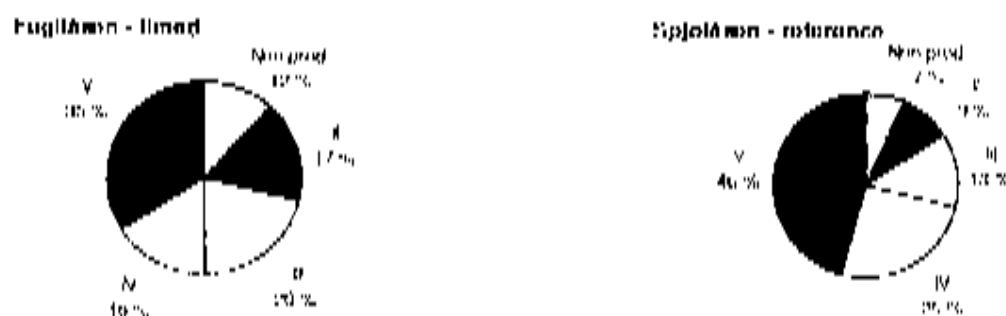


Figure 23. Ca concentrations in soil solution in plots in the lined catchment.

## 4.5 Forest investigations

### 4.5.1 Forest condition

In **Figure 24** the distribution of different cutting classes is shown in the two catchments. The sampling intensity is approximately 3% for both areas, and the stand composition is fairly alike. In the reference area there is less young and middle aged forest. Here, about 70% of the forest area is classified as old production forest or old forest, while in the limed catchment this proportion is 50%.



**Figure 24.** Area distribution of different cutting classes in the two catchments. Cutting classes according to following definition: I - Forest under regeneration, II - Regenerated areas and young forest, III - Young production forest, IV - old production forest, V - old forest

**Table 10** and **Table 11** show aggregated figures for forest condition in the two catchments. The share of Scots pine to Norway spruce is higher in the reference area (20/30) compared to the limed area (56/44). A small fraction of broad-leaved trees in the middle aged and old stands is typical for this region. The clear-cut areas in the catchments are mainly dominated by birch or trembling aspen.

**Table 10.** Some key figures for the forest stands in the two catchments.

	Fugleåsene - limed area	Sjøfåsene - reference area
Area	84.1 ha	40.8 ha
Total standing volume	9920 m <sup>3</sup>	6027 m <sup>3</sup>
Mean diameter	140 mm	157 mm
Mean height	10.4 m	11.4 m
Average volume per tree	0.157 m <sup>3</sup>	0.187 m <sup>3</sup>
Mean age	67 yr	77 yr
Number of sample plots	122	56
Sample intensity	2.9%	2.7%
Site index (II <sub>60</sub> )	13.4	11.8

**Table 11.** Volume in different cutting classes and by tree species in the two catchments.

Engelåsen - lined catchment							
Volume (m <sup>3</sup> )							
Cutting class	Spruce	Pine	Birch	Aspen	Oak	Other broadleaves	Sum
III	230	445	177	177	124	18	1171
IV	703	990	92	144	101	3	2033
V	2516	3028	192	502	306	51	6592
Sum	3449	4460	461	823	531	72	9796

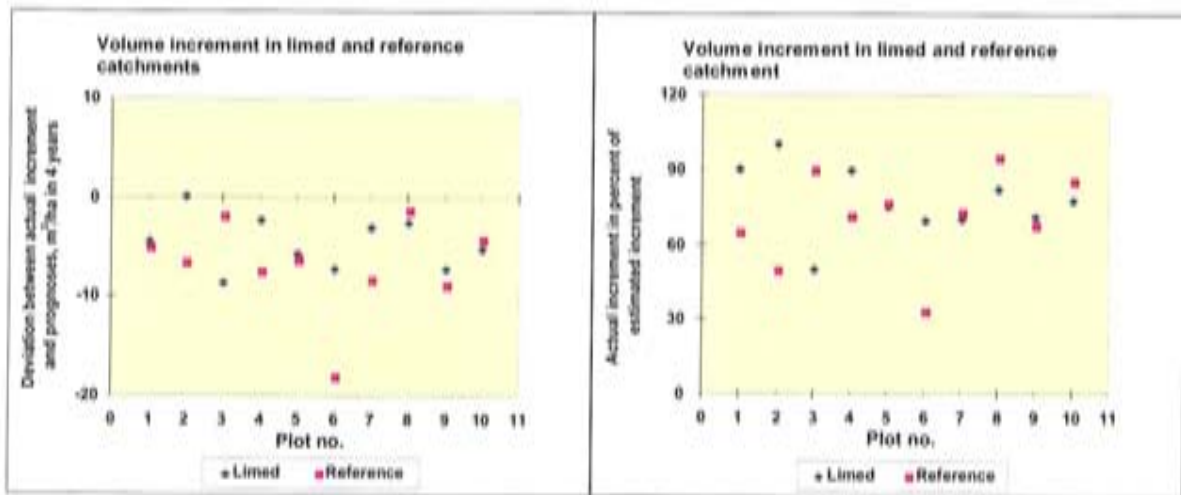
  

Spjotåsen - reference catchment							
Volume (m <sup>3</sup> )							
Cutting classes	Spruce	Pine	Birch	Aspen	Oak	Other broadleaves	Sum
III	48	231	8	108	16	3	414
IV	467	1015	36	114	1	1	1635
V	1050	2318	117	314	57	47	3803
Sum	1565	3564	161	433	74	50	5847

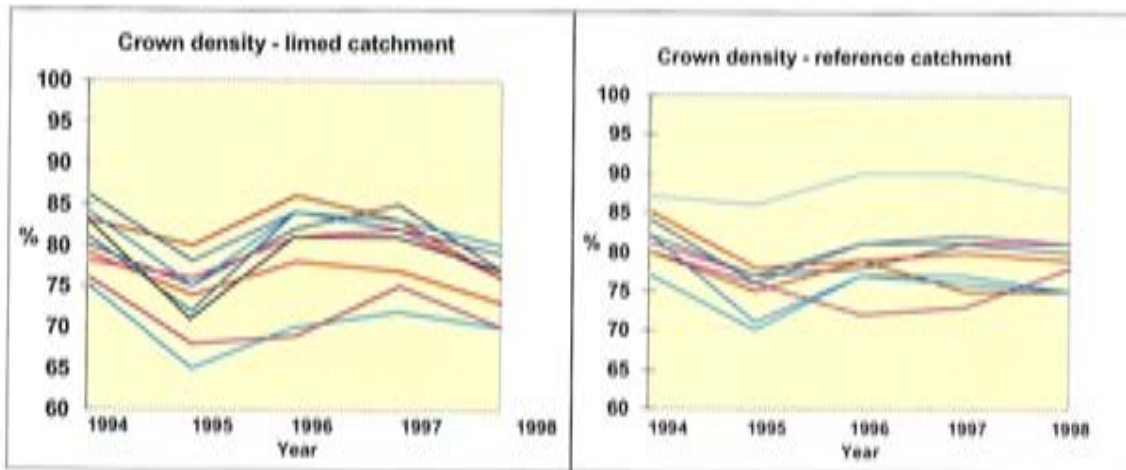
#### 4.5.2 Intensive monitoring plots

The increment in the 4-year period 1994-1998 is compared to the estimated increment based on increment function (**Figure 25**). The overall level of estimated increment is 70% for the reference catchment and 77% for the lined catchment. This difference is not significant. As seen from **Figure 25** the variation in residuals is rather large. One plot, no 6 in the reference catchment, has a very low increment compared to that expected from the function. This is probably caused by low vitality and declining trees. The reason for this has not been investigated further.

