

Accepted Manuscript

This is an Accepted Manuscript of the following article:

Anqi Luo, Huihuang Chen, Xiaofei Gao, Laurence Carvalho, Yuanyuan Xue, Lei Jin, Jun Yang.
Short-term rainfall limits cyanobacterial bloom formation in a shallow eutrophic
subtropical urban reservoir in warm season.
Science of The Total Environment. Volume 827, 25 June 2022, 154172.

The article has been published in final form by Elsevier at
<http://dx.doi.org/10.1016/j.scitotenv.2022.154172>

© 2022. This manuscript version is made available under the

CC-BY-NC-ND 4.0 license

<http://creativecommons.org/licenses/by-nc-nd/4.0/>

1 **Short-term rainfall limits cyanobacterial bloom formation in a**
2 **shallow eutrophic subtropical urban reservoir in warm season**

3 Anqi Luo^{1,2}, Huihuang Chen^{1,2}, Xiaofei Gao^{1,3}, Laurence Carvalho^{4,5}, Yuanyuan Xue¹,
4 Jun Yang^{1,*}

5 ¹ *Aquatic EcoHealth Group, Fujian Key Laboratory of Watershed Ecology, Key Laboratory of*
6 *Urban Environment and Health, Institute of Urban Environment, Chinese Academy of Sciences,*
7 *Xiamen 361021, China*

8 ² *University of Chinese Academy of Sciences, Beijing 100049, China*

9 ³ *College of Fisheries, Henan Normal University, Xinxiang 453007, China*

10 ⁴ *UK Centre for Ecology & Hydrology, Penicuik, EH45 8EP, United Kingdom*

11 ⁵ *Norwegian Institute for Water Research, Oslo, NO-0579, Norway*

12 ***Corresponding author:** Jun Yang, E-mail address: jyang@iue.ac.cn.

13 **Abstract**

14 The global increase in dominance of toxic blooms of cyanobacteria has severely
15 impacted aquatic ecosystems and threatened human health for decades. Although it has
16 been shown that high levels of rainfall may inhibit the growth of cyanobacterial blooms,
17 it is still unclear how cyanobacteria respond to short-term rainfall events. Based on five-
18 year (2016–2020) high-frequency (half-week) sampling data from an eutrophic, urban,
19 subtropical reservoir, we explored the short-term effects of rainfall events on
20 cyanobacterial biomass (CBB) by constructing general additive models of CBB in rainy
21 periods during warm (April to September) and cool (December and January) months.
22 We find evidence in support of the hypotheses that short-term rainfall events
23 significantly reduce CBB in warm months, but the opposite response was observed in
24 the cool months. We also highlight a difference in the factors explaining CBB decreases

25 in warm months (precipitation, air temperature, relative humidity, dissolved oxygen and
26 total phosphorus) compared with factors explaining the response of CBB in cool
27 months (sunshine hours, pH and total carbon). In particular, meteorological factors
28 (precipitation, wind speed and sunlight) drive changes in water temperature and hydro-
29 dynamics of the reservoir causing a rapid reduction of CBB after rainfall events. This
30 varying response of cyanobacteria to short-term rainfall events in the eutrophic
31 subtropical reservoir may also be expected in temperate or cool lakes as climate change
32 effects become stronger.

33 **Keywords:** Climate change; phytoplankton; Precipitation; Inland water;

34 Meteorological factors

35 **1. Introduction**

36 Water is a natural resource, which is essential to all ecosystems and societies. In
37 recent decades, climate warming and human activities have caused a further
38 deterioration of global freshwater quality, through increasing the magnitude and
39 frequency of cyanobacterial blooms in inland freshwater systems (Sinha et al., 2017;
40 Huisman et al., 2018; Ho et al., 2019). Harmful cyanobacterial blooms are serious
41 issues that threaten human health and the water supply of millions of people with
42 associated huge economic losses (Carmichael et al., 2001; Michalak et al., 2013; Qin
43 et al., 2015; Carmichael and Boyer, 2016).

44 Climate variables are one of the main driving factors of cyanobacterial biomass at
45 different latitudes (Richardson et al., 2019; Giani et al., 2020). Rainfall events can affect
46 specific regional and watershed environments through impacts on surface runoff,
47 temperature and wind direction (Piao et al., 2010; Ho and Michalak, 2019; Weber et al.,
48 2020) and thus change the physico-chemical conditions in water bodies, further
49 boosting or abating cyanobacterial blooms (Reichwaldt et al., 2012; Yang et al., 2017).
50 Hydrological alterations can also affect water quality, ecosystem structure and function
51 (Webster et al., 2005; Piao et al., 2010; Yang et al., 2017). Shallow lakes in particular
52 are highly sensitive to wind-induced turbulence (Qin et al., 2015). Previous studies have
53 found that the phytoplankton (algae) biomass mostly decreases after rainfall events
54 (Havens et al., 2017). It has also been observed that high flushing rates can affect the
55 dominance of cyanobacteria (Carvalho et al., 2011; Tang et al., 2021). The timing,
56 frequency and volume of heavy rainfall events are key factors influencing
57 cyanobacterial biomass (Ahn et al., 2002; Yang et al., 2017; Richardson et al., 2019).

58 Lake ecosystems have unique characteristics with differences in morphometry,
59 hydrology, and human activities in individual watersheds. These differences result in

60 varying responses, such as frequency and magnitude of harmful algal blooms, to
61 environmental drivers (Paerl H et al., 2016). Rainfall-induced hydrological changes of
62 watersheds may benefit cyanobacterial growth by increasing nutrient inputs (Michalak
63 et al., 2013), but large storm events can have a negative effect on cyanobacterial blooms
64 by flushing cells from the waterbody (Reichwaldt and Ghadouani, 2012). The
65 urbanization caused by human activities can increase surface runoff and decrease
66 precipitation infiltration (Paul and Meyer, 2001; Grimm et al., 2008). Moreover,
67 urbanization-driven increases of point source pollution and non-point source pollution
68 can facilitate the occurrence and development of harmful algal blooms (Havens et al.,
69 2003; Hall and Leavitt., 1999).

70 A key challenge is to understand how climate change affects shorter-term
71 meteorological factors, such as precipitation, temperature and wind in specific regions,
72 and how these factors interact to produce changes in environmental conditions in a lake,
73 and ultimately, changes in water quality (Michalak, 2016). Moreover, it is necessary to
74 strengthen our understanding of how phytoplankton communities respond to short-term
75 storm events at fine scale (Stockwell et al., 2020). However, the bulk of previous studies
76 have primarily explored the relationship between seasonal averages of cyanobacterial
77 biomass and weather (Bouvy et al., 2003; Robson and Hamilton, 2004; Hampel et al.,
78 2020) or single extreme precipitation events (Zhang et al., 2012; Paerl et al., 2020; Gao
79 et al., 2021; Qin et al., 2021). Although we know rainfall can have a strong influence
80 on the occurrence of cyanobacterial blooms (Mu et al., 2019), we still don't fully
81 understand how these rainfall events change lake conditions and consequently affect
82 cyanobacterial biomass across space and time. Furthermore, there are no publications
83 that use long-term, high-frequency monitoring data to analyse the cyanobacteria-
84 rainfall relationship in sub-tropical climates. In view of these gaps, this study aimed to:

85 1) explore the cyanobacterial response in a sub-tropical reservoir to short-term rainfall
86 events using high-frequency (half a week) monitoring data; 2) construct models
87 explaining the dynamics of cyanobacterial biomass (CBB) during rainy periods in both
88 warm and cool months; 3) explore the specific driving factors behind the response of
89 CBB to rainfall episodes.

90 We hypothesize that: 1) the short-term decline of cyanobacterial biomass is relevant
91 to the magnitude and frequency of rainfall events; 2) short-term rainfall events affect
92 the cyanobacterial biomass-environment relationship differently in warm versus cool
93 months.

