

Accepted Manuscript

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This is an Accepted Manuscript of the following article:

Mengzhu Wang, Yindong Tong, Cen Chen, Xianhua Liu, Yiren Lu, Wei Zhang, Wei He, Xuejun Wang, Shen Zhao, Yan Lin. Ecological risk assessment to marine organisms induced by heavy metals in China's coastal waters. *Marine Pollution Bulletin*. Volume 126, 2018, Pages 349-356, ISSN 0025-326X.

The article has been published in final form by Elsevier at

<http://dx.doi.org/10.1016/j.marpolbul.2017.11.019>

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1 **Ecological risk assessment to marine organisms induced by heavy metals in**  
2 **China's coastal waters**

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## 20 **Abstract**

21 China's coastal environment has been heavily affected by the loading of terrestrial  
22 pollutants in recent decades, and quantitative risk assessment is urgently needed to  
23 assess the ecological risks of China's coastal environment. We assessed the ecological  
24 risks induced by five heavy metals (including Cu, Zn, Pb, Hg and As) in China's  
25 coastal waters for three groups of marine organisms (including crustacean, fish and  
26 mollusc) based on data obtained from a nationwide unified coastal environment  
27 monitoring program consisting of 301 sampling sites. The results show that higher  
28 heavy metal concentrations occurred more frequently in the Bohai Sea and in the  
29 estuaries of major sea-going rivers. The ecological risks decreased in the following  
30 order: Bohai Sea > Yellow Sea > South China Sea > East China Sea. There was  
31 generally low ecological risk, but certain hotspots existed near Tianjin and Jinzhou,  
32 which had relatively high ecological risks caused by Cu and Zn.

## 33 **Keywords**

34 Ecological risk assessment; heavy metals; marine organisms; China's coastal water;  
35 species sensitivity distributions

## 36 **1. Introduction**

37 In recent decades, China's coastal water environment has faced serious  
38 contamination due to continuous pollutant loadings from the terrestrial system (Du et  
39 al., 2013; Liu et al., 2013; State Oceanic Administration, China, 2015, 2016).  
40 Generally, these pollutant loadings can be divided into two major categories: nutrient  
41 elements (i.e., nitrogen and phosphorus) and toxic pollutants (i.e., heavy metals and  
42 organic pollutants) (Li et al., 2014; Ministry of Environmental Protection, China,  
43 2015; Selman et al., 2015; Tong et al., 2015). Unlike nutrients, heavy metals in water  
44 are especially of concern for marine organisms and seafood consumers due to their  
45 toxicity, bio-concentration and non-degradation (Morillo et al., 2002; Loska and  
46 Wiechula, 2003; Xing et al., 2013). In 2015, the riverine discharges of heavy metals  
47 into seas were reported to be ~21,000 tons, including 3,318 tons of Cu, 16,243 tons of

48 Zn, 858 tons of Pb, 49 tons of Hg and 3,188 tons of As (State Oceanic Administration,  
49 China, 2015). In addition to the riverine discharges, the contribution from direct  
50 discharges by coastal cities is also significant. For instance, up to ~18 tons of Pb was  
51 directly discharged into China's seas in 2015 (Ministry of Environmental Protection,  
52 China, 2015).

53 These heavy metals in coastal waters could have seriously negative impacts on  
54 living organisms. Kim et al. (2003) reported that in the coastal waters of Southeast  
55 Korea, the decreased genetic diversity of organisms occurred more frequently in the  
56 regions polluted by heavy metals. Johnston et al. (2009) estimated that an ~30%-50%  
57 reduction of species richness was caused by marine contaminants, especially heavy  
58 metals. In addition to heavy metals with high acute toxicity (i.e., Pb and Hg),  
59 less-toxic heavy metals in water could pose even more serious ecological risk  
60 concerns for aquatic organisms (Fu et al., 2017). Fu et al. (2016, 2017) reported that  
61 the ecological risks to freshwater organisms caused by Cu and Zn were much higher  
62 than those of other heavy metals. The greatest probabilities of ecological risks to  
63 aquatic species were ~100% for Cu and ~50% for Zn, whereas the risks caused by  
64 other toxic heavy metals (<0.1%) could almost be neglected (Fu et al., 2016, 2017).

65 Ecological risk assessment (ERA) is used to estimate the possibility of the potential  
66 ecological effects if aquatic organisms are exposed to one or several environmental  
67 stressors (Solomon et al., 1996; Steen et al., 1999). Usually, the ecological risk is  
68 expressed as the percentage of the affected organisms relative to all of the organisms  
69 belonging to the same or similar groups of species. The US Environmental Protection  
70 Agency established a complete database (ECOTOX database) based on the EC<sub>50</sub>  
71 (median effect concentration) and LC<sub>50</sub> (median lethal concentration) values of  
72 various aquatic organisms in fresh or sea water if exposed to different concentrations  
73 and species of heavy metals. To apply the toxicological data to develop favourable  
74 and reasonable environmental standards for the entire ecosystem or to assess the  
75 hazards of pollutants at the ecological level, interspecies correlation estimation (ICE)