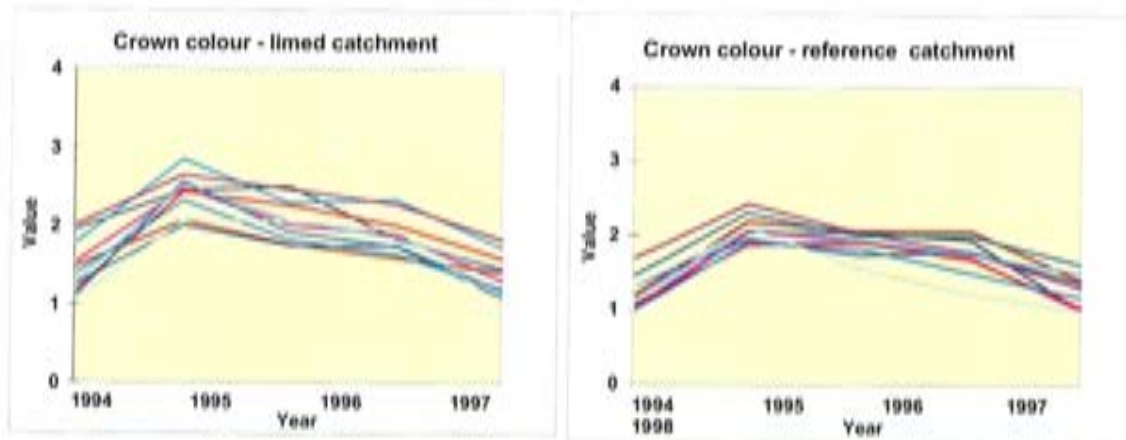
The average crown density and crown colour are shown in **Figure 26** and **Figure 27**. The crown density is also relatively low, with many of the stands with less than 85% crown density for most of the period. The level varies little from year-to-year, no effect of timing can be extracted. The same holds for the crown colour. The year-to-year variation is more pronounced for crown colour. An increase in colour value (more yellow trees) was found in 1995. Then a steady decrease in value (more green trees) has been noted in both catchments.



**Figure 25.** Differences between measured and estimated increment on the intensive monitoring plots. Residuals (m<sup>3</sup> ha<sup>-1</sup> in 4 years) (left) and measured increment in percent of estimated (right).

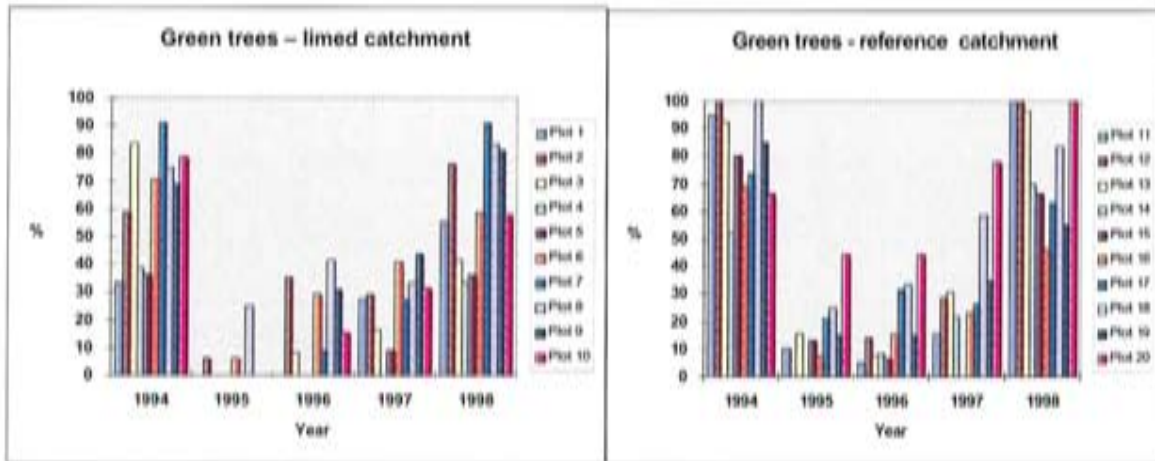


**Figure 26.** Average crown density on intensive monitoring plots in the period in the two areas.



**Figure 27.** Average crown colour in the intensive monitoring plots during the period in the two catchments.

In **Figure 28** the percentage amount of green trees in each plot can be seen. For most of the plots an almost complete recovery in crown colour compared to 1994 has occurred. Still some stands have low proportions of green trees.



**Figure 28.** The amount of green trees on the intensive monitoring plots in the limed catchment and reference catchment in the different years.

## 5. Discussion

The choice of coarse dolomite and the dose of  $3 \text{ t ha}^{-1}$  for liming the 80 ha forested ecosystem in Gjerstad was partly based on recommendations for forest soil liming in Sweden, as summarised by Nihlgård *et al.* (1996). Results from the whole-catchment liming at Tjønnsstrond (Traaen *et al.* 1997) with the same dose of more fine-powdered calcite indicated that significant improvement of the water quality was likely. Higher doses may change the soil pH too much, whereas ion exchange of accumulated Al and  $\text{H}^+$  in the soil with the added Ca and Mg and production of organic acids from decomposition of organic matter may increase the acidity of the soil solution if lower doses are used.

Due to the anticipated variability of many important factors for acidity and liming effects within heterogeneous catchments, mixed forests and climatic regions, we chose a fixed dose and a paired catchment experiment in a forest ecosystem that might be regarded as typical for large acidified areas in Norway. An initial split in the limed catchment for stream water analyses and parallel sampling of soil solution at different stations were included to study in-catchment variability.

### 5.1 Soil chemistry

Sites for soil chemistry was selected mainly for vegetation purposes. Large gradients in nutrient status and water supply were two central criteria for this selection. As a result large variability in most of the soil chemical compounds was expected, and this is indeed what was found.

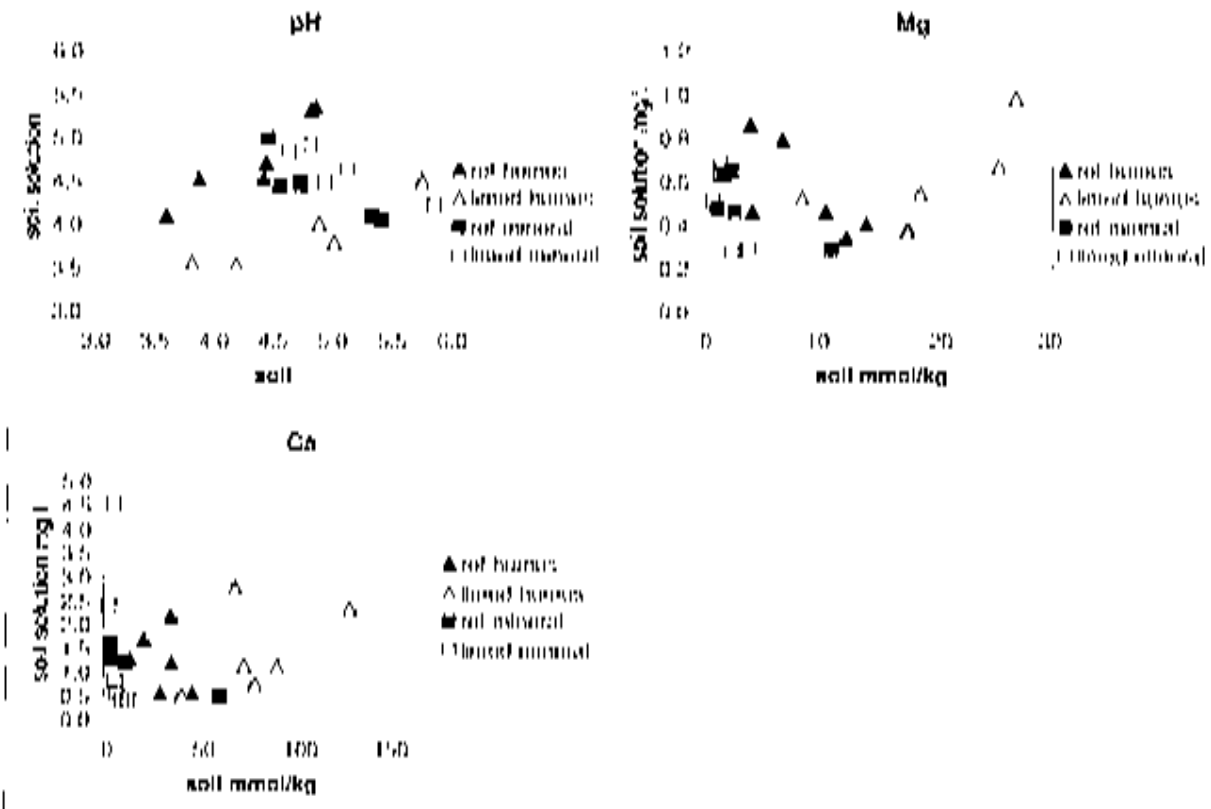
The chemistry of the soils at the two Gjerstad catchments is typical for soils in southern Norway in coniferous forests on nutrient-poor, siliceous moraine. Soils on well-drained sites are acidic and exhibit podsollic character, while those on adjacent wetter areas are typically organic-rich and humified. Both the cation exchange capacity (CEC) and the base saturation, two key parameters with

respect to acidification of soil and mitigation by liming, are well within the ranges expected for coniferous forest soils of southern Norway.