94 **2. Materials and methods**

95 *2.1. Study area, sampling and dataset*

96 High-frequency field sampling was conducted in the Xinglinwan Reservoir which
97 is located in the lower reaches of the Houxi River (about 25 km long), Xiamen City,
98 southeast China (Fig. 1a). The study site belongs to the subtropical maritime monsoon
99 climate and the average temperature of the year is 20.9 °C, the average wind speed is
100 5–6 m/s, the average annual rainfall is 1357 mm. Xinglinwan Reservoir is a shallow
101 and eutrophic reservoir with low transparency and long hydraulic retention time (Peng
102 et al., 2020; Yang et al., 2022). The total storage capacity of Xinglinwan Reservoir is
103 about 2568 m³, the average water depth is about 2.5 m and the water level (with mean
104 of about -0.5 m) always below the sea level (Fig. S1). The Xinglinwan Reservoir was
105 heavily influenced by agricultural activities and urbanization-induced land-use change
106 and sewage effluent in the past decades (Zhu et al., 2019). It's also worth noting that
107 Xinglinwan Reservoir is an enclosed water body as a dam blocks the outflow with
108 limited flushing (Fig. 1a), therefore precipitation across the natural catchment is the
109 major source of water for this reservoir and seawater intrusion in the bottom water layer

110 is stronger in winter (dry and cool) than in summer (wet and warm) seasons (Mo et al.,
111 2021).

112 Water samples were normally taken from the surface water (0.5 m depth) at two
113 stations (station L: 24° 36' 21" N, 118° 03' 37"; station G: 24° 36' 09" N, 118° 03' 59"
114 E) twice a week (Tuesday and Friday) from 6th January 2016 to 29th December 2020.
115 Our dataset (N = 1064) consists of a range of cyanobacterial and physicochemical data
116 of Xinglinwan Reservoir, the meteorological data of Xiamen City. In this study, the
117 cyanobacterial chlorophyll-*a* was used to represents the cyanobacterial biomass (CBB).
118 The cyanobacterial chlorophyll-*a* was specifically measured using a PHYTO-PAM
119 Phytoplankton Analyzer (Heinz Walz GmbH, Effeltrich, Germany), taking three
120 replicate measurements and calculate the average value for each sample. The physical
121 data (water temperature, pH, dissolved oxygen, turbidity and salinity) were measured
122 at the stations using a Hydrolab DS5 multiparameter water quality analyzer (Hach
123 Company, Loveland, CO, USA). The chemical data including total carbon (TC), total
124 nitrogen (TN), total organic carbon (TOC), ammonium-nitrogen (NH₄-N), nitrate-
125 nitrogen (NO₃-N), nitrite-nitrogen (NO₂-N), total phosphorus (TP), orthophosphate-
126 phosphorus (PO₄-P) were measured following standard methods (Greenberg et al.,
127 1992). The meteorological data (precipitation, evaporation, the average relative
128 humidity, min relative humidity, sunshine hours, average air temperature, max air
129 temperature, min air temperature, average surface temperature, max surface
130 temperature, min surface temperature, the average wind speed and max wind speed)
131 were extracted from the China Meteorological Data Service Center (CMDC) of the
132 National Meteorological Information Center in China (<http://data.cma.cn>). In addition,
133 daily temperature range was calculated by the difference between the max and min air
134 temperatures. Table S1 presents a descriptive analysis of all variables.

135 **2.2. Defining rainfall events and grouping**

136 In this study, we examined the cyanobacterial response within 7 days following
137 rainfall events. The total precipitation (mm) was calculated for X days prior to sampling
138 ($Pcpt_x$), where X indicates 1, 2, 3, 4, 5, 6 and 7 days, and used to define whether rainfall
139 events had occurred in the X days before the sampling time (day). If $Pcpt_x > 0$, it was
140 classified as a “rain” period. Otherwise, it was classified as a “non-rain” period. The
141 year was divided into 6 bi-monthly periods (“Tgroup”) for considering the seasonal
142 changes in CBB: T1 (sampled in February and March, N = 165), T2 (sampled in April
143 and May, N = 175), T3 (sampled in June and July, N = 174), T4 (sampled in August
144 and September, N = 208), T5 (sampled in October and November, N = 166) and T6
145 (sampled in December and January, N = 176). The T2–T4 are considered the warm
146 months, whereas the T6 covers the cool months. For each “Tgroup”, we used the
147 nonparametric Mann-Whitney *U* test to assess the significant differences in the CBB
148 between rain and non-rain periods. Only Tgroups with significant differences ($P < 0.05$)
149 in CBB were analyzed further.

150 **2.3. Factor analysis**

151 Factor analysis (FA) was used to remove the collinearity of variables in the dataset
152 and identify common factors in the environment variables that represent the correlated
153 variables. Initially, parallel analysis was used to identify the optimal number of factors
154 explaining CBB during rainy periods in warm and cool months, respectively. Through
155 the factor loading matrix, the variables with the highest factor scores were selected as
156 the common factors (Table S2; Table S3). The FA was performed using the package of
157 R studio 3.6.3 (R Core Team, 2020).

158 **2.4. Generalized additive model**

159 The identified factors that significantly influenced the effects of rainfall in warm
160 or cool months were used to model CBB based on a generalized additive model (GAM).
161 The factors with the highest score selected in FA were separately fitted to CBB in warm
162 or cool months, and the factors with P value less than 0.05 were selected as explanatory
163 factors. The selected covariates from FA were used as explanatory factors for CBB.
164 The covariate factors that had high influence on the R^2 -adjust of the two response
165 variables were identified, in order to explore the role of these explanatory factors in the
166 rainfall-CBB relationship.

167 Two regression functions were constructed based on CBB and their explanatory
168 factors in rain periods of the warm months (y_1) or the cool months (y_2). The y_1 ($N =$
169 350) is the regression function of CBB and corresponding explanatory variables
170 (precipitation, temperature, relative humidity, dissolved oxygen and total phosphorus)
171 in rain period of warm months (T2, T3 and T4). The y_2 ($N = 72$) is the regression
172 function of CBB and corresponding explanatory variables (sunshine hours, pH and total
173 carbon) in rain period of cool months (T6). In addition, difference tests were performed
174 again for these significant explanatory variables between rain and non-rain datasets
175 using the nonparametric Mann-Whitney U test. GAM was performed using the mgcv
176 package of R studio 3.6.3 ([R Core Team, 2020](#)).

177 ***2.5. The PLS path model***

178 Partial least squares path modeling (PLS-PM) is a general framework that can
179 analyse structural relationships between variables ([Tenenhaus et al., 2005](#)). We
180 constructed a statistic for detecting the relationship between meteorological factors and
181 CBB. In this study, PLS-PM described the causal links between the meteorological
182 factors in rain (precipitation, wind speed and sunlight) or non-rain (wind speed and
183 sunlight) periods, environmental factors (water temperature and dissolved oxygen et al),

184 and CBB. PLSPM was performed with the plsmpm package in a R studio 3.6.3 (R Core
185 Team, 2020).

186 **3. Result**

187 *3.1. Seasonal changes of cyanobacterial biomass in rain and non-rain periods*

188 The plot of seasonality in cyanobacterial biomass (CBB) showed a clear pattern
189 with an increase from T1 to T3 and general decrease from T3 to T6 (Fig. 1b; Fig. S1).
190 Meanwhile, from T2 to T4 (the warm months – April to September), CBB showed a
191 significant ($P < 0.05$) negative response to rainfall over the previous 1–4 days
192 (particularly 1–3 days) prior to the sampling day, compared with non-rainy periods
193 during the same months (Fig. 1b; Fig. 2a). The negative response to rainfall appeared
194 to increase from T2 to T4 consistent with the changes in the intensity of sunshine hours,
195 and the frequency of precipitation (Fig. 2b). After October (T5 and T6), the CBB
196 showed a tendency to increase after rainfall events, particularly in T6 in response to
197 rainfall over the previous 1–3 days (Fig. 2a).