76 models and a species sensitivity distributions (SSDs) method has been proposed and  
77 widely applied during the past decades (Traas et al., 2002; Wheeler et al., 2002;  
78 Azevedo et al., 2015; Zheng et al., 2015). The ICE mode and SSDs were used to  
79 assess ecological risks when they were first introduced in the 1970s (Kooijman, 1987).  
80 These were used to build statistical distribution models on the basis of the varying  
81 sensitivities of species to different pollutants and to provide assessments of a certain  
82 region's ecological risks (Wheeler et al., 2002). Lenwood et al. (1998) applied SSDs  
83 to calculate the concentrations of Cu and Ca that could protect 90% of the species in  
84 the freshwater and saltwater systems of the Chesapeake Bay of the United States. The  
85 90% protected concentrations of Cu were 8.3 mg/L in freshwater and 6.3 mg/L in  
86 saltwater and those of Ca were 5.1 mg/L and 31.4 mg/L, respectively. The derived  
87 value could be applied as a reference value for the development of environmental  
88 quality standards. Van Sprang et al. (2004) estimated that the no-effect concentration  
89 for aquatic organisms based on dissolved Zn was 34.2 mg/L in Dutch surface water  
90 and estimated that there were only very limited ecological risks. Mu et al. (2014)  
91 integrated quantitative ion character-activity relationships and SSDs to derive the  
92 saltwater quality criteria for 34 different metals or metalloids. Even when considering  
93 the uncertainty due to the application of different correlation models and variances in  
94 the toxicity data between different species (He et al., 2017), ICE and SSDs are still  
95 important tools to assess the ERA caused by heavy metals to organisms.

96 Although marine organisms are facing increasing pressure from terrestrial pollutant  
97 loadings to China's coastal waters, the ecological risks to these organisms have never  
98 been previously assessed by a unified dataset. In this study, we compared the heavy  
99 metal concentrations (including Cu, Zn, Pb, Hg and As) in China's four adjacent seas  
100 (including the Bohai Sea, Yellow Sea, East China Sea and South China Sea) based on  
101 a nationwide dataset obtained from 301 sampling sites in 2015. To establish the  
102 linkage between aquatic heavy metal concentrations and the ecological risks of  
103 marine organisms, the ICE and SSDs methods based on the US EPA ECOTOX

104 database were applied to estimate the potential ecological risks for crustacean, fish  
105 and mollusc in coastal waters. The coastal regions near the estuaries of major  
106 sea-going rivers were addressed in this study. This study attempts to describe the  
107 current status of heavy metal contamination in China's coastal waters and to pinpoint  
108 the major heavy metals, regions and sensitive aquatic species with high ecological  
109 risks, which is important for future policy-making regarding the control of heavy  
110 metal pollution and the protection of coastal environments.

## 111 **2. Methods and materials**

### 112 **2.1 Concentrations of heavy metals in China's coastal waters**

113 The dataset applied in this study was derived from the coastal water quality  
114 monitoring program, which was conducted by the Ministry of Environmental  
115 Protection, China, in 2015 (see Table S1). The dataset includes the concentrations of  
116 five major inorganic heavy metals, including Cu, Zn, Pb, Hg and As, from a total of  
117 301 monitoring sites distributed in four seas (35 sites in the Bohai Sea, 68 sites in the  
118 Yellow Sea, 95 sites in the East China Sea and 103 sites in the South China Sea as  
119 shown in Figure S1). The water sampling locations were usually set ~30-50 km away  
120 from coastlines to avoid direct interferences from human activities. All of the field  
121 samplings were conducted during the period between August and October in 2015 and  
122 based on the guidelines of "Specifications on Spot Location of Monitoring Sites  
123 Related to Coastal Area Environment (HJ730-2014)". Approximately 500 mL of  
124 filtered mixed water samples (~1 meter below sea surface, ~10 meters below the sea  
125 surface and ~2 meters above the seabed, filtered by 0.45  $\mu\text{m}$  Millipore filter) were  
126 collected.  $\text{HNO}_3$  (guaranteed reagent, GR) was added to make the pH of the water  
127 lower than 2 before sample analysis. The detection limits for Cu, Zn, Pb, Hg and As  
128 were 0.100, 1.550, 0.150, 0.001 and 0.200  $\mu\text{g/L}$ , respectively (HJ730-2014). During  
129 data analysis, for sampling sites with measured values that were lower than the  
130 detection limit, their values were assumed to be half of the detection limits.

### 131 **2.2 Ecological risk assessment**

132 As recommended by previous studies (Newman et al., 1995, 2000; Zheng et al.,  
133 2015, 2016), we used the EC<sub>50</sub> or LC<sub>50</sub> values from laboratory tests as an effective  
134 indicator of marine organisms exposed to heavy metals. The EC<sub>50</sub> or LC<sub>50</sub> data of the  
135 selected heavy metals were obtained from the US EPA ECOTOX database (US EPA,  
136 2017). In this study, we selected crustacean, fish and mollusc as the target marine  
137 organisms, which are the most common marine organisms in China's coastal waters;  
138 they also have good quantity and available ecotoxicity data. The following principles  
139 were applied during the selection of ecotoxicity data (Durda and Preziosi, 2000; Liu et  
140 al., 2011; Zheng et al., 2015): I. The ecotoxicity data for the organisms were obtained  
141 in a sea water environment. II. The forms of heavy metals exposed to organisms in the  
142 laboratory test were inorganic, and organic forms (i.e., methyl mercury) were  
143 excluded in the data screening due to a lack of monitoring data. III. Eight or more  
144 previously reported toxicity data points were required to construct ICE models and  
145 SSD curves for one group of species. For instance, the SSD curve for As to mollusc is  
146 not applicable in this study due to the limited reported ecotoxicity data. The  
147 ecotoxicity data for different groups of marine organisms are provided in Table S2.  
148 The specific process of calculating SSDs and its uncertainty was performed by the  
149 SSD generator (US EPA, 2016).