Whether liming has resulted in changes in soil chemistry such as increase base saturation remains to be seen. A resampling of soils is scheduled for 1999. Liming at Tjønnstrand, for example, resulted in significant increase in base saturation (Tjønnen *et al.* 1997).

## 5.2 Soil solution chemistry

In principal one should expect a close relation between chemistry of soil and chemistry of soil solution at each individual site. The wide range in soil chemistry should be reflected in a corresponding wide range of soil solution chemistry. This is in part the case for pH in the humus layer. In both catchments soils with low pH (water) yield soil solution with low pH (Figure 29). This is the case for the humus layer. The mineral soil horizons do not show this relationship, however. Similarly for both Mg and Ca there is a tendency for the humus layers with higher contents of exchangeable base cations to have higher concentrations in soil solution (Figure 30). Again there is no such relationship for the mineral horizons.



**Figure 29.** A comparison of soil and soil solution chemistry in the humus and mineral soil horizons in two catchments at Tjønerstad. The soils were sampled in summer 1994. The soil solution data are means for samples from each lysimeter collected November 1994 through November 1997.

During the first 2.5 years following liming there have been very few indications of changes in soil solution chemistry. A small increase in Mg at only some of the locations has taken place. This was expected due to the slow movement of the dissolved base cations in other experiments (Brahmer 1994; Nohstedt 1997). This contrasts with the rapid change of runoff water chemistry.

The most striking result from the Gjerstad experiment is the relatively uniform, acid and Al-rich runoff in the sub-catchments and the great variability in soil solution (Figure 30). This indicates that soil solution collected by the Poremat tension lysimeters is not representative of runoff. Similar observations on the lack of similarity between soil solution and runoff comes from other whole catchment experiments such as the RAIN (Hauhs 1998) and CLIMEX (Jenkins *et al.* 1996) projects at Risdalsheia near Grimstad, about 60 km SW of Gjerstad and the NITREX project at Gårdsjön, near Gothenburg Sweden (Staines and Kjømås 1998). In both these cases runoff responded more rapidly and consistently to changes in the chemical composition of deposition, while lysimeter data showed wide variations ranging from little or no response over many years, to immediate and persistent response. As at Gjerstad the lysimeter samples show great heterogeneity spatially.

There are several possible explanations for the lack of similarity between soil solution and runoff. First, the sampling technique for soil solution probably does not collect water in volumes proportional to that in runoff. Soil solution is comprised of water in pore spaces over a continuum of sizes, from macropores that have water only during saturated conditions, to micropores that retain water by capillary tension. The Poremat lysimeters under tension will collect water from a fraction of these pore sizes, but most probably some will be over-represented. The macropore flow will be underrepresented in volume. And since the lysimeters at Gjerstad were only evaluated 1 day each 14 days, the samples represent only a small time fraction as well.

Second, water movement varies greatly in time and space though heterogeneous uneven terrain such as is characteristic of the forested catchments at Gjerstad, and also Risdalsheia and Gårdsjön. The dominant pathway from soil surface to the stream is horizontal, not vertical. Incoming precipitation (or throughfall) percolates down through the soil undergoing chemical change underway until an impermeable surface is reached – either bedrock or a water-saturated zone. The water then moves horizontally towards the stream channel, and may alter chemical composition further, especially in the discharge regions near the stream. The runoff in the stream is a composite sample of water “packages” that have moved vertically and horizontally through the soil at a wide variety of distances and contact times.

Although lysimeter water is not representative for runoff, it may, however, give a good picture of the water quality available for uptake by roots. Thus the response of trees and ground vegetation to the liming treatment may first occur when the chemical composition of soil solution changes. At the NITREX experiment at Gårdsjön the runoff began to respond to N additions already during the first year of treatment, while the soil solution responded after 2-3 years and the vegetation yet later (Kjømås *et al.* 1998).

### 5.3 Runoff chemistry

In runoff, the most striking effects of the dolomite treatment at Gjerstad were increased pH and concentrations of base cations, slightly reduced Al transport and change of Al species from highly toxic to non-toxic species. These effects are probably due to dissolution of dolomite in the topsoil and subsequent ion exchange of base cations in the humic layer. Lateral surface and sub-surface flow transport excess and non-exchanged Ca and Mg to the stream.

Rapid, almost immediate, improvements in the stream water quality were found. This occurred even though the finest fraction of the dolomite was removed before spreading, in contrast to the liming material used in e.g. the Gårdsjön liming experiment, where the particles were in the range 0-30 µm, a particle size fraction that was almost completely removed from the Gjerstad dolomite.



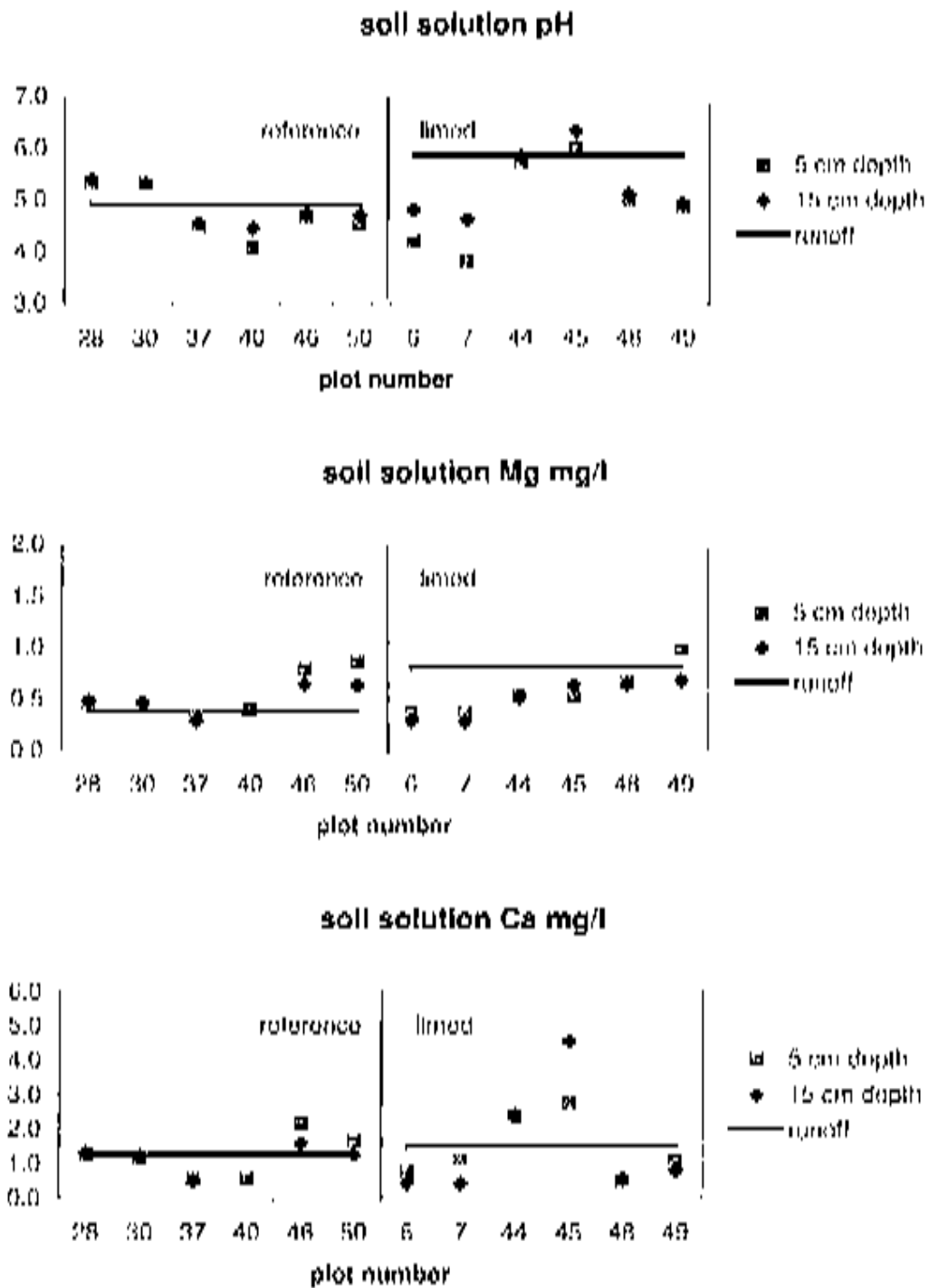


Figure 30. Comparison of mean concentrations in soil solution (6 sites at each catchment) and in runoff for samples collected during the post-liming period November 1994 through November 1997.