198 Considering samplings with prior rain events, the meteorological factors of
199 precipitation, air temperature, relative humidity and sunshine hours in warm months
200 (T2, T3 and T4) were all higher than that in the cool months (T6), in particular air
201 temperature and sunshine hours in T3 and T4, and precipitation in T4 (Table 1). T1 to
202 T4 had more than 100 days of rain, while T5 had only 58 days with rain. Days of rainfall
203 in T6 (72 days) is slightly higher than that in T5 (58 days), but it is still less than that
204 from T1 to T4 (Table 1). The CBB significantly decreased in T2, T3 and T4 (the warm
205 months) with the higher average values of sunshine hours and precipitation and higher
206 frequency of sunshine hours > 5 h and days when precipitation > 0 (Fig. 2). Conversely,
207 the average value of CBB significantly increased in T6 (the cool months) corresponding
208 with lower average values and frequency of sunshine hours and precipitation (Fig. 2).

209 Unexpectedly, during the 3-day rain period when cyanobacterial biomass decreased in
210 Xinglinwan Reservoir, our data showed a lower wind speed (low average wind speed
211 and low frequency of wind speed $> 3\text{m/s}$) in warm months (Fig. 2b; Fig. 2c).

212 **3.2. Explanatory factors of CBB in warm or cool months**

213 In the warm months, this long-term and high-frequency study of Xinglinwan
214 Reservoir demonstrates that an increase of rainfall within 1–3 days can significantly
215 reduce the CBB, meanwhile the dissolved oxygen (DO) decreased in general (Fig. S2),
216 simultaneously. Considering the decreasing response of CBB after rainfall events (Fig.
217 2a), the significant explanatory variables ($P < 0.05$) included precipitation (Pcpt), air
218 temperature (AT), relative humidity (RH), dissolved oxygen (DO) and total phosphorus
219 (TP). These factors explained 33.1% of variance in the warm months ($P < 0.001$) of the
220 meteorological factors (Table 2). AT (R^2 -adjusted = 0.16) and DO (R^2 -adjusted = 0.12)
221 explained more than Pcpt, AT and RH and DO also showed a fit with CBB. The fitting
222 of 3-days precipitation and CBB showed a significant negative response in rain period
223 of warm months (Fig. 3). However, DO and total phosphorus showed different patterns
224 in GAM fitting performance in warm months between rain period and non-rain period
225 (Fig. 3).

226 Similarly, the explanatory factors in the cool months differed before and after
227 short-term rainfall events. A non-linear relationship was observed between sunshine
228 hours and total carbon with CBB in rainfall periods of the cool months (Fig. 4). During
229 the cool months, the increasing response of CBB to rainfall events was significantly
230 explained ($P < 0.05$) by daily sunshine hours (SH), pH and total carbon (TC), which
231 explained 49.3% of variance in T6 ($P < 0.001$) (Table 3). In addition, the GAM fitting
232 results of sunshine hours and CBB, pH and CBB, total carbon and CBB before and after
233 rain period also showed differences in cool months. Specifically, the fitting curve of

234 these significant factors are almost nonlinear in the period of rain, while tends to be
235 linear or stable in the period of non-rain (Fig. 4). The GAM fitting of TC performed
236 better (R^2 -adjusted = 0.18) than SH or pH (Table 3). Total carbon showed a slight trend
237 of decreasing after rainfall in general the cool months (Fig. 4), but this didn't make a
238 significant contribution to the CBB decreased (Fig. S2).

239 *3.3. The meteorological-environment relationship and the environment-CBB* 240 *relationship in rain or non-rain periods*

241 The relationship between CBB and meteorological variables (precipitation, wind
242 speed and sunlight) was significantly different between rain and non-rain periods. A
243 combination of precipitation, wind speed and sunlight significant resulted in the
244 decrease of CBB in rain periods ($P < 0.05$), while no significant causal relationship was
245 found on the PLS-PM results between CBB and the inputted combination of wind speed
246 and sunlight in non-rain periods (Fig. 5). In detail, when Xinglinwan Reservoir
247 experienced short-term rainfall, the overall combined effect of these meteorological
248 factors (precipitation, wind speed and sunlight) had a significant negative correlation
249 (path coefficient = -0.40) with the environmental factors (water temperature, pH,
250 dissolved oxygen) which had a positive correlation (path coefficient = 0.43) with
251 cyanobacterial biomass (Fig. 5a). However, The PLS-PM inputted combination of wind
252 speed and sunlight have a significant positive relationship (path coefficient = 0.55) with
253 the combined effect of water temperature, pH, and DO, which did not have a significant
254 relationship (path coefficient = 0.07) with cyanobacterial biomass in non-rain periods
255 (Fig. 5b). In addition, the combination of wind speed and sunlight contributed less (path

256 coefficient = -0.17) to the reduction of cyanobacterial biomass in non-rain conditions
257 (Fig. 5b).

258 **4. Discussion**

259 Descriptions of rainfall events indirectly controlling cyanobacterial biomass in
260 subtropical reservoirs or lakes have, until now, been mostly based on extreme rainfall
261 events, such as tropical storms, or a relatively long-term “seasonal average” temporal
262 scale (Michalak, 2016; Yang et al., 2017; Gao et al., 2021). Few studies have analysed
263 the short-term effects of rainfall events on cyanobacterial biomass at a high temporal
264 resolution (Michalak, 2016). This paper focused on the effects of rainfall events on the
265 environmental conditions that led to changes in cyanobacterial biomass in the short-
266 term.

267 ***4.1 Changes of explanatory factors of CBB after short-term rainfall***

268 In warm months of the Xinglinwan Reservoir, dissolved oxygen (DO) decreased
269 after the short-term rainfall in general. Short term rainfall seems to affect the
270 relationship between cyanobacterial biomass (CBB) and DO only at low concentrations
271 in Xinglinwan Reservoir. The relationship between CBB and DO may be correlative,
272 rather than causal, with rainfall affecting both these parameters. Stormwater runoff can
273 increase the inputs of organic matter to aquatic ecosystems, which could explain
274 declining DO concentrations and increasing turbidity in lake waters (McCabe et al.,
275 2021). However, the duration and extent of this impact is related to the frequency and
276 intensity of rainfall events in a specific reservoir (Li et al., 2015). Mostly, the small or
277 moderate rainfall in Houxi River may have limited the effects of disturbance in the
278 relationship between DO and CBB. At the same time, DO may fluctuate greatly during

279 a short-term rainfall period. In poor-quality rivers of temperate maritime climates of the
280 UK, increased rainfall events can flush away algal blooms and result in a large diurnal
281 variation in DO (Whitehead et al., 2009). The increasing vertical stratification and
282 precipitation in summer may cause variations in hypoxia volumes (Zhou et al., 2014).
283 The decline in DO took place in the presence of strong precipitation and southwesterly
284 winds in Xinglinwan Reservoir during the warm months.

285 In cool months, rainfall events changed the fitted relationship between CBB and
286 environmental factors to sunshine hours, pH, and total carbon and with CBB increasing
287 in Xinglinwan Reservoir. Sunshine hours have been observed as the primary factor
288 affecting cyanobacterial growth (Becker et al., 2010), and have been shown to be
289 positively correlated with cyanobacterial abundance (Zhang et al., 2012; Hu et al.,
290 2018). Sunshine hours, as an explanatory factor in y_2 , was significantly decreased by
291 3-days of rainfall, meanwhile an increase of CBB sensitivity to sunlight hours. Weak
292 sunlight and temperature may result in decreased cyanobacterial biomass, conversely
293 increased available light, may facilitate underwater photosynthetic active radiation,
294 which encourages phytoplankton growth (Álvarez et al., 2005). The relationship
295 between pH and CBB in Xinglinwan Reservoir changed slightly with rainfall in cool
296 months and exhibited a linear fitting in general, except for a slightly positive
297 relationship in rain periods. At the same time, a nonlinear relationship was observed
298 during the rainfall events between the total carbon and CBB.