150 By using the data of organisms belonging to similar groups of species from  
151 laboratory tests, SSD curves could assist in the interpretation of site data for stressor  
152 identification and risk assessment by relating them to the proportions of species  
153 expected to be affected at the prescribed concentrations, which is used to represent the  
154 ecological risks to marine organisms. SSDs could model the variations in the  
155 sensitivity of different groups of species to a certain stressor (i.e., heavy metals). For  
156 instance, HC<sub>5</sub> (5% Hazardous Concentration) represents the exposure concentration,  
157 which causes 5% of the target organisms (i.e., fish and mollusc) to experience  
158 negative impacts, which are predefined by different endpoints, including reduced  
159 mobility and fertility. Several interspecies correlation models have been applied to

160 describe the distribution of toxicological data when building SSD curves. Different  
161 selections of the fitting models (i.e., log-normal distribution, log-logistic model and  
162 bootstrap regression) could cause variances in the estimation of ecological risks to  
163 marine organisms (Newman et al., 2000; Durda & Preziosi, 2000; Wheeler et al.,  
164 2002). Among these models, the log-normal model was recommended by US EPA for  
165 the assessment of the ecological risks of heavy metals (US EPA, 2016). The SSD  
166 curves for different marine organisms and heavy metals are provided in Figure S2.  
167 Based on the SSD curves and monitoring data in the coastal waters, we calculated the  
168 ecological risks and evaluated the proportions of marine organisms expected to be  
169 affected at prescribed concentrations for crustacean, fish and mollusc, respectively.  
170 To describe the general risk to all organisms (regardless of species), three groups of  
171 marine organisms (crustacean, fish and mollusc) were also mixed to construct the  
172 SSD curves and estimate the general risks induced by a single heavy metal to all  
173 organisms. In addition to the risks induced by a single heavy metal, the following  
174 equation was applied to estimate the effects of multiple substances, i.e., the potential  
175 affected fraction (msPAF) of marine organisms by the five heavy metals (Kong et al.,  
176 2011; Hu et al., 2012; Jiang et al., 2012; He et al., 2014):

$$177 \quad \text{msPAF} = 1 - (1 - P_{\text{Cu}}) \times (1 - P_{\text{Zn}}) \times (1 - P_{\text{Pb}}) \times (1 - P_{\text{Hg}}) \times (1 - P_{\text{As}})$$

178 Where  $P_{\text{Cu}}$ ,  $P_{\text{Zn}}$ ,  $P_{\text{Pb}}$ ,  $P_{\text{Hg}}$  and  $P_{\text{As}}$  are the ecological risks for organisms under the  
179 exposure of Cu, Zn, Pb, Hg and As in the seas, respectively.

### 180 **3. Results**

#### 181 **3.1 Concentrations of heavy metals in China's coastal waters**

182 Figure 1 shows the concentrations of selected heavy metals in 301 sampling sites  
183 distributed in China's four seas in 2015. Unlike serious nutrient pollution (i.e.,  
184 nitrogen and phosphorus) (Li et al., 2014; Selman et al., 2015), heavy metal  
185 concentrations in China's coastal waters were relatively low for the majority of the  
186 sampling sites, and higher measured values usually occurred in the Bohai Sea. For Cu,  
187 ~95% of the sampling sites had concentrations of lower than 5  $\mu\text{g/L}$  (Figure 1), which



188 was the Grade I limit by China's sea water quality standard (see Table S3). The  
189 average measured Cu concentrations in the Bohai Sea, Yellow Sea, East China Sea  
190 and South China Sea were  $3.7\pm 2.5$ ,  $1.6\pm 1.5$ ,  $0.8\pm 0.5$  and  $1.2\pm 1.0$   $\mu\text{g/L}$ , respectively.  
191 For Zn, most of the sampling sites had concentrations of lower than 20  $\mu\text{g/L}$ , except  
192 for some sites in the Bohai Sea, where Zn concentrations ranged from 5.6 to 49.6  
193  $\mu\text{g/L}$  and ~60% of the monitored values exceeded 20  $\mu\text{g/L}$ . The lowest Zn  
194 concentrations were observed in the East China Sea, with a range of 1.6-15.8  $\mu\text{g/L}$ .  
195 For Pb, ~50% of the sites in the Bohai Sea had concentrations of higher than 1  $\mu\text{g/L}$ ,  
196 which is the Grade I limit based on China's sea water standard (Table S3). For Hg, the  
197 difference in the spatial distributions among the four seas was not significant, and  
198 ~98% of monitored values were lower than 0.05  $\mu\text{g/L}$ , which was the Grade I limit  
199 based on China's sea water standard (Tong et al., 2017). The As concentrations in the  
200 four seas were far below the Grade I limit based on China's sea water standard (i.e.,  
201  $0.8\pm 1.1$   $\mu\text{g/L}$  in the Bohai Sea,  $1.1\pm 0.9$   $\mu\text{g/L}$  in the Yellow Sea,  $1.8\pm 0.7$   $\mu\text{g/L}$  in the  
202 East China Sea, and  $1.0\pm 0.9$   $\mu\text{g/L}$  in the South China Sea).