The post-liming increase in runoff concentrations of Ca and Mg was not significantly different in the Gjerstad experiment. Higher mobility of Mg relative to Ca was reflected in the relative increase. Whereas 0.5 % of the added Ca was transported out, 2.0 % of the added Mg was transported out of the limed catchment during the first 2.5 years.

Although the annual net transport of Ca and Mg was less than 1 % of the amount added, this was sufficient to cause a rapid response in runoff water quality after liming. The increase in water quality persisted throughout the period. No decreasing trends were documented after the initial increase in pH, base cations or ANC, and a longer monitoring period is needed to draw conclusions regarding duration and thereby mean annual liming costs.

These changes due to liming have turned the potentially toxic runoff into a water quality adequate for survival of acid sensitive organisms such as trout.

The results from Gjerstad generally agree with other comparable experiments with liming terrestrial areas. Hultberg *et al.* (1995) reported similar effects on runoff after application of 6 t ha<sup>-1</sup> of extremely fine ground dolomite (particle size range: 0-30 µm) to a small catchment at the Gårdsjö research site, Sweden. A significant reduction of inorganic Al and Mn of 40-50 % was found. As in Gjerstad the mean annual flux of Mg was less than 1 % (7 % in 9 years). Fransman and Nihlgård (1995) also found increased pH, Ca and Mg and reduced Al, Fe and Mn after forest soil liming. Other experiments from Sweden with moderate lime doses (3-4 t ha<sup>-1</sup>) showed only minor effects on pH and Al concentration in stream water during a 3-year period (Westling and Skarby 1993).

In Schuchsee, Black Forest in Germany a minor increase in stream-water Ca was registered one year after liming with 4 t ha<sup>-1</sup> of pelleted dolomite limestone, but the increase in Mg was as rapid as in Gjerstad (Bodmer 1994). Marked increases (after an initial peak) in both Ca and Mg in the soil solution at 30-cm depth were seen the year following liming. Increased Al-mobilisation at 30 cm was ascribed to the initial ion exchange processes after liming. Al did not increase in the stream water the first year after liming.

Kreutzer (1995) reported increased pH in the humic layer and in drainage water from this layer in the 6-year period after liming. No pH change in soil solution at 20 cm depth or drainage water from this layer was found. Drastic deprotonation of functional groups of the humic matrix was reported and both the cation exchange capacity and the base saturation increased. This stored buffer capacity in the topsoil. Six years after liming, dissolution of the added dolomite was regarded as 100 % and 70 % of the added Ca and 30 % of the Mg were still present in the humic layer. The low number for Mg reflects the more mobile nature of this element compared to Ca. Dissolution of lime was described by the equation:  $m(t) = A \cdot e^{-t/T}$ , where A is the mass of carbonate added in tonnes, m is mass of carbonates remaining in tonnes, and T is time in years. According to this equation 80 % of the added dolomite in Gjerstad should be dissolved at the end of 1996.

Liming of the heathland Tjønnstrand-catchment in Norway with a 3 t ha<sup>-1</sup> dose of calcite powder resulted in a rapid and long-lasting increase in both pH and Ca and a more than 50 % reduction in reactive Al (Tjønn *et al.* 1997). Liming of an upland, forested catchment (subcatchment IV) at Woods Lake, Adirondack Mountains, New York, USA, also resulted in improvement in water quality (Crimo and Driscoll 1996). Liming of wetlands may increase the water quality of runoff, although the duration may be significantly shorter due to the more favourable dissolution properties of flushed bog surfaces and effective draining of dissolved liming material (Hindar *et al.* 1996).

Although the results from some of these experiments indicate only minor changes in pH, base cations and Al, even small water quality changes or differences may be of significance for the survival of

acid-sensitive fish and invertebrates. The significance of minor water quality changes may also be important for the duration of adequate liming effects on streamwater, tending to increase duration and thereby reducing costs if being part of a liming programme for aquatic systems.

Potential undesirable effects, such as increased  $\text{NO}_3^-$  leaching and mobility of organically-complexed trace metals (Fe, Cu and Pb), may be expected after forest soil liming. So far, a significant but minor increase in the median concentration of  $\text{NO}_3^-$  due to liming was found after 2-5 years of liming at Gjerstad. The seasonal variability was large and concentrations in the limed catchment were within the concentration range in the reference most of the year. A clear increase in concentration was found the second winter after liming. This "winter/spring" trend was also found at Tjønnstrønd (Thunet *et al.* 1997). A possible explanation may be that liming in general stimulates decomposition of organic matter and that the result of this change is insignificant or masked during the first period after liming due to a delay in the build up of the microflora and uptake of the produced  $\text{NO}_3^-$  by the soil and vegetation during summer. Nitrification may still be significant in the snow covered soil during winter, however, and the accumulated  $\text{NO}_3^-$  from the dormant season may be flushed out with meltwater during spring melt, thus increasing the concentrations in late winter/spring. A longer data record is needed to draw any conclusion on the course of  $\text{NO}_3^-$  in this experiment.

Wardvinge *et al.* (1996) summarised liming effects on runoff from several forested catchments in Sweden and found no increased N leaching during periods of 9-10 years. An alternative hypothesis to increased N leaching after liming was pointed out, increased N leaching due to shortage of base cations after several years of nitrification was regarded as relevant for areas in Sweden that receive high amounts of atmospheric N. If liming would counteract this was not confirmed on, however.

At Schluchsee, increased  $\text{NO}_3^-$  in soil seepage from the humus layer and both 30 and 80 cm soil depth were recorded (Brahmer 1994). As a consequence of the pH increase of two units in the O horizon a shift in the population of nitrifiers from heterotrophic to autotrophic organisms took place (Feyer *et al.* 1995). All mineralised N was converted to  $\text{NO}_3^-$  in the limed topsoil, whereas  $\text{NH}_4^+$  dominated the inorganic N fraction in the reference. However, no change in streamwater was found the following 3 years after liming (Brahmer 1994; Feyer *et al.* 1995).

Kreutzer (1995) reported loss of N from the humus layer after 7 years of liming and also an increase of about  $1 \text{ mg l}^{-1}$  in the  $\text{NO}_3^-$  concentration in the drainage water leaving the root zone at 40-cm soil depth. The data record indicates a doubling of the  $\text{NO}_3^-$  concentration at this depth due to liming already after one year. This is explained by a transformation of organically bound N from the humus layer to dissolved organic N which was translocated to deeper soil depth and converted to  $\text{NO}_3^-$  by stimulated nitrification in the mineral soil. No data for streamwater was presented, but increased concentrations of drainage water  $\text{NO}_3^-$  was calculated from soil solution data and a flux model. N deposition in Högwald ( $12 \text{ kg N ha}^{-1}$ ) is of the same order of magnitude as in the Birkenes area but higher than at Sollhornfjell closer to the Gjerstad site.

Increased mobility of potentially harmful trace metals complexed in the humic layer, especially Cu and Pb, may be expected after forest soil liming, if liming results in increased decomposition of organic matter (Kreutzer 1995). However, mobility of trace metals also depends on soil pH and increasing pH may stop leaching of metals related to accelerated soil acidification (Hoff 1988). No increasing trends or elevated concentrations of the 10 trace metals relative to the reference stream were documented after the forest soil liming with dolomite in Gjerstad. This was probably due to the pH increase in runoff water, the moderate dose ( $3 \text{ t ha}^{-1}$ ) and low or moderate deposition of some of the trace metals in the area relative to e.g. German sites. Deposition of Cd, Cu and Pb in 1992-1993 at the Deuselbach monitoring station 200 km north of the Black Forest in Germany was higher than at Sollhornfjell in Norway (Berj *et al.* 1996) by factors of 6.8, 4.6 and 1.5, respectively. Increased pH will increase the retention of some metals, whereas organically-complexed metals not will be mobilised if decomposition of organic matter is insignificant for this to occur. Hoff (1988), in a

review of forest soil liming, concluded that the risk of trace metal mobilisation for limed soils is probably not more pronounced than for acidified soils.