299

300 **4.2 Precipitation reduced the cyanobacterial biomass in Xinglinwan Reservoir in**
301 **short term**

302 The spatial and temporal extent of cyanobacterial blooms is normally affected by
303 changing patterns in meteorological and hydrological conditions of specific watershed
304 (Vaicute et al., 2021). In Xinglinwan Reservoir, the increase in the intensity and
305 frequency of short-term rainfall, ultimately led to the decline of CBB in both warm and
306 cool months. Precipitation is considered to be the driving factor for the reduction of
307 cyanobacterial biomass in Xinglinwan Reservoir as the results of the path analysis in
308 rainy period and non-rain period are opposite due to the loading of the precipitation
309 factors (Fig. 5). An 18-year dataset analysis from seven shallow lakes in Florida showed
310 evidence that hydrologic factors (rainfall and flow) can suppress cyanobacterial blooms
311 (Havens et al., 2017). The increasing frequency of rainfall events results in higher
312 precipitation, lower temperatures and increases in wind speed, which increase the
313 disturbance in lakes. Alongside this, there are increased loss rates through flushing that
314 reduced the biomass of phytoplankton in a short-time in Xinglinwan Reservoir.

315 Lower sunlight appears as another significant factor in the decline of CBB in the
316 rainy period while no significant contribution to the increases of CBB in 3-day non-
317 rain period in Xinglinwan Reservoir. It is well known that warmer temperatures favour
318 the growth of bloom-forming cyanobacteria because they create the warm and static
319 conditions that reduce vertical mixing in the lake (Michalak, 2013; Woolway et al.,
320 2019). The sensitivity of cyanobacterial dynamics to climatic conditions can, however,
321 vary by specific region (Shi et al., 2017). Our results highlight high sunlight is
322 correlated with increased CBB in non-rain periods, but the effects were not significant
323 in the short-term (within 3 days). In the warm months of Xinglinwan Reservoir, the
324 average value of wind speed and the frequency of strong wind are less than those in the

325 cool months, but the cyanobacterial biomass decreases significantly during rainfall in
326 the warm months. In the end, the negative effects of rainfall events (especially
327 precipitation increase and low sunshine) appear to be the main reasons for the decrease
328 of CBB in Xinglinwan Reservoir in the short-term.

329 In addition, Xinglinwan reservoir has been in a state of eutrophication with a low
330 resource utilization efficiency (RUE) of cyanobacteria compared to the upstream low
331 nutrient reservoirs (Yang et al., 2022). In this study, we only considered the response of
332 cyanobacterial biomass to a rainfall event within three days. Although precipitation can
333 cause nutrient inputs through increased surface runoff, much of these nutrients are in
334 complex forms and so are not readily available for cyanobacteria to use efficiently
335 within the first three days in Xinglinwan Reservoir. As Alvarez-Cobelas et al., (2006)
336 indicated, rainfall events can have a delayed effect on groundwater fluxes in seepage
337 lakes, and catchment run-off may not have an immediate effect on lake nutrient
338 concentrations, resulting in limited direct effects of nutrients on cyanobacterial blooms
339 from rainfall in the short term. Following this process, it seems impractical to consider
340 nutrient factors in the path analysis at Xinglingwan Reservoir. Another reason cannot
341 be ignored is that a given cyanobacterial bloom may move to the central water body
342 (Umehara et al., 2015). Stations L and G are located at the inflow of Xinglinwan
343 Reservoir, where cyanobacteria can be transferred to open water by physical processes
344 such as water flushing and dilution caused by rainfall.

345 ***4.3 The reasons that cyanobacterial biomass increased in T6 (cool months)***

346 Unexpectedly, CBB significantly increased following the rainfall in the cool
347 months within 3 days in Xinglinwan Reservoir, unlike the warm months (Fig. 2a).
348 Besides, large differences in cyanobacterial communities were found in Xinglinwan
349 Reservoir between summer (July) and winter (January). Further, individual species

350 responses to rainfall in warm or cool months may differ from these more general
351 responses observed for cyanobacteria as a whole. The effect of precipitation and wind
352 on CBB was great as in periods of T1 vs T6 and T2 vs T5 there were similar values for
353 air temperature, relative humidity and daily sunshine hours but the significance of CBB
354 change from rain to non-rain periods was different (Figs. 1b; 2a) and appeared to be
355 strongly influenced by precipitation and wind. During the cool months (T6), the CBB
356 was relatively low in Xinglinwan Reservoir, while meteorological conditions in
357 Xinglinwan Reservoir during the cool months showed a relatively weak intensity of
358 sunshine (less than 5 hours), relatively high frequencies of high wind speeds (more than
359 3 m/s), a relatively low frequency of precipitation and a decline of CBB. This is not,
360 however, the first study that shows cyanobacteria can increase in subtropical regions in
361 cool months. Many studies have already indicated negative relationships between wind
362 speed and sunshine hours to cyanobacteria, such as *Microcystis* (Jöhnk et al., 2008;
363 Zhang et al., 2012; Qu et al., 2019). One example was observed in Lake Chaohu, where
364 the cyanobacterial biomass in winter was highest sometimes while the phytoplankton
365 abundance in general was lowest in summer (Jiang et al., 2014; Guan et al., 2020).
366 Specifically, this may be because these studies are in, or include, subtropical climates
367 where cool months are still relatively warm and favourable to cyanobacteria. The
368 growth and the dominance of cyanobacteria will be promoted when the average air
369 temperature is mostly over 20 °C (Anneville et al., 2015; Yang et al., 2017; Weber et
370 al., 2020). This may happen in Xinglinwan Reservoir which has relatively warm days
371 in winter. On the other hand, the lower temperature of the cool months (compared to
372 the warm months) can reduce the loss rate of some cyanobacteria by lower predation
373 pressure (Ma et al., 2016). In this way, the weakening effect in the cool months could
374 be due to less intense and frequency of rainfall events on cyanobacterial biomass,

375 compared to the high intensity storms more typical of the warm months (i.e. high-
376 frequency storm-driven perturbations in warm months). Our study highlights the need
377 for more published studies from subtropical regions, where patterns observed between
378 “cool” and “warm” months may be very different from the dominant published studies
379 from cool temperate and boreal regions. This complex or multiple responses of
380 cyanobacteria to short-term rainfall events in the eutrophic subtropical reservoir may
381 also be expected in temperate or cool lakes as climate change effects become stronger
382 in the future.

383 **5. Conclusion**

384 To elucidate the short-term effects of rainfall events on cyanobacterial biomass,
385 we analyzed high-frequency (half a week) sampling data of Xinglinwan Reservoir
386 combined with local meteorological information from 2016 to 2020. We considered the
387 condition of rainfall events, and examined the 3-day response of cyanobacteria to
388 specific environmental factors in both warm and cool months. We observed that in both
389 warm and cool months, short-term rainfall (with 3 days) significantly explained reduced
390 or boosted cyanobacterial biomass, respectively. Furthermore, this correlative
391 relationship between proximal meteorological (especially precipitation) forcing factors
392 and the reduction of cyanobacterial biomass was shown to be significant and we have
393 provided potential causative mechanisms to explain these observations. The approaches
394 and findings of this study offer an important insight into short-term response of
395 cyanobacterial biomass to rainfall in subtropical monsoon regions. It can provide an
396 analytical framework for the study of the relationship between cyanobacteria and
397 rainfall events in the future. Finally, it is also an important study to understand the
398 impacts of hydrological variability that informs water quality managers when they are

399 developing strategies to reduce harmful cyanobacterial blooms in subtropical water
400 bodies.

401 **Declaration of competing interest**

402 The authors declare that they have no known competing financial interests or personal
403 relationships that have influenced the work reported in this paper.