203 Spatial variances in heavy metal concentrations along the coastal regions were  
204 observed (Figure 1), and the higher concentrations usually occurred at the estuaries of  
205 the major sea-going rivers (i.e., the Yangtze and Pearl Rivers). In coastal waters, the  
206 concentrations of heavy metals largely depended on terrestrial pollutant loadings,  
207 water volume and the water exchange rate with the open ocean (Lin et al., 2016;  
208 Sharples et al., 2016). It has been reported that the residence time of riverine water on  
209 the continental shelf around China could last from months to years before the water is  
210 completely exchanged with the open ocean (Lin et al., 2016; Sharples et al., 2016),  
211 and limited exchange with the open ocean could prolong the residence time of heavy  
212 metals in China's coastal waters. The main reason for the poor exchange capacity of  
213 the Bohai Sea is its semi-enclosed bay. Also, within the same sea, relatively higher  
214 concentrations usually occurred at the estuaries of the major sea-going rivers. For  
215 instance, the Cu and As concentrations at the estuaries of the Pearl River were higher

216 than those at the other monitoring sites in the South China Sea (Figure 1). This fact  
217 could probably be explained by huge riverine pollutant loadings to the sea. For  
218 example, sea-going rivers were believed to be important Hg contributors to coastal  
219 waters (Liu et al., 2016). Among all of the sources (e.g., riverine discharges, coastal  
220 erosion), riverine Hg discharge was dominant and occupied ~70% of the terrestrial Hg  
221 loadings (Liu et al., 2016). In 2015, riverine discharges of heavy metals into the seas  
222 were reported to be ~21,000 tons (3,318 tons of Cu, 16,243 tons of Zn, 858 tons of Pb,  
223 49 tons of Hg and 3,188 tons of As) (State Oceanic Administration, China, 2015).

### 224 **3.2 Interspecies correlation estimation based on the ECOTOX database**

225 The parameters used to obtain the ICE values for different marine organisms under  
226 the exposure of the selected heavy metals were calculated based on the ECOTOX  
227 database, and the HC<sub>5</sub> value for different organisms was also calculated. As shown in  
228 Figure S2, if exposed to the same concentration levels (<1000 µg/L) of heavy metals,  
229 the hazardous levels of crustacean, fish and mollusc for each pollutant ranked as: Hg  
230 > Cu > As > Zn > Pb, Hg > Cu > Pb > Zn > As, and Hg > Cu > Zn > Pb, respectively.  
231 This point could also be reflected in the HC<sub>5</sub> value. For instance, the HC<sub>5</sub> values for  
232 Hg to marine organisms (1.3 µg/L for crustacean, 29.8 µg/L for fish and 9.0 µg/L for  
233 mollusc, respectively) were generally lower than those of the other heavy metals  
234 (Table 1). The corresponding HC<sub>5</sub> values for three groups of marine organisms by Cu  
235 were 19.5(9.4-40.5) µg/L, 48.4(29.1-80.3) µg/L, and 9.2(4.4-19.2) µg/L, respectively,  
236 whereas for Zn, they were 75.8(44.8-128.2) µg/L, 116.2(27.4-493.8) µg/L, and  
237 202.6(66.7-615.8) µg/L, respectively.

## 238 **4. Discussion**

239 First, we evaluated the geographic differences in the general ecological risks for all  
240 organisms (regardless of species types) induced by Cu, Zn, Pb, Hg and As in China's  
241 four adjacent seas. Among the four seas, the Bohai Sea had the highest ecological risk  
242 of all selected heavy metals except As, and the lowest values usually occurred in the  
243 South China Sea, except for some sites at the estuary of the Pearl River (shown in

244 Figure 2). The risk induced by Cu to all of the selected organisms in the Bohai Sea  
245 was highest among the four seas, but the potential risk did not exceed 2.1% (Figure  
246 2). For the other three seas, the potential ecological risks caused by Cu were all lower  
247 than 1.0%. The ecological risks induced by Zn were slightly lower than those induced  
248 by Cu in China's coastal waters, and the lowest values were observed in the East  
249 China Sea. The ecological risks caused by Pb in the four seas decreased in the  
250 following order: Bohai Sea > Yellow Sea > South China Sea > East China Sea (Figure  
251 2). Although marine organisms were more sensitive to Hg (i.e., lower HC<sub>5</sub> values  
252 compared with other heavy metals, as shown in Table 1), the higher potential  
253 ecological risks were usually caused by Cu and Zn (Figure 2) due to their higher  
254 concentrations in coastal seawater. The coastal regions with the highest potential  
255 ecological risks for each heavy metal are shown in Figure 3. Due to their higher  
256 concentrations, the major river estuaries (i.e., the Pearl and Yangtze Rivers) and the  
257 Bohai Sea were hotspots in terms of ecological risks. For instance, the ecological risk  
258 of Cu at the monitoring sites near Tianjin City in the Bohai Sea could be higher than  
259 2%.

260 Second, we evaluated the single ecological risks for each group of species induced  
261 by Cu, Zn, Pb, Hg and As in China's four adjacent seas. For different organism types,  
262 the heavy metals with the highest ecological risks were almost the same. For  
263 crustacean, the heavy metal with the highest ecological risk was Zn, and the  
264 ecological risk caused by Cu was just slightly lower than that of Zn. For fish and  
265 mollusc, the heavy metals were Zn and Cu, respectively (Figure 4). Significant  
266 variances in ecological risks could be observed among different seas. Due to the  
267 higher concentrations of heavy metals, the organisms in the Bohai Sea usually had the  
268 highest ecological risks, especially for the sampling sites near Jinzhou, Tianjin and  
269 Tangshan (Figure S3) where the potential ecological risk of Cu to mollusc could be  
270 even higher than 5%, which should draw the attention of local environmental  
271 authorities. The ecological risks caused by Pb varied significantly among different

272 groups of organisms, e.g., fish could have higher risk levels, while the risks to  
273 mollusc and crustacean could be neglected. The risks caused by Hg and As in the four  
274 seas were low due to their lowest concentrations in the water (Figure 4).