The decrease in Mn, Zn and Co may or may not be desirable, depending on the resulting concentrations. Both Mn and Co are elements of importance for flora and fauna and increased retention, e.g. as the result of oxidation of inorganic monomeric Mn to MnO<sub>2</sub> at pH 5 and subsequent precipitation, may result in shortage. Mn limitation in lime-rich soils has been reported and appears to be common (Pearson and Adams 1967). In some acidified areas in Sweden large concentrations of dissolved inorganic Mn ( $> 1 \text{ mg l}^{-1}$ ) have been reported and a reduction is probably desirable. In Gjerstad, however, the concentrations are very low ( $5\text{--}15 \mu\text{g l}^{-1}$ ) and a reduction to  $< 5 \mu\text{g l}^{-1}$  after liming signals less availability of Mn for vegetation and forests and might be considered as a potential problem. For 985 Norwegian lakes the median Mn concentration was  $2.5 \mu\text{g l}^{-1}$  (Skjelkvåle *et al.* 1999), about the same level as measured at LIM-1 and LIM-4 the last 15 years. Higher concentrations should be expected in streams with higher concentrations of organic matter such as is characteristic of forested ecosystems.

The liming effects on soil solution and stream water depend on hydrology, soil permeability and topography, i.e. the contact between runoff and limestone material. Steep slopes, thin soils and high precipitation, as in the Gjerstad area, was expected to promote rapid response in stream water due to anticipated dominance of overland and subsurface flow. Soil stratification in hillslope profiles is characterised by compact basal layers (Feger 1999), which also may promote lateral flow. Poorly mineralised humus, forming hydrophobic layers in the organic top layer of the coniferous forest may act in the same way. Less clear cut effects in stream water in other forested ecosystems may be due to different topography, soil properties and climatic conditions. Possibilities of undesirable effects, like increased leaching of N and trace metals, is dependent on the deposition history of these elements in the actual sites. As deposition may vary to a great extent, e.g. between German and Norwegian sites, different results should be expected and this is indeed also what has been found. This confirms the rationale in selecting Norwegian sites for research activities before drawing conclusions on possible effects of forest soil liming in Norway.

Terrestrial liming, as the forest soil liming in Gjerstad and the alternative strategies represented by whole catchment liming of non-forested catchments (Traaen *et al.* 1997) and wetland liming (Hindri *et al.* 1996), has certain advantages compared to more traditional (in Norway) lake liming and lime dosing methods. The N transport from the acidified catchment to watercourses will be reduced, and a more stable water quality is achieved throughout the year. Low doses of coarse-grained dolomite will probably minimise the undesirable effects on water quality. Terrestrial liming, also the forest soil liming concept, may therefore represent an interesting supplement to other liming methods for aquatic systems. A longer data record is needed, however, to draw conclusions on NO<sub>x</sub> leaching and costs.

## 5.4 Forest condition

An overall conclusion from the tree stand investigations is that the liming has so far shown no significant effects on tree growth and tree vitality (crown density and crown colour).

The causes for the general low level of increment could be several. In specific periods deviation from the function may occur due to e.g. climatic factors. The increment function can be said to represent the average climate for a long period. Another explanation could be that the function is based on forest stands that have been treated regularly with thinning and therefore represent high vitality and high selvsko stands.

From other investigations we know that increment in old pine and spruce forests will be little or even negatively affected for more than 20 years after liming (Stad *et al.* 1996). To evaluate the forest growth more detailed increment analyses should be performed 10 years after liming. Then increment

cores should be taken and a calibration to pre-burning increment could be done. Then more information on long-term effects of this low dose dolomite application can be extracted.

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## Appendix A. Water and soil chemistry

### Abbreviation and references for sampling stations

LIM 1: Stream outlet of catchment 1 in the limed catchment Englehåsen

LIM 2: Stream outlet of catchment 2 in the limed catchment Englehåsen

LIM 3: Stream outlet of catchment 3 in the limed catchment Englehåsen

LIM 4: Main stream outlet of the limed catchment Englehåsen

REF: Stream outlet of reference catchment Spjøtåsen

Limed: 1 - unlimed; 2 - limed

PI-06 to PI-xx: Sampling stations for soil solution in the limed catchment Englehåsen

SP-28 to SP-xx: Sampling stations for soil solution in the limed catchment Spjøtåsen





S. No.	DATE	TIME	PT	DR	CR	MR	AR	AC	TCG	MR	R	C	SUP	AVGT	NOG	INT	W/M	TRF
U1P	13-Mar-84	12	43	11	11	11	11	11	11	11	11	11	11	11	11	11	11	11
U1Q	13-Mar-84	13	43	11	11	11	11	11	11	11	11	11	11	11	11	11	11	11
U1G	13-Mar-84	14	43	11	11	11	11	11	11	11	11	11	11	11	11	11	11	11
U1H	13-Mar-84	15	43	11	11	11	11	11	11	11	11	11	11	11	11	11	11	11
REF	13-Mar-84	16	43	11	11	11	11	11	11	11	11	11	11	11	11	11	11	11
U1J	28-Mar-84	17	43	11	11	11	11	11	11	11	11	11	11	11	11	11	11	11
U1K	28-Mar-84	18	43	11	11	11	11	11	11	11	11	11	11	11	11	11	11	11
U1L	28-Mar-84	19	43	11	11	11	11	11	11	11	11	11	11	11	11	11	11	11
REF	28-Mar-84	20	43	11	11	11	11	11	11	11	11	11	11	11	11	11	11	11
U1N	11-Apr-84	21	43	11	11	11	11	11	11	11	11	11	11	11	11	11	11	11
U1O	11-Apr-84	22	43	11	11	11	11	11	11	11	11	11	11	11	11	11	11	11
U1P	11-Apr-84	23	43	11	11	11	11	11	11	11	11	11	11	11	11	11	11	11
U1Q	11-Apr-84	24	43	11	11	11	11	11	11	11	11	11	11	11	11	11	11	11
U1R	11-Apr-84	25	43	11	11	11	11	11	11	11	11	11	11	11	11	11	11	11
U1S	11-Apr-84	26	43	11	11	11	11	11	11	11	11	11	11	11	11	11	11	11
U1T	11-Apr-84	27	43	11	11	11	11	11	11	11	11	11	11	11	11	11	11	11
REF	11-Apr-84	28	43	11	11	11	11	11	11	11	11	11	11	11	11	11	11	11
U1V	25-Apr-84	29	43	11	11	11	11	11	11	11	11	11	11	11	11	11	11	11
U1W	25-Apr-84	30	43	11	11	11	11	11	11	11	11	11	11	11	11	11	11	11
U1X	25-Apr-84	31	43	11	11	11	11	11	11	11	11	11	11	11	11	11	11	11
U1Y	25-Apr-84	32	43	11	11	11	11	11	11	11	11	11	11	11	11	11	11	11
REF	25-Apr-84	33	43	11	11	11	11	11	11	11	11	11	11	11	11	11	11	11
U1Z	2-May-84	34	43	11	11	11	11	11	11	11	11	11	11	11	11	11	11	11
U1A	2-May-84	35	43	11	11	11	11	11	11	11	11	11	11	11	11	11	11	11
REF	2-May-84	36	43	11	11	11	11	11	11	11	11	11	11	11	11	11	11	11
U1B	15-May-84	37	43	11	11	11	11	11	11	11	11	11	11	11	11	11	11	11
U1C	15-May-84	38	43	11	11	11	11	11	11	11	11	11	11	11	11	11	11	11
U1D	15-May-84	39	43	11	11	11	11	11	11	11	11	11	11	11	11	11	11	11
U1E	15-May-84	40	43	11	11	11	11	11	11	11	11	11	11	11	11	11	11	11
REF	15-May-84	41	43	11	11	11	11	11	11	11	11	11	11	11	11	11	11	11
U1F	30-May-84	42	43	11	11	11	11	11	11	11	11	11	11	11	11	11	11	11
U1G	30-May-84	43	43	11	11	11	11	11	11	11	11	11	11	11	11	11	11	11
U1H	30-May-84	44	43	11	11	11	11	11	11	11	11	11	11	11	11	11	11	11
REF	30-May-84	45	43	11	11	11	11	11	11	11	11	11	11	11	11	11	11	11