404 **Acknowledgments**

405 This work was supported by the Strategic Priority Research Program of the Chinese
406 Academy of Sciences (XDA23040302), the National Natural Science Foundation of
407 China (91851104, 32001152, and 92047204), and the Natural Science Foundation of
408 Fujian Province (2019J02016 and 2020J05089). Laurence Carvalho was supported by
409 the Natural Environment Research Council award number NE/R000131/1 as part of the
410 SUNRISE programme delivering National Capability. We thank Prof. Hans W. Paerl
411 for the constructive comments.

412 **Author contributions**

413 J.Y. conceived the ideas and designed the experiments; H.C. and X.G. collected and
414 determined the samples; A.L. and J.Y. analyzed the data and led the writing of the
415 manuscript; L.C. and Y.X. discussed the analytical approach and contributed to the
416 writing. All authors contributed to revisions and approved the final version of the
417 manuscript.

418 **Supplementary materials**

419 Supplementary material associated with this article can be found online.

420 **References**

421 Ahn, C.Y., Chung, A.K., Oh, H.M., 2002. Rainfall, phycocyanin, and N:P ratios related to cyanobacterial
422 blooms in a Korean large reservoir. *Hydrobiologia* 474, 117–124.

423 Álvarez, E., E. Nogueira, J. L. Acuña, M. López-Álvarez, and J. A. Sostres., 2009. Short-term dynamics
424 of late-winter phytoplankton blooms in a temperate ecosystem (Central Cantabrian Sea, Southern
425 Bay of Biscay). *J. Plankton Res.* 31: 601–617.

426 Alvarez-Cobelas, M., Cirujano, S., Rojo, C., Rodrigo, M.A., Piña, E., Rodríguez-Murillo, J.C., Montero,
427 E., 2006. Effects of changing rainfall on the limnology of a Mediterranean, Flowthrough-Seepage
428 Chain of Lakes. *Int. Rev. Hydrobiol.* 91(5), 466–482.

429 Anneville, O., Domaizon, I., Kerimoglu, O., Rimet, F., Jacquet, S., 2015. Blue-green algae in a
430 “greenhouse century”? new insights from field data on climate change impacts on cyanobacteria
431 abundance. *Ecosystems* 18(3), 441–458.

432 Becker, V., Caputo, L., Ordonez, J., Marce, R., Armengol, J., Crossetti, L.O., Huszar, V.L., 2010. Driving
433 factors of the phytoplankton functional groups in a deep Mediterranean reservoir. *Water Res.* 44(11),
434 3345–3354.

435 Bouvy, M., Nascimento, S.M., Molica, R.J.R., Ferreira, A., Huszar, V., Azevedo, S.M.F.O., 2003.
436 Limnological features in Tapacurá Reservoir (northeast Brazil) during a severe drought.
437 *Hydrobiologia* 493, 115–130.

438 Carmichael, W.W., Azevedo, S.M.F.O., AN, J.S., Molica, R.J.R., Jochimsen, E.M., Lau, S., Rinehart,
439 G.R.S., Eaglesham, G.K., 2001. Human fatalities from cyanobacteria: chemical and biological
440 evidence for cyanotoxins. *Environ. Health. Persp.* 109, 663–668.

441 Carmichael, W.W., Boyer, G.L., 2016. Health impacts from cyanobacteria harmful algae blooms:
442 implications for the North American Great Lakes. *Harmful Algae* 54, 194–212.

443 Carvalho, L., Miller, C.A., Scott, E.M., Codd, G.A., Davies, P.S., Tyler, A.N., 2011. Cyanobacterial
444 blooms: statistical models describing risk factors for national-scale lake assessment and lake
445 management. *Sci. Total. Environ.* 409, 5353–5358.

446 Cho, S., Lim, B., Jung, J., Kim, S., Chae, H., Park, J., Park, S., Park, J.K., 2014. Factors affecting algal
447 blooms in a man-made lake and prediction using an artificial neural network. *Measurement* 53, 224–
448 233.

449 Ferrenberg, S., Reed, S.C., and Belnap, J., 2015. Climate change and physical disturbance cause similar
450 community shifts in biological soil crusts. *Proc. Natl Acad. Sci. USA.* 112, 12116–12121.

451 Gao, X.F., Chen, H.H., Gu, B.H., Jeppesen, E., Xue, Y.Y., Yang, J., 2021. Particulate organic matter as
452 causative factor to eutrophication of subtropical deep freshwater: role of typhoon (tropical cyclone)
453 in the nutrient cycling. *Water Res.* 188, 116470.

454 Giani, A., Taranu, Z.E., von Ruckert, G., Gregory-Eaves, I., 2020. Comparing key drivers of
455 cyanobacteria biomass in temperate and tropical systems. *Harmful Algae.* 97, 101859.

456 Greenberg, A., Clesceri, L., Eaton, A. 1992. *Standard Methods for the Examination of Water and*
457 *Wastewater.* 16th ed. Washington DC, USA: American Public Health Association.

458 Grimm, N.B., Faeth, S.H., Golubiewski, N.E., Redman, C.L., Wu, J.G., Bai, X.M., et al., 2008. Global
459 change and the ecology of cities. *Science* 319, 756–760.

460 Guan, Y., Zhang, M., Yang, Z., Shi, X., Zhao, X., 2020. Intra-annual variation and correlations of
461 functional traits in *Microcystis* and *Dolichospermum* in Lake Chaohu. *Ecol. Indic.* 111, 106052.

462 Hall, R.I and Leavitt, P.R., 1999. Effects of agriculture, urbanization, and climate on water quality in the
463 northern Great Plains. *Limnol. Oceanogr.* 44, 739–756.

464 Hampel, J.J., McCarthy, M.J., Aalto, S.L., Newell, S.E., 2020. Hurricane disturbance stimulated

465 nitrification and altered ammonia oxidizer community structure in Lake Okeechobee and St. Lucie
466 Estuary (Florida). *Front. Microbiol.* 11, 1541.

467 Havens, K.E., James, R.T., East, T.L.; Smith, V.H., 2003. N:P ratios, light limitation, and cyanobacterial
468 dominance in a subtropical lake impacted by non-point source nutrient pollution. *Environ. Pollut.*
469 122, 379–390.

470 Havens, K.E., Ji, G., Beaver, J.R., Fulton, R.S., Teacher, C.E., 2017. Dynamics of cyanobacteria blooms
471 are linked to the hydrology of shallow Florida lakes and provide insight into possible impacts of
472 climate change. *Hydrobiologia* 829, 43–59.

473 Ho, J.C., Michalak, A.M., Pahlevan, N., 2019. Widespread global increase in intense lake phytoplankton
474 blooms since the 1980s. *Nature* 574(7780), 667–670.

475 Hu, M.Q., Zhang, Y.C., Ma, R.H., Zhang, Y.X., 2018. Spatial and temporal dynamics of floating algal
476 blooms in Lake Chaohu in 2016 and their environmental drivers. *Environ. Sci.* 39, 87–99.

477 Huisman, J., Codd, G.A., Paerl, H.W., Ibelings, B.W., Verspagen, J.M.H., Visser, P.M., 2018.
478 Cyanobacterial blooms. *Nat. Rev. Microbiol.* 16(8), 471–483.

479 Jiang, Y.J., He, W., Liu, W.X., Qin, N., Ouyang, H.L., Wang, Q.M., Kong, X.Z., He, Q.S., Yang, C., Yang,
480 B., Xu, F.L., 2014. The seasonal and spatial variations of phytoplankton community and their
481 correlation with environmental factors in a large eutrophic Chinese lake (Lake Chaohu). *Ecol. Indic.*
482 40, 58–67.

483 Jing, Y., Zhang, Y., Hu, M., Chu, Q., Ma, R., 2019. MODIS-satellite-based analysis of long-term
484 temporal-spatial dynamics and drivers of algal blooms in a plateau Lake Dianchi, China. *Remote*
485 *Sens.* 11(21), 2582.

486 Jöhnk, K.D., Huisman, J.E.F., Sharples, J., Sommeijer, B.E.N., Visser, P.M., Stroom, J.M., 2008. Summer
487 heatwaves promote blooms of harmful cyanobacteria. *Glob. Chang. Biol.* 14(3), 495–512.