275 It must be noted that benthic invertebrates may mostly ingest heavy metals in the  
276 sediments (Lin et al., 2013; Guo et al., 2016); however, due to limited data, the risk  
277 assessment conducted in this study only evaluated the risks due to exposure in the  
278 water column. Guo et al. (2016) also noted that a significant increase in Hg levels was  
279 observed with increasing trophic levels of fish, and fish were found to have a high  
280 bioaccumulation factor for Hg. Therefore, the low risk of Hg due to exposure in the  
281 water column does not mean that the overall risk for exposure to Hg is low. Rather,  
282 the majority of Hg in fish potentially accumulated through the water and food web  
283 (Tong et al., 2017), and fish with high trophic levels should receive special attention.

284 Third, we evaluated the msPAF induced by all heavy metals to all organisms in  
285 China's coastal waters (Figure 5). In most of the risk distributions, the risk levels in  
286 the four seas decreased in the order of Bohai Sea > Yellow Sea > South China Sea >  
287 East China Sea. Additionally, for the ecological risks of all heavy metals to all  
288 organisms, higher levels usually occurred at the estuaries of major rivers flowing into  
289 seas. All of the calculated msPAFs values were lower than 5%, which was a safe  
290 value for aquatic organisms (Aldenberg and Slob, 1993; Newman et al., 2000). For  
291 the msPAF values of the five major heavy metals, the Bohai Sea was the highest  
292 among the four seas, where ~25% of the sites were above the risk of 2.5% and the  
293 highest risk in the Bohai Sea exceeded 3.0%, such as at the sampling sites in Jinzhou,  
294 Tangshan and Tianjin. The msPAF in the Yellow Sea was the second highest, and was  
295 much lower than that in the Bohai Sea. Almost all sites had a risk of less than 1.0%,  
296 and only one site in Dalian, Liaoning Province, had a relatively high risk of 3.5%.  
297 From the risk distribution in the Yellow Sea, we could see that all of the sampling  
298 sites in Dalian increased the average risk probability level of this area. For the South

299 China Sea, the percentage of the monitoring sites with msPAS values lower than 0.5%  
300 was ~77%.

## 301 **5. Conclusions**

302 The Bohai Sea contains the highest concentrations of Cu, Zn, Hg and Pb among the  
303 four Chinese adjacent seas, whereas the East China Sea contains the highest  
304 concentration of As. The river estuaries, such as the Yangtze River Estuary and the  
305 Pearl River Estuary, usually have higher concentrations of heavy metals compared  
306 with other regions.

307 Due to the high concentrations of heavy metals, Bohai Sea also poses the highest  
308 ecological risks for crustacean, fish and mollusc among the four Chinese adjacent  
309 seas. Most of the sites showed the affected proportion to be below 2.0%, and all sites  
310 had an affected proportion below 5.0%.

311 In general, the ecological risk level is low in all Chinese adjacent seas based on the  
312 current evaluation. However, hotspot areas may still exist, such as industrial cities  
313 along the Bohai Sea and the Pearl River Delta. Certain monitoring sites near Tianjin  
314 City and Jinzhou in the Bohai Sea had high ecological risks by Cu (>2% of the total)  
315 and Zn (>2.2% of the total), which should draw the attention of environmental  
316 authorities in the future.

317 The joint ecological risk posed by the five heavy metals to the three target  
318 organisms in coastal waters is relatively low, and the risk decreases in the following  
319 order: Bohai Sea > Yellow Sea > South China Sea > East China Sea. For crustacean,  
320 the heavy metal causing the highest ecological risk was Zn, and the ecological risk  
321 caused by Cu was slightly lower than that of Zn. For fish and mollusc, their  
322 corresponding heavy metals were Zn and Cu, respectively.

## 323 **Acknowledgements**

324 This study is funded by the National Natural Science Foundation of China (Grant  
325 #41501517, 41630748, and 41671492) and the Natural Science Foundation of Tianjin  
326 (Grant #16JCQNJC08300).

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#### 443 **Figure captions**