S. N°	DTC	J. W.	Q. R.	C. S.	M. B.	A. S.	A. C.	T. C.	L. S.	R.	C.	S. U. F.	R. E. S.	N. O. E. T.	T. M. V.	N. U. M. T. E. L.	T. E. P.
101	12-Jan-94	2	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10
102	12-Jan-94	2	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10
103	12-Jan-94	2	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10
104	12-Jan-94	2	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10
REF	12-Jan-94	1	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10
105	21-Jan-94	2	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10
106	21-Jan-94	2	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10
107	21-Jan-94	2	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10
108	21-Jan-94	2	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10
109	21-Jan-94	2	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10
110	21-Jan-94	2	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10
111	30-Jan-94	2	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10
112	30-Jan-94	2	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10
113	30-Jan-94	2	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10
114	30-Jan-94	2	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10
REF	30-Jan-94	1	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10
115	15-Feb-94	2	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10
116	15-Feb-94	2	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10
117	15-Feb-94	2	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10
118	15-Feb-94	2	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10
REF	15-Feb-94	1	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10
119	14-Mar-94	2	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10
120	14-Mar-94	2	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10
121	14-Mar-94	2	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10
122	14-Mar-94	2	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10
REF	14-Mar-94	1	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10
123	21-Mar-94	2	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10
124	21-Mar-94	2	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10
125	21-Mar-94	2	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10
126	21-Mar-94	2	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10
REF	21-Mar-94	1	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10
127	5-Apr-94	2	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10
128	5-Apr-94	2	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10
129	5-Apr-94	2	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10
130	5-Apr-94	2	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10
REF	5-Apr-94	1	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10

Site	DATE	L-rod	PH	Ca	MG	AL	AC	TOC	NH	NO <sub>3</sub>	SLUG	AMG	NO <sub>2N</sub>	TAN	KMn4	TSP
LM1	16-Aug-96	2	6.1	15	1.6	155	158	4	13	13	13	13	13	13	13	13
LM2	16-Aug-96	2	5.9	11	1.5	152	155	4	12	12	12	12	12	12	12	12
LM3	16-Aug-96	2	5.8	12	1.5	150	153	4	11	11	11	11	11	11	11	11
LM4	16-Aug-96	2	5.5	14	1.4	148	151	4	10	10	10	10	10	10	10	10
REF	16-Aug-96	1	4.8	17	1.1	138	141	4	7	7	7	7	7	7	7	7
LM1	25-Aug-96	2	5.9	12	1.5	152	155	4	11	11	11	11	11	11	11	11
LM2	25-Aug-96	2	5.7	9	1.3	148	151	4	9	9	9	9	9	9	9	9
LM3	25-Aug-96	2	5.8	11	1.4	150	153	4	10	10	10	10	10	10	10	10
LM4	25-Aug-96	2	5.8	11	1.4	150	153	4	10	10	10	10	10	10	10	10
REF	25-Aug-96	1	5.1	13	1.1	141	144	4	7	7	7	7	7	7	7	7
LM1	26-Aug-96	3	6.2	14	1.5	162	165	4	11	11	11	11	11	11	11	11
LM2	26-Aug-96	3	6.0	11	1.3	158	161	4	10	10	10	10	10	10	10	10
LM3	26-Aug-96	3	6.1	15	1.6	161	164	4	11	11	11	11	11	11	11	11
LM4	26-Aug-96	3	6.1	15	1.6	161	164	4	11	11	11	11	11	11	11	11
REF	26-Aug-96	1	5.1	13	1.1	141	144	4	7	7	7	7	7	7	7	7
LM1	30-Aug-96	3	6.2	15	1.6	162	165	4	11	11	11	11	11	11	11	11
LM2	30-Aug-96	3	6.0	12	1.4	160	163	4	10	10	10	10	10	10	10	10
LM3	30-Aug-96	3	6.2	15	1.6	162	165	4	11	11	11	11	11	11	11	11
LM4	30-Aug-96	3	6.1	14	1.5	161	164	4	10	10	10	10	10	10	10	10
REF	30-Aug-96	1	5.1	13	1.1	141	144	4	7	7	7	7	7	7	7	7
LM1	31-Aug-96	3	6.1	14	1.5	161	164	4	10	10	10	10	10	10	10	10
LM2	31-Aug-96	3	6.0	12	1.4	160	163	4	10	10	10	10	10	10	10	10
LM3	31-Aug-96	3	6.1	14	1.5	161	164	4	10	10	10	10	10	10	10	10
LM4	31-Aug-96	3	6.0	13	1.4	160	163	4	10	10	10	10	10	10	10	10
REF	31-Aug-96	1	4.9	13	1.1	140	143	4	7	7	7	7	7	7	7	7
LM1	29-Sep-96	3	6.1	14	1.5	161	164	4	10	10	10	10	10	10	10	10
LM2	29-Sep-96	3	6.0	12	1.4	160	163	4	10	10	10	10	10	10	10	10
LM3	29-Sep-96	3	6.1	14	1.5	161	164	4	10	10	10	10	10	10	10	10
LM4	29-Sep-96	3	6.0	13	1.4	160	163	4	10	10	10	10	10	10	10	10
REF	29-Sep-96	1	5.1	13	1.1	141	144	4	7	7	7	7	7	7	7	7



S ac	DATE	U-99	14	Da	Mo	Av	Ac	700	Na	K	Cl	SLP	AMC	MOON	TUM	RASH	TCP
REF	12-01-96	2	54	1	1	188	28	5	18	3	29	11	28	45	260	45	4
REF	12-02-96	2	53	13	1	147	20	6	18	3	30	11	28	49	260	45	4
REF	12-03-96	1	48	13	16	133	18	6	18	3	30	11	28	49	260	45	4
REF	12-04-96	2	55	16	16	133	18	6	18	3	30	11	28	49	260	45	4
REF	12-05-96	2	51	16	16	133	18	6	18	3	30	11	28	49	260	45	4
REF	12-06-96	1	4	16	17	132	18	6	18	3	30	11	28	49	260	45	4
REF	12-07-96	2	53	17	15	131	16	6	18	3	30	11	28	49	260	45	4
REF	12-08-96	2	53	15	15	131	16	6	18	3	30	11	28	49	260	45	4
REF	12-09-96	1	49	15	15	131	16	6	18	3	30	11	28	49	260	45	4
REF	12-10-96	2	53	15	15	131	16	6	18	3	30	11	28	49	260	45	4
REF	12-11-96	2	53	15	15	131	16	6	18	3	30	11	28	49	260	45	4
REF	12-12-96	2	53	15	15	131	16	6	18	3	30	11	28	49	260	45	4
REF	12-13-96	1	49	15	15	131	16	6	18	3	30	11	28	49	260	45	4
REF	12-14-96	2	53	15	15	131	16	6	18	3	30	11	28	49	260	45	4
REF	12-15-96	2	53	15	15	131	16	6	18	3	30	11	28	49	260	45	4
REF	12-16-96	2	53	15	15	131	16	6	18	3	30	11	28	49	260	45	4
REF	12-17-96	2	53	15	15	131	16	6	18	3	30	11	28	49	260	45	4
REF	12-18-96	2	53	15	15	131	16	6	18	3	30	11	28	49	260	45	4
REF	12-19-96	2	53	15	15	131	16	6	18	3	30	11	28	49	260	45	4
REF	12-20-96	2	53	15	15	131	16	6	18	3	30	11	28	49	260	45	4
REF	12-21-96	2	53	15	15	131	16	6	18	3	30	11	28	49	260	45	4
REF	12-22-96	2	53	15	15	131	16	6	18	3	30	11	28	49	260	45	4
REF	12-23-96	2	53	15	15	131	16	6	18	3	30	11	28	49	260	45	4
REF	12-24-96	2	53	15	15	131	16	6	18	3	30	11	28	49	260	45	4
REF	12-25-96	2	53	15	15	131	16	6	18	3	30	11	28	49	260	45	4
REF	12-26-96	2	53	15	15	131	16	6	18	3	30	11	28	49	260	45	4
REF	12-27-96	2	53	15	15	131	16	6	18	3	30	11	28	49	260	45	4
REF	12-28-96	2	53	15	15	131	16	6	18	3	30	11	28	49	260	45	4
REF	12-29-96	2	53	15	15	131	16	6	18	3	30	11	28	49	260	45	4
REF	12-30-96	2	53	15	15	131	16	6	18	3	30	11	28	49	260	45	4
REF	12-31-96	2	53	15	15	131	16	6	18	3	30	11	28	49	260	45	4