488 Li, X., Huang, T.L., Ma, W.X., Sun, X., Zhang, H.H., 2015. Effects of rainfall patterns on water quality
489 in a stratified reservoir subject to eutrophication: Implications for management. *Sci. Total. Environ.*
490 521–522, 27–36.

491 Ma, J., Qin, B., Paerl, H.W., Brookes, J.D., Hall, N.S., Shi, K., Zhou, Y., Guo, J., Li, Z., Xu, H., Wu, T.,
492 Long, S., 2016. The persistence of cyanobacterial (*Microcystis* spp.) blooms throughout winter in
493 Lake Taihu, China. *Limnol Oceanogr.* 61(2), 711–722.

494 McCabe, K.M., Smith, E.M., Lang, S.Q., Osburn, C.L., Benitez-Nelson, C.R., 2021. Particulate and
495 dissolved organic matter in stormwater runoff influences oxygen demand in urbanized headwater
496 catchments. *Environ. Sci. Technol.* 55(2), 952–961.

497 Michalak, A.M., 2013. Anderson, E.J., Beletsky, D., Boland, S., Bosch, N.S., Bridgeman, T.B., Chaffin,
498 J.D., Cho, K., Confesor, R., Daloglu, I., Depinto, J.V., Evans, M.A., Fahnenstiel, G.L., He, L., Ho,
499 J.C., Jenkins, L., Johengen, T.H., Kuo, K.C., Laporte, E., Liu, X., McWilliams, M.R., Moore, M.R.,
500 Posselt, D.J., Richards, R.P., Scavia, D., Steiner, A.L., Verhamme, E., Wright, D.M., Zagorski, M.A.
501 Record-setting algal bloom in Lake Erie caused by agricultural and meteorological trends consistent
502 with expected future conditions. *Proc. Natl Acad. Sci. USA.* 110(16), 6448–6452.

503 Michalak, A.M., 2016. Study role of climate change in extreme threats to water quality. *Nature* 534, 340–
504 350.

505 Mo, Y.Y., Peng, F., Gao, X.F., Xiao, P., Logares, R., Jeppesen, E., Ren, K.X., Xue, Y.Y., Yang, J., 2021.
506 Low shifts in salinity determined assembly processes and network stability of microeukaryotic
507 plankton communities in a subtropical urban reservoir. *Microbiome* 9, 128.

508 Mu, M., Wu, C., Li, Y., Lyu, H., Fang, S., Yan, X., Liu, G., Zheng, Z., Du, C., Bi, S., 2019. Long-term

509 observation of cyanobacteria blooms using multi-source satellite images: a case study on a cloudy
510 and rainy lake. *Environ. Sci. Pollut. Res.* 26(11), 11012–11028.

511 Paerl, H.W., Gardner, W.S., Havens, K.E., Joyner, A.R., McCarthy, M.J., Newell, S.E., Qin, B., Scott,
512 J.T., 2016. Mitigating cyanobacterial harmful algal blooms in aquatic ecosystems impacted by
513 climate change and anthropogenic nutrients. *Harmful Algae* 54, 213–222.

514 Paerl, R.W., Venezia, R.E., Sanchez, J.J., Paerl, H.W., 2020. Picophytoplankton dynamics in a large
515 temperate estuary and impacts of extreme storm events. *Sci. Rep.* 10, 22026.

516 Paul, M.J., Meyer, J.L., 2001. Streams in the urban landscape. *Ann. Rev. Ecol. Syst.* 32, 333–365.

517 Peng, F., Guo, Y.Y., Isabwe, A., Chen, H.H., Wang, Y.M., Zhang, Y.P., Zhu, Z.X., Yang, J., 2020.
518 Urbanization drives riverine bacterial antibiotic resistome more than taxonomic community at
519 watershed scale. *Environ. Int.* 137, 105524.

520 Piao, S., Ciais, P., Huang, Y., Shen, Z., Peng, S., Li, J., Zhou, L., Liu, H., Ma, Y., Ding, Y., Friedlingstein,
521 P., Liu, C., Tan, K., Yu, Y., Zhang, T., Fang, J., 2010. The impacts of climate change on water
522 resources and agriculture in China. *Nature* 467(7311), 43–51.

523 Qin, B., Deng, J., Shi, K., Wang, J., Brookes, J., Zhou, J., Zhang, Y., Zhu, G., Paerl, H.W., Wu, L., 2021.
524 Extreme Climate Anomalies Enhancing Cyanobacterial Blooms in Eutrophic Lake Taihu, China.
525 *Water Resources Res.* 57(7), e2020WR029479.

526 Qin, B., Li, W., Zhu, G., Zhang, Y., Wu, T., Gao, G., 2015. Cyanobacterial bloom management through
527 integrated monitoring and forecasting in large shallow eutrophic Lake Taihu (China). *J Hazard*
528 *Mater.* 287, 356–363.

529 Qu, M., Anderson, S., Lyu, P., Malang, Y., Lai, J., Liu, J., Jiang, B., Xie, F., Liu, H.H.T., Lefebvre, D.D.,
530 Wang, Y.S., 2019. Effective aerial monitoring of cyanobacterial harmful algal blooms is dependent
531 on understanding cellular migration. *Harmful Algae.* 87, 101620.

532 R Core Team, 2020. R: A Language and Environment for Statistical Computing. R Foundation for
533 Statistical Computing, Vienna, Austria <https://www.r-project.org>.

534 Reichwaldt, E.S., Ghadouani, G., 2012. Effects of rainfall patterns on toxic cyanobacterial blooms in a
535 changing climate: Between simplistic scenarios and complex dynamics. *Water Res.* 46, 1372–1393.

536 Richardson, J., Feuchtmayr, H., Miller, C., Hunter, P.D., Maberly, S.C., Carvalho, L., 2019. Response of
537 cyanobacteria and phytoplankton abundance to warming, extreme rainfall events and nutrient
538 enrichment. *Glob. Chang. Biol.* 25(10), 3365–3380.

539 Robson, B.J., Hamilton, D.P., 2004. Three-dimensional modelling of a *Microcystis* bloom event in the
540 Swan River estuary, Western Australia. *Ecol. Model.* 174(1-2), 203–222.

541 Salmaso, N., 2010. Long-term phytoplankton community changes in a deep subalpine lake: responses to
542 nutrient availability and climatic fluctuations. *Freshwater Biol.* 55(4), 825–846.

543 Shi, K., Zhang, Y.L., Zhou, Y.Q., Liu, X.H., Zhu, G.W., Qin, B.Q., Gao, G., 2017. Long-term MODIS
544 observations of cyanobacterial dynamics in Lake Taihu: responses to nutrient enrichment and
545 meteorological factors. *Sci. Rep.* 7, 40326.

546 Sinha, E., Michalak, A.M., Balaji, V., 2017. Eutrophication will increase during the 21st century as a
547 result of precipitation changes. *Science* 357, 405–408.

548 Stockwell, J.D., Doubek, J.P., Adrian, R., Anneville, O., Carey, C.C., Carvalho, L., De Senerpont Domis,
549 L.N., Dur, G., Frassl, M.A., Grossart, H.P., Ibelings, B.W., Lajeunesse, M.J., Lewandowska, A.M.,
550 Llamas, M.E., Matsuzaki, S.S., Nodine, E.R., Noges, P., Patil, V.P., Pomati, F., Rinke, K., Rudstam,
551 L.G., Rusak, J.A., Salmaso, N., Seltnann, C.T., Straile, D., Thackeray, S.J., Thiery, W., Urrutia-
552 Cordero, P., Venail, P., Verburg, P., Woolway, R.I., Zohary, T., Andersen, M.R., Bhattacharya, R.,

553 Hejzlar, J., Janatian, N., Kpodonu, A., Williamson, T.J., Wilson, H.L., 2020. Storm impacts on
554 phytoplankton community dynamics in lakes. *Glob. Chang. Biol.* 26(5), 2756–2784.