444 Figure 1. Concentrations of Cu, Zn, Pb, Hg and As in China's coastal waters

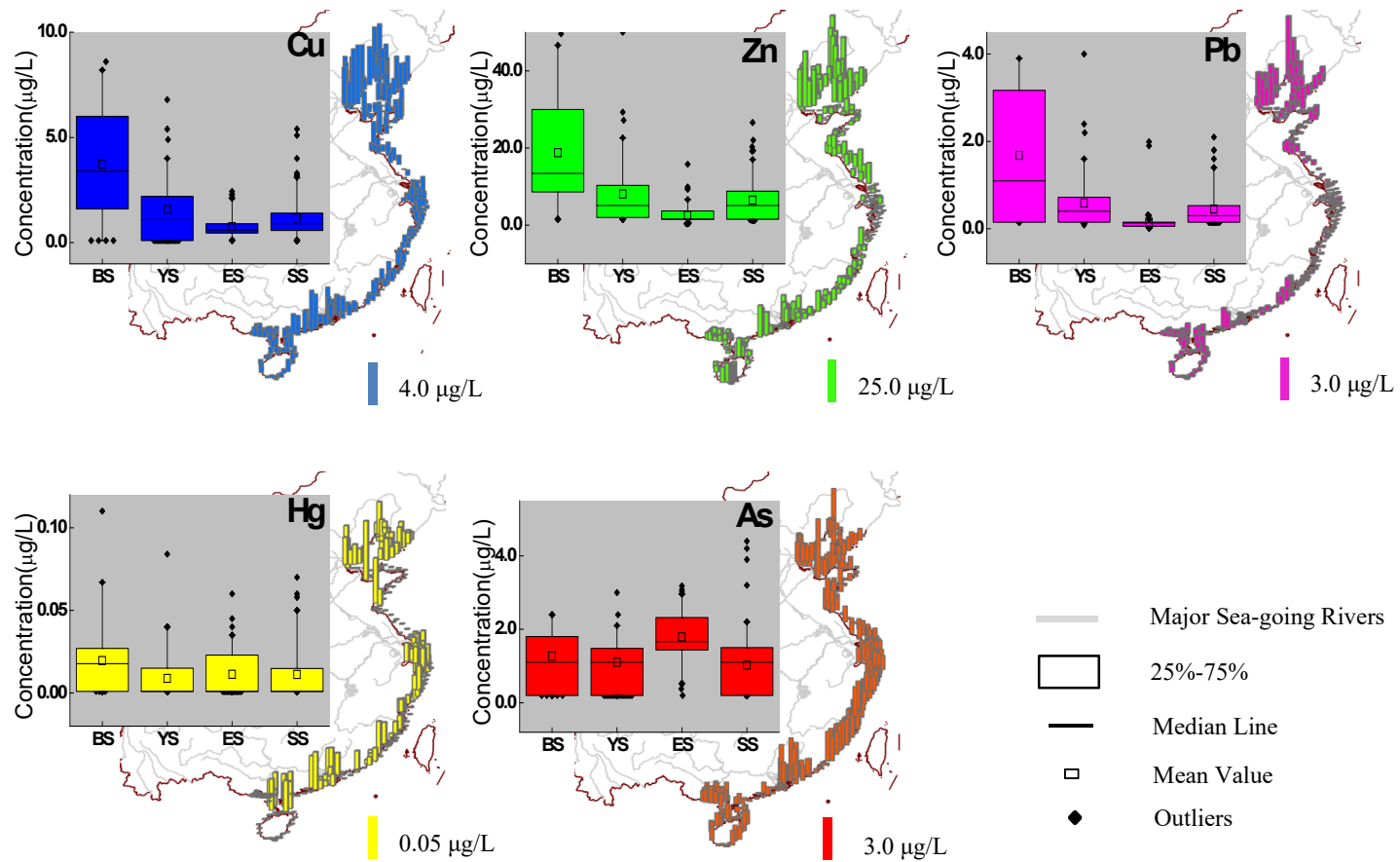
445 Figure 2. Ecological risks faced by all of the selected organisms induced by heavy metals

446 Figure 3. Coastal regions with the highest potential ecological risks

447 Figure 4. Ecological risks to crustacean, fish and mollusc by Cu, Zn, Pb, Hg and As

448 Figure 5. Combined ecological risks of multiple substance in China's coastal waters





**Figure 1. Concentrations of Cu, Zn, Pb, Hg and As in China's coastal waters**

(BS: Bohai Sea; YS: Yellow Sea; ES: East China Sea; SS: South China Sea)