St. No	DATE	JRS	HR	Ca	Mg	Al	Si	TOC	Fe	K	S	Sulf	PHI	MOH	TSI	N+H	TPP
UW4	2-Jan-98	3	11	14	19	150	16	9	18	14	29	59	11	14	126	12	1
UW4	2-Jan-98	3	11	14	19	150	16	9	18	14	29	59	11	14	126	12	1
PEF	2-Jan-98	3	11	14	19	150	16	9	18	14	29	59	11	14	126	12	1
UW4	2-Jan-98	3	11	14	19	150	16	9	18	14	29	59	11	14	126	12	1
UW4	2-Jan-98	3	11	14	19	150	16	9	18	14	29	59	11	14	126	12	1
PEF	2-Jan-98	3	11	14	19	150	16	9	18	14	29	59	11	14	126	12	1
UW4	4-Dec-98	3	11	14	19	150	16	9	18	14	29	59	11	14	126	12	1
UW4	4-Dec-98	3	11	14	19	150	16	9	18	14	29	59	11	14	126	12	1
PEF	4-Dec-98	3	11	14	19	150	16	9	18	14	29	59	11	14	126	12	1
UW4	3-Dec-98	3	11	14	19	150	16	9	18	14	29	59	11	14	126	12	1
UW4	3-Dec-98	3	11	14	19	150	16	9	18	14	29	59	11	14	126	12	1
PEF	3-Dec-98	3	11	14	19	150	16	9	18	14	29	59	11	14	126	12	1
UW4	8-Dec-98	3	11	14	19	150	16	9	18	14	29	59	11	14	126	12	1
UW4	8-Dec-98	3	11	14	19	150	16	9	18	14	29	59	11	14	126	12	1
PEF	8-Dec-98	3	11	14	19	150	16	9	18	14	29	59	11	14	126	12	1
UW4	3-Jan-99	3	11	14	19	150	16	9	18	14	29	59	11	14	126	12	1
UW4	3-Jan-99	3	11	14	19	150	16	9	18	14	29	59	11	14	126	12	1
PEF	3-Jan-99	3	11	14	19	150	16	9	18	14	29	59	11	14	126	12	1
UW4	5-Jan-99	3	11	14	19	150	16	9	18	14	29	59	11	14	126	12	1
UW4	5-Jan-99	3	11	14	19	150	16	9	18	14	29	59	11	14	126	12	1
PEF	5-Jan-99	3	11	14	19	150	16	9	18	14	29	59	11	14	126	12	1
UW4	5-May-99	3	11	14	19	150	16	9	18	14	29	59	11	14	126	12	1
UW4	5-May-99	3	11	14	19	150	16	9	18	14	29	59	11	14	126	12	1
PEF	5-May-99	3	11	14	19	150	16	9	18	14	29	59	11	14	126	12	1
UW4	23-Jul-99	3	11	14	19	150	16	9	18	14	29	59	11	14	126	12	1
UW4	23-Jul-99	3	11	14	19	150	16	9	18	14	29	59	11	14	126	12	1
PEF	23-Jul-99	3	11	14	19	150	16	9	18	14	29	59	11	14	126	12	1
UW4	2-Jul-99	3	11	14	19	150	16	9	18	14	29	59	11	14	126	12	1
UW4	2-Jul-99	3	11	14	19	150	16	9	18	14	29	59	11	14	126	12	1
PEF	2-Jul-99	3	11	14	19	150	16	9	18	14	29	59	11	14	126	12	1
UW4	1-Aug-99	3	11	14	19	150	16	9	18	14	29	59	11	14	126	12	1
UW4	1-Aug-99	3	11	14	19	150	16	9	18	14	29	59	11	14	126	12	1
PEF	1-Aug-99	3	11	14	19	150	16	9	18	14	29	59	11	14	126	12	1
UW4	4-Sep-99	3	11	14	19	150	16	9	18	14	29	59	11	14	126	12	1
UW4	4-Sep-99	3	11	14	19	150	16	9	18	14	29	59	11	14	126	12	1
PEF	4-Sep-99	3	11	14	19	150	16	9	18	14	29	59	11	14	126	12	1
UW4	13-Sep-99	3	11	14	19	150	16	9	18	14	29	59	11	14	126	12	1
UW4	13-Sep-99	3	11	14	19	150	16	9	18	14	29	59	11	14	126	12	1
PEF	13-Sep-99	3	11	14	19	150	16	9	18	14	29	59	11	14	126	12	1
UW4	14-Oct-99	3	11	14	19	150	16	9	18	14	29	59	11	14	126	12	1
UW4	14-Oct-99	3	11	14	19	150	16	9	18	14	29	59	11	14	126	12	1
PEF	14-Oct-99	3	11	14	19	150	16	9	18	14	29	59	11	14	126	12	1
UW4	4-Nov-99	3	11	14	19	150	16	9	18	14	29	59	11	14	126	12	1
UW4	4-Nov-99	3	11	14	19	150	16	9	18	14	29	59	11	14	126	12	1
PEF	4-Nov-99	3	11	14	19	150	16	9	18	14	29	59	11	14	126	12	1
UW4	12-Nov-99	3	11	14	19	150	16	9	18	14	29	59	11	14	126	12	1
UW4	12-Nov-99	3	11	14	19	150	16	9	18	14	29	59	11	14	126	12	1
PEF	12-Nov-99	3	11	14	19	150	16	9	18	14	29	59	11	14	126	12	1
UW4	15-Nov-99	3	11	14	19	150	16	9	18	14	29	59	11	14	126	12	1
UW4	15-Nov-99	3	11	14	19	150	16	9	18	14	29	59	11	14	126	12	1
PEF	15-Nov-99	3	11	14	19	150	16	9	18	14	29	59	11	14	126	12	1



Ln	W	Tab	Days	Links	HT log ft.	Latc MSF	MSF mg/L	SO <sub>4</sub> mg/L	S mg/L	WHAM mg/L	NO <sub>3</sub> -N mg/L	DO <sub>5</sub> mg/L	Ca mg/L	Mg mg/L	Fe mg/L	TDS mg/L
1	00	10-16-98	1	1	4.26	2.47	1.4	1.4	1	1	1	1	1	1	1	1
2	00	20-10-98	2	2	4.22	2.27	1.4	1.4	1	1	1	1	1	1	1	1
3	00	30-10-98	3	3	4.14	2.22	1.4	1.4	1	1	1	1	1	1	1	1
4	00	30-10-98	4	4	4.25	2.24	1.4	1.4	1	1	1	1	1	1	1	1
5	00	30-10-98	5	5	4.26	2.27	1.4	1.4	1	1	1	1	1	1	1	1
6	00	30-10-98	6	6	4.26	2.27	1.4	1.4	1	1	1	1	1	1	1	1
7	00	30-10-98	7	7	4.1	2.1	1.4	1.4	1	1	1	1	1	1	1	1
8	00	30-10-98	8	8	4.26	2.27	1.4	1.4	1	1	1	1	1	1	1	1
9	00	30-10-98	9	9	4.26	2.27	1.4	1.4	1	1	1	1	1	1	1	1
10	00	30-10-98	10	10	4.26	2.27	1.4	1.4	1	1	1	1	1	1	1	1
11	00	30-10-98	11	11	4.26	2.27	1.4	1.4	1	1	1	1	1	1	1	1
12	00	30-10-98	12	12	4.26	2.27	1.4	1.4	1	1	1	1	1	1	1	1
13	00	30-10-98	13	13	4.26	2.27	1.4	1.4	1	1	1	1	1	1	1	1
14	00	30-10-98	14	14	4.26	2.27	1.4	1.4	1	1	1	1	1	1	1	1
15	00	30-10-98	15	15	4.26	2.27	1.4	1.4	1	1	1	1	1	1	1	1
16	00	30-10-98	16	16	4.26	2.27	1.4	1.4	1	1	1	1	1	1	1	1
17	00	30-10-98	17	17	4.26	2.27	1.4	1.4	1	1	1	1	1	1	1	1
18	00	30-10-98	18	18	4.26	2.27	1.4	1.4	1	1	1	1	1	1	1	1
19	00	30-10-98	19	19	4.26	2.27	1.4	1.4	1	1	1	1	1	1	1	1
20	00	30-10-98	20	20	4.26	2.27	1.4	1.4	1	1	1	1	1	1	1	1
21	00	30-10-98	21	21	4.26	2.27	1.4	1.4	1	1	1	1	1	1	1	1
22	00	30-10-98	22	22	4.26	2.27	1.4	1.4	1	1	1	1	1	1	1	1
23	00	30-10-98	23	23	4.26	2.27	1.4	1.4	1	1	1	1	1	1	1	1
24	00	30-10-98	24	24	4.26	2.27	1.4	1.4	1	1	1	1	1	1	1	1
25	00	30-10-98	25	25	4.26	2.27	1.4	1.4	1	1	1	1	1	1	1	1
26	00	30-10-98	26	26	4.26	2.27	1.4	1.4	1	1	1	1	1	1	1	1
27	00	30-10-98	27	27	4.26	2.27	1.4	1.4	1	1	1	1	1	1	1	1
28	00	30-10-98	28	28	4.26	2.27	1.4	1.4	1	1	1	1	1	1	1	1
29	00	30-10-98	29	29	4.26	2.27	1.4	1.4	1	1	1	1	1	1	1	1
30	00	30-10-98	30	30	4.26	2.27	1.4	1.4	1	1	1	1	1	1	1	1