555 Tang, Q.H., Lei, L.M., Zhao, L., Gu, J.G., Xiao, L.J., Han, B.P., 2021. Interactive effect of water level
556 and flushing rate on population dynamics of a harmful cyanobacterial species: *Raphidiopsis*
557 *raciborskii*. *Ecotoxicology* 30, 936–944.

558 Tenenhaus, M., Vinzi, V.E., Chatelin, Y.M., Lauro, C., 2005. PLS path modeling. *Comput. Stat. Data An.*
559 48(1), 159–205.

560 Umehara, A., Komorita, T., Tai, A., Takahashi, T., Orita, R., Tsutsumi, H., 2015. Short-term dynamics of
561 cyanobacterial toxins (microcystins) following a discharge from a coastal reservoir in Isahaya Bay,
562 Japan. *Mar. Pollut. Bull.* 92, 73–79.

563 Vaiciute, D., Bucas, M., Bresciani, M., Dabuleviciene, T., Gintauskas, J., Mezine, J., Tiskus, E.,
564 Umgiesser, G., Morkunas, J., de Santi, F., Bartoli, M., 2021. Hot moments and hotspots of
565 cyanobacteria hyperblooms in the Curonian Lagoon (SE Baltic Sea) revealed via remote sensing-
566 based retrospective analysis. *Sci. Total Environ.* 769, 145053.

567 Weber, S.J., Mishra, D.R., Wilde, S.B., Kramer, E., 2020. Risks for cyanobacterial harmful algal blooms
568 due to land management and climate interactions. *Sci. Total Environ.* 703, 134608.

569 Webster, P.J., Holland, G.J., Gurry, J.A., Chang, H.R., 2005. Changes in tropical cyclone number,
570 duration, and intensity in a warming environment. *Science* 309, 1845–1846.

571 Whitehead, P.G., Wilby, R.L., Battarbee, R.W., Kernan, M., Wade, A.J., 2009. A review of the potential
572 impacts of climate change on surface water quality. *Hydrolog. Sci. J.* 54(1), 101–123.

573 Woolway, R., Weyhenmeyer, G.A., Schmid, M., Dokulil M.T., Eyto, E., Maberly, S.C., May, L.,
574 Merchant, C.J., 2019. Substantial increase in minimum lake surface temperatures under climate
575 change. *Climatic Change* 155, 81–94.

576 Xie, L.Q., Xie, P., Tang, H.J., 2003. Enhancement of dissolved phosphorus release from sediment to lake
577 water by *Microcystis* blooms — an enclosure experiment in a hyper-eutrophic, subtropical Chinese
578 lake. *Environ. Pollut.* 122, 391–399.

579 Yang, J.R., Lv, H., Isabwe, A., Liu, L.M., Yu, X.Q., Chen, H.H., Yang, J., 2017. Disturbance-induced
580 phytoplankton regime shifts and recovery of cyanobacteria dominance in two subtropical reservoirs.
581 *Water Res.* 120, 52–63.

582 Yang, Y.G., Chen, H.H., Mamun, A.A., Ndayishimiye, J.C., Yang, J.R., Isabwe, A., Luo, A.Q., Yang, J.,
583 2022. Urbanization reduces resource use efficiency of phytoplankton community by altering the
584 environment and decreasing biodiversity. *J Environ. Sci.* 112, 140–151.

585 Zhang, M., Duan, H., Shi, X., Yu, Y., Kong, F., 2012. Contributions of meteorology to the phenology of
586 cyanobacterial blooms: implications for future climate change. *Water Res.* 46, 442–452.

587 Zhou, Y.T., Scavia, D., Michalak, A.M., 2014. Nutrient loading and meteorological conditions explain
588 interannual variability of hypoxia in Chesapeake Bay. *Limnol. Oceanogr.* 59(2), 373–384.

589 Zhu, Z.X., Gao, X.F., Peng, F., Chen, H.H., Tang, L.N., Yang, J., 2017. Relationship between water
590 quality and landscape characteristics of the Houxi River watershed in Xiamen City along a rural-
591 urban gradient. *Acta. Ecologica. Sinica* 39(6), 2021–2033.

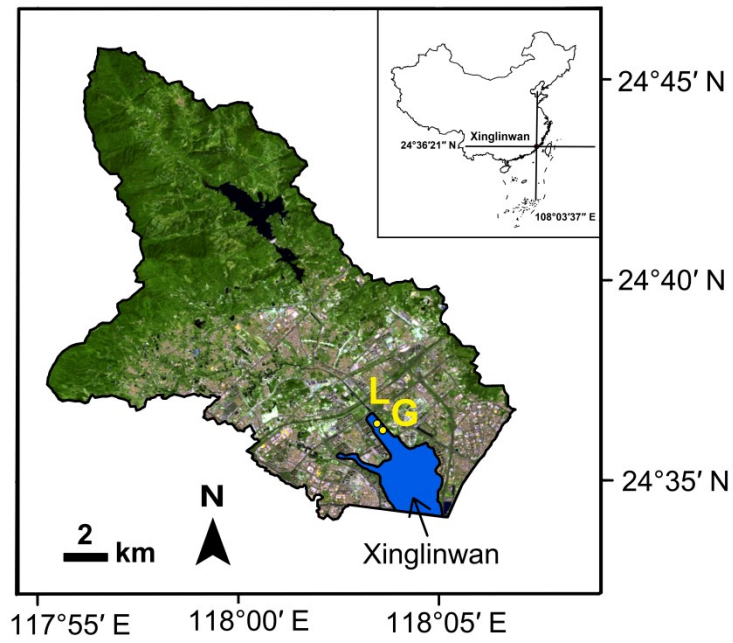


Fig. 1. Map of Houxi River watershed showing the sampling sites in Xinglinwan Reservoir, Southeast China. Stations G and L are two sampling sites.

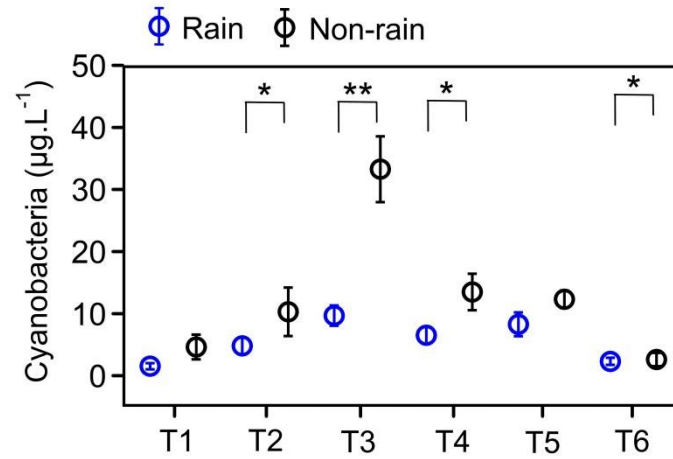


Fig. 2. Seasonal differences in cyanobacterial biomass between rain and non-rain periods from T1 to T6. T1, February and March; T2, April and May; T3, June and July; T4, August and September; T5, October and November; T6, December and January. (*, $P < 0.05$; **, $P < 0.001$)