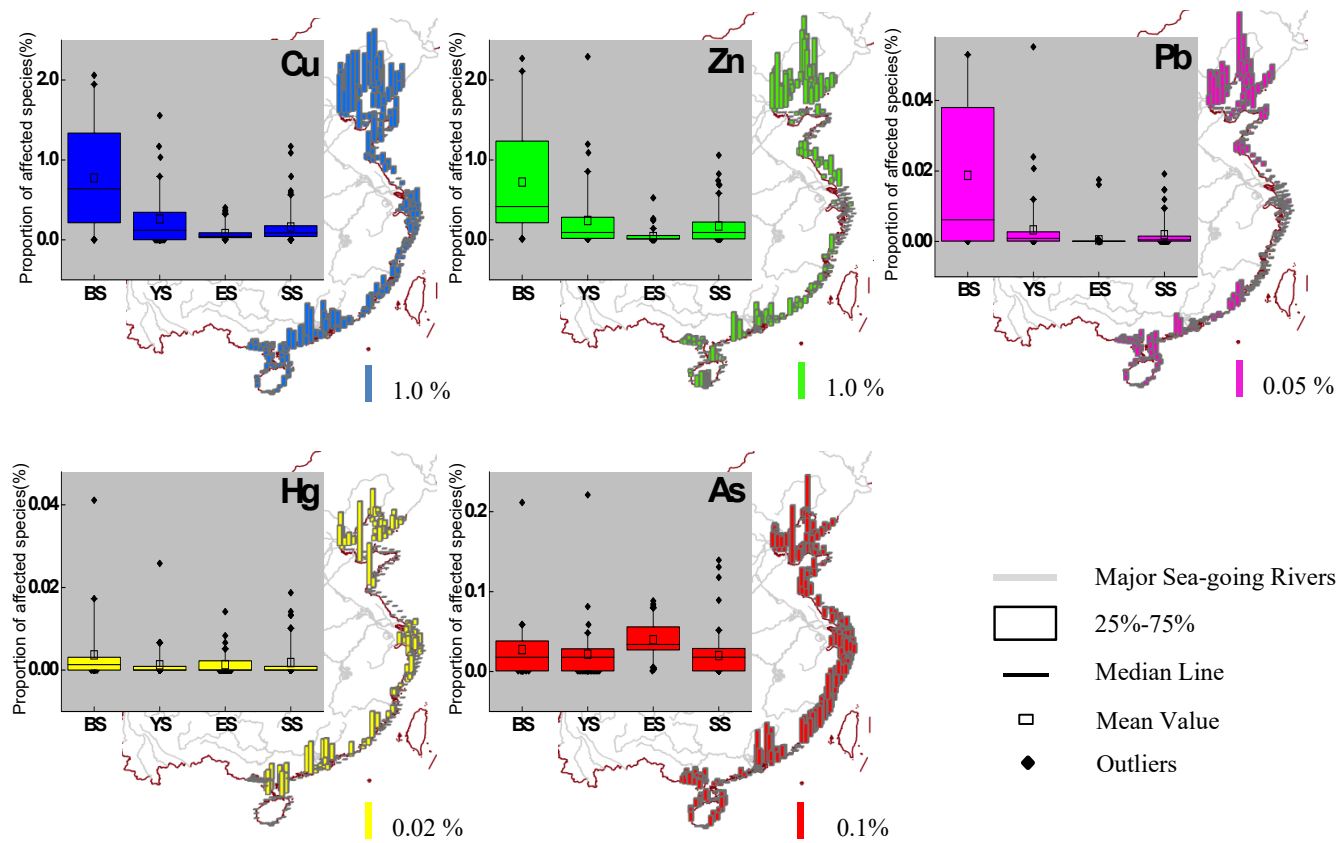


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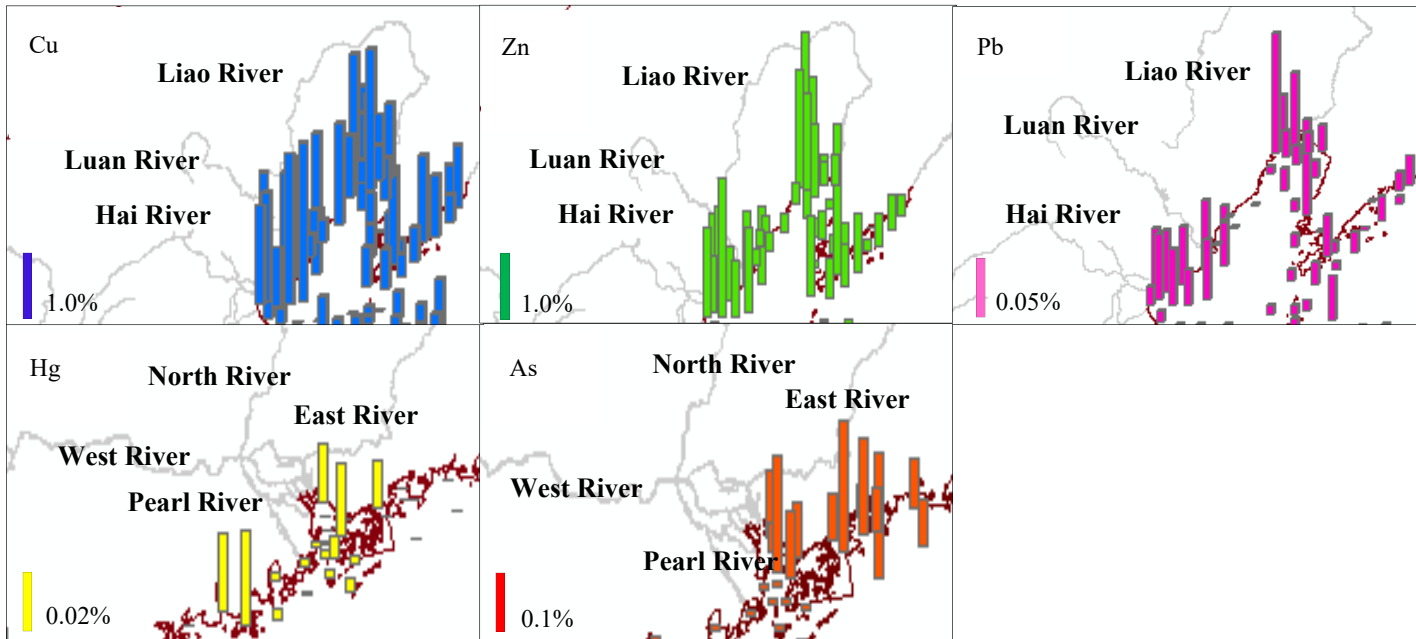
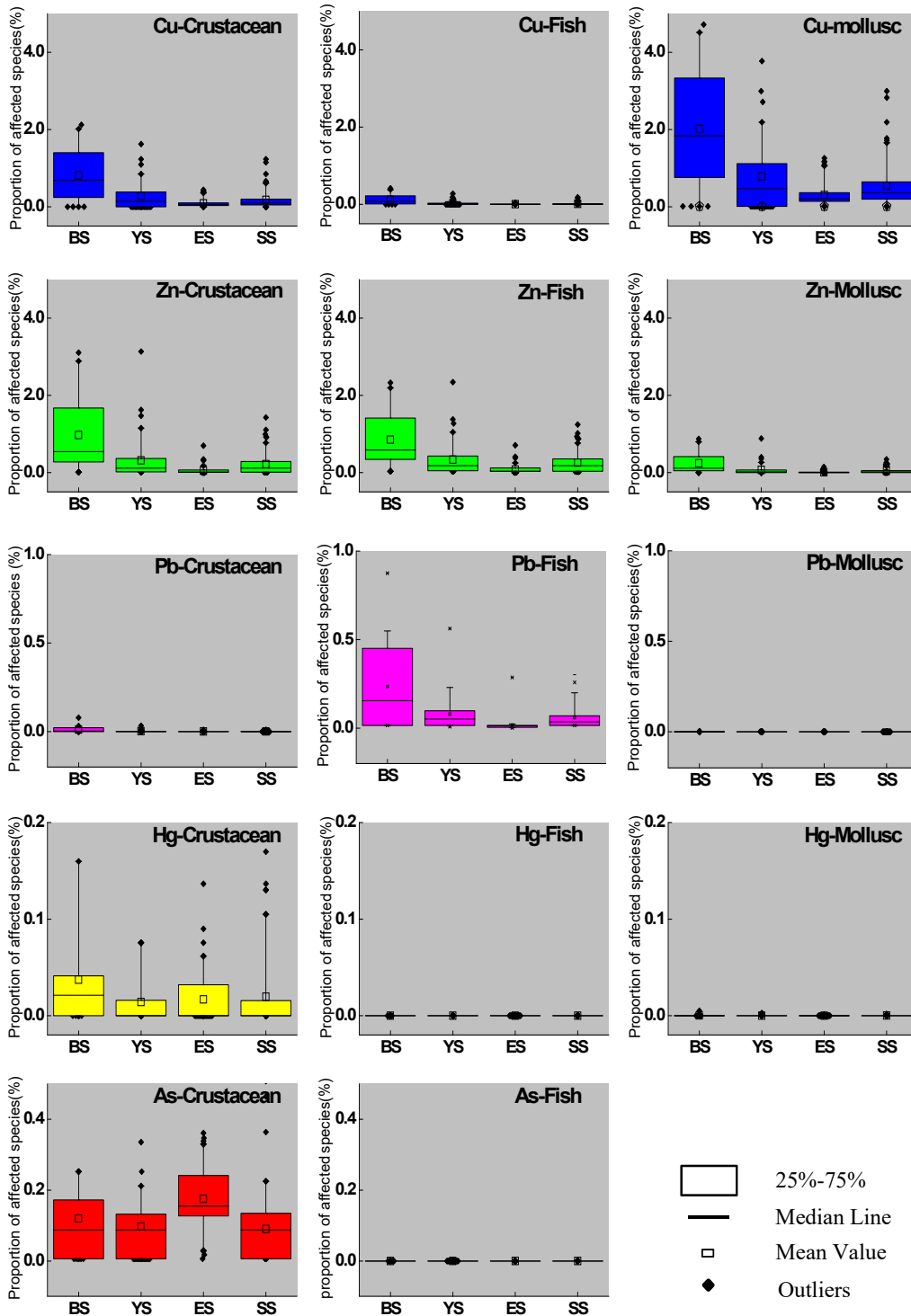
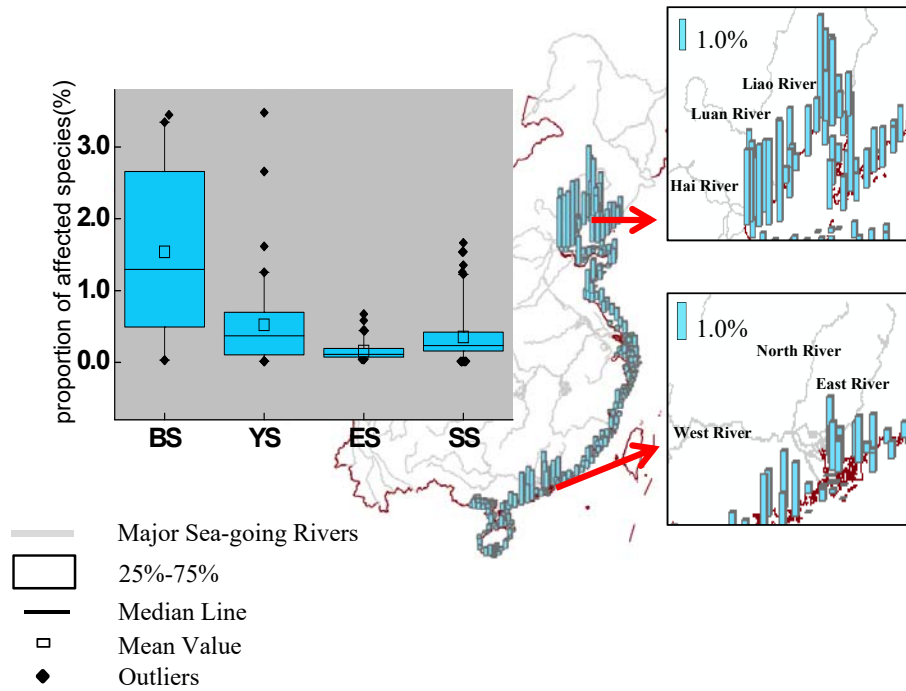


Figure 3. Coastal regions with the highest potential ecological risks



**Figure 4. Ecological risks to crustacean, fish and mollusc by Cu, Zn, Pb, Hg and**

**As**



**Figure 5. Combined ecological risks of multiple substance in China's coastal waters**