Loc	rc	Date	Dist	Unit	PH	Land	AUR	D	SOL	S	Area	MOON	Test	CT	By	TR	X	AR	Mo	DC
1	1	1-1-68	15	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10
2	2	1-15-68	15	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10
3	3	1-20-68	15	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10
4	4	1-25-68	15	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10
5	5	1-30-68	15	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10
6	6	2-5-68	15	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10
7	7	2-10-68	15	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10
8	8	2-15-68	15	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10
9	9	2-20-68	15	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10
10	10	2-25-68	15	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10
11	11	2-30-68	15	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10
12	12	3-5-68	15	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10
13	13	3-10-68	15	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10
14	14	3-15-68	15	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10
15	15	3-20-68	15	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10
16	16	3-25-68	15	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10
17	17	3-30-68	15	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10
18	18	4-5-68	15	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10
19	19	4-10-68	15	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10
20	20	4-15-68	15	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10
21	21	4-20-68	15	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10
22	22	4-25-68	15	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10
23	23	4-30-68	15	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10
24	24	5-5-68	15	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10
25	25	5-10-68	15	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10
26	26	5-15-68	15	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10
27	27	5-20-68	15	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10
28	28	5-25-68	15	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10
29	29	5-30-68	15	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10
30	30	6-5-68	15	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10
31	31	6-10-68	15	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10
32	32	6-15-68	15	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10
33	33	6-20-68	15	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10
34	34	6-25-68	15	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10
35	35	6-30-68	15	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10
36	36	7-5-68	15	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10
37	37	7-10-68	15	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10
38	38	7-15-68	15	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10
39	39	7-20-68	15	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10
40	40	7-25-68	15	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10
41	41	7-30-68	15	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10
42	42	8-5-68	15	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10
43	43	8-10-68	15	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10
44	44	8-15-68	15	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10
45	45	8-20-68	15	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10
46	46	8-25-68	15	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10
47	47	8-30-68	15	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10
48	48	9-5-68	15	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10
49	49	9-10-68	15	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10
50	50	9-15-68	15	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10

Lot	no	date	Dept	Line#	Qty	Unit	Price	Amount	Disc	Net	Tax	Total
001	01	01/01/01	01	01	100	EA	1.00	100.00		100.00		100.00
002	02	01/01/01	01	02	200	EA	2.00	400.00		400.00		400.00
003	03	01/01/01	01	03	300	EA	3.00	900.00		900.00		900.00
004	04	01/01/01	01	04	400	EA	4.00	1600.00		1600.00		1600.00
005	05	01/01/01	01	05	500	EA	5.00	2500.00		2500.00		2500.00
006	06	01/01/01	01	06	600	EA	6.00	3600.00		3600.00		3600.00
007	07	01/01/01	01	07	700	EA	7.00	4900.00		4900.00		4900.00
008	08	01/01/01	01	08	800	EA	8.00	6400.00		6400.00		6400.00
009	09	01/01/01	01	09	900	EA	9.00	8100.00		8100.00		8100.00
010	10	01/01/01	01	10	1000	EA	10.00	10000.00		10000.00		10000.00









Loc	Ac	Dep	Depth	Line	W	Land	LJK	G	SOx	S	MnO <sub>2</sub>	Fe-4	Cr	Mn	Ag	C	M	Al	TC	
					mg/l	mg/l	mg/l	mg/l	mg/l	mg/l	mg/l	mg/l	mg/l	mg/l	mg/l	mg/l	mg/l	mg/l	mg/l	mg/l
23	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10
24	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10
25	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10
26	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10
27	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10
28	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10
29	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10
30	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10
31	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10
32	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10
33	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10
34	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10
35	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10
36	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10
37	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10
38	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10
39	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10
40	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10
41	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10
42	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10
43	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10
44	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10
45	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10
46	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10
47	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10
48	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10
49	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10
50	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10	10











## Appendix B. Hydrology

Abbreviation and references for sampling stations:

LIM 1: Stream outlet of catchment 1 in the lined catchment Fuglehåsen

LIM 4: Main stream outlet of the lined catchment Fuglehåsen

REF: Stream outlet of reference catchment Spjotåsen

DATE	LIM 1 mg/l/pond	LIM 4 mg/l/pond	REF mg/l/pond	DATE	LIM 1 mg/l/pond	LIM 4 mg/l/pond	REF mg/l/pond
1 Mar 94				12 Jun 96	600	2070	307
12 Mar 94	41900	21455	35210	18 Apr 96	25200	60120	22005
20 Mar 94	15050	20080	14520	27 Apr 96	10070	42030	13410
31 Apr 94	30137	150127	60315	28 Apr 96	16000	29703	11203
01 Apr 94	27255	47200	25250	7 May 96	0707	20570	7030
25 Apr 94	28121	48820	20500	20 May 96	10797	24244	8553
2 May 94	30072	62329	28001	29 May 96	7474	10413	6707
9 May 94	14000	24480	12150	29 Jun 96	2010	9437	3157
18 May 94	5002	1725	2450	7 Sep 96	7400	10105	5007
25 Aug 94	12747	30574	3079	30 Sep 96	21941	51525	21210
31 Aug 94	22017	57605	10000	16 Oct 96	11030	28623	10000
11 Sep 94	25015	50405	22700	20 Oct 96	20700	50004	20100
19 Sep 94	30121	58337	30000	7 Nov 96	17440	27884	15207
29 Sep 94	3100	5000	372	21 Nov 96	8094	10400	7407
12 Oct 94	1700	2420	1000	4 Dec 96	3001	20174	6213
26 Oct 94	3774	17110	15001	8 Dec 96	10000	20127	17010
31 Oct 94	25504		7212	21 Dec 96	7070	10010	2002
3 Nov 94	11000	00000	0000	3 Mar 97		104800	41015
14 Nov 94	5000	10042	4001	5 May 97		100724	30067
21 Nov 94	0625	18408	7014	21 Jun 97		32745	0000
6 Dec 94	6146	11000	4007	2 Jul 97		17000	5711
12 Dec 94	20200	38500	14980	7 Aug 97		5000	2007
21 Dec 94	10412	27807	11001	4 Sep 97		0250	0000
31 Dec 94	0000	10172	0010	18 Sep 97		10720	0000
3 Jan 95	1102	2000	1000	14 Oct 97		35000	10427
15 Feb 95	50000	110000	00000	4 Nov 97		10000	2000
14 Mar 95	31000	60007	20700	10 Nov 97		20700	0600
22 Mar 95	25101	40771	20000	19 Nov 97		30004	10000
5 Apr 95	40000	20001	30000	21 Dec 97		100000	01217
16 Apr 95	30700	00750	20000				
25 Apr 95	30700	50070	10001				
8 May 95	20000	40000	10000				
26 May 95	3001	10020	7704				
8 Jun 95	11000	10700	7004				
13 Jun 95	10004	24107	10057				
27 Jun 95	22007	31000	10172				
7 Sep 95	1000	1000	1000				
14 Sep 95	17015	20000	11000				
18 Sep 95	27045	50004	22000				
4 Oct 95	10000	20700	10000				
20 Oct 95	10100	30074	10771				
20 Nov 95	4770	11410	3707				
4 Dec 95	2775	0001	1000				
10 Dec 95	4000	10707	2007				
31 Dec 95	000	2007	075				