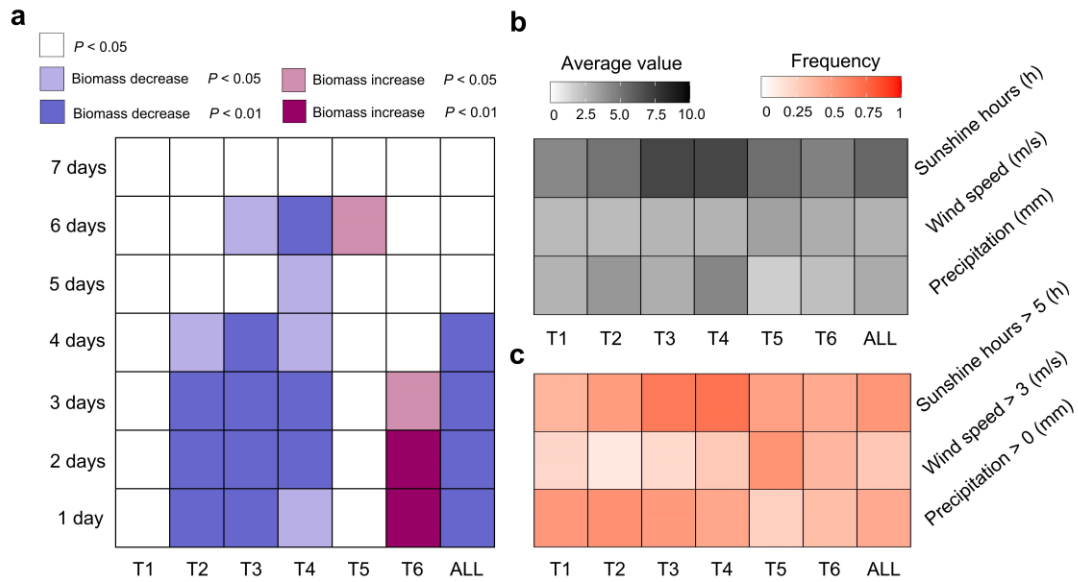


Fig. 3. The response of cyanobacteria biomass to rain events in 1-7 day/days (a), and the meteorological conditions from T1 to T6 (b, c). For (a), the colored squares represented significant differences in cyanobacterial biomass before and after rainfall events. Darkness or lightness of color represented the level of significant difference in cyanobacteria. The purple color represents a significant decrease in cyanobacterial biomass after rainfall, while the pink color represents a significant increase in cyanobacterial biomass after rainfall. X day/days ($X = 1, 2, 3, 4, 5, 6$) represented the total rainfall for X days prior to sampling. For (b), the colored squares represented the average of meteorological factors during T1-T6. (c) the occurrence frequency of conditionally meteorological factors in the period of T1-T6.

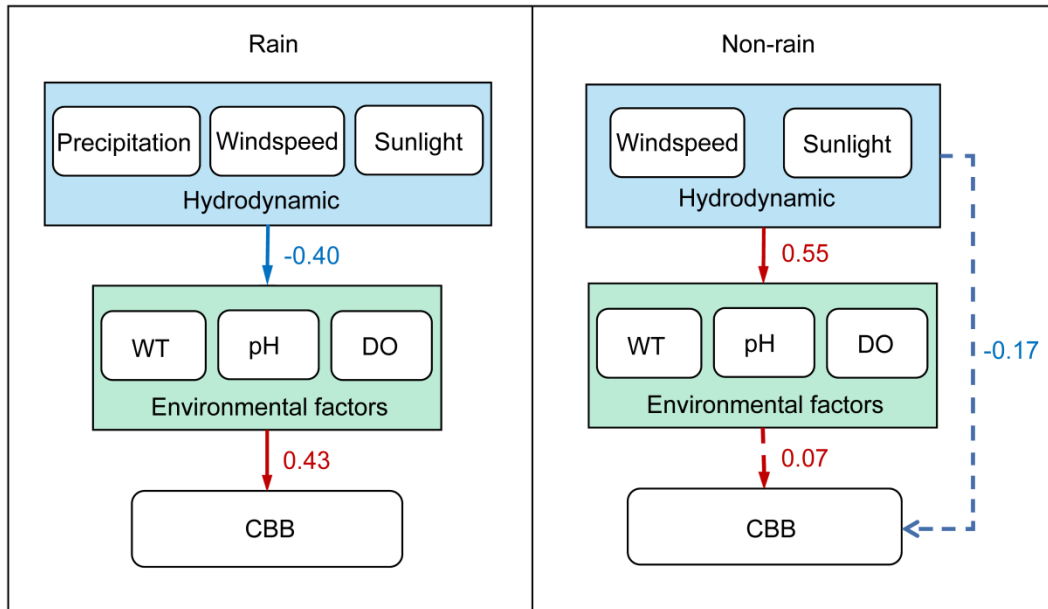


Fig. 4. Path analysis of the relationship between hydrodynamic and cyanobacterial biomass. The red and blue lines represent positive and negative correlations, respectively. The number represents the significant path coefficient. The solid lines represented the significant paths and dotted lines represented the non-significant paths. The blue rectangle represents the hydrodynamic driving factors. The green rectangle represents the environmental factors. Windspeed, the average wind speed of the three days prior to sampling; Precipitation, the total precipitation of the three days prior to sampling; Sunlight, the average sunshine hours of the three days prior to sampling; Environmental factors composed of water temperature (WT), pH, dissolved oxygen (DO).

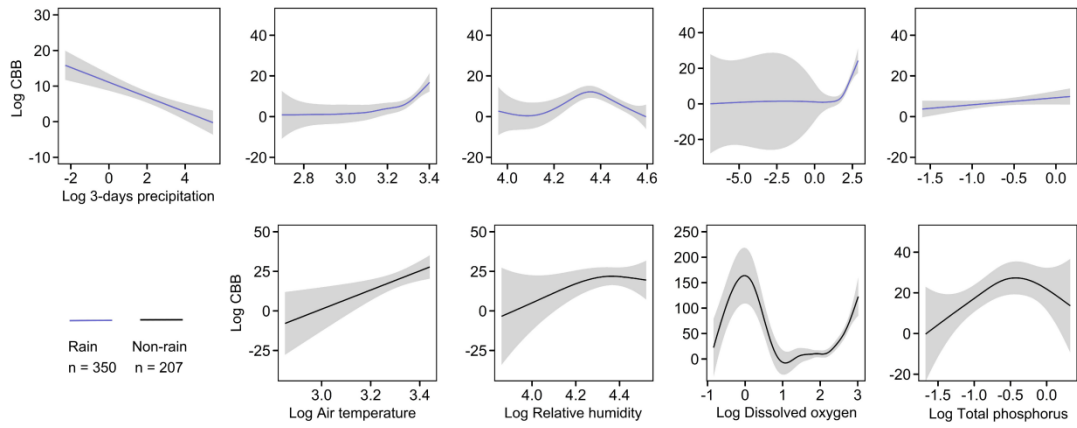


Fig. 5. Fitted plot of generalized additive models between response variables (y_1) and each of significant explanatory factors. y_1 is the difference between expected cyanobacterial biomass (CBB) derived from 95% quantile regression for warmer season (T2, T3 and T4, April to September).

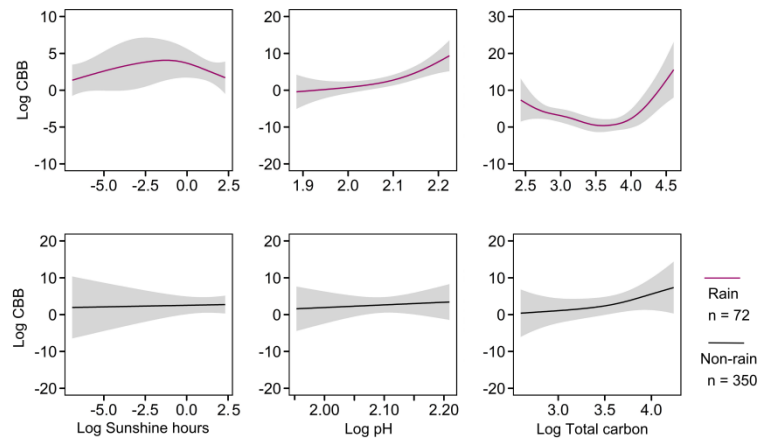


Fig. 6. Fitted plot of generalized additive models between response variables (y_2) and each of significant explanatory factors. y_2 is the difference between expected cyanobacterial biomass (CBB) derived from 95% quantile regression for colder season (T6, December and January).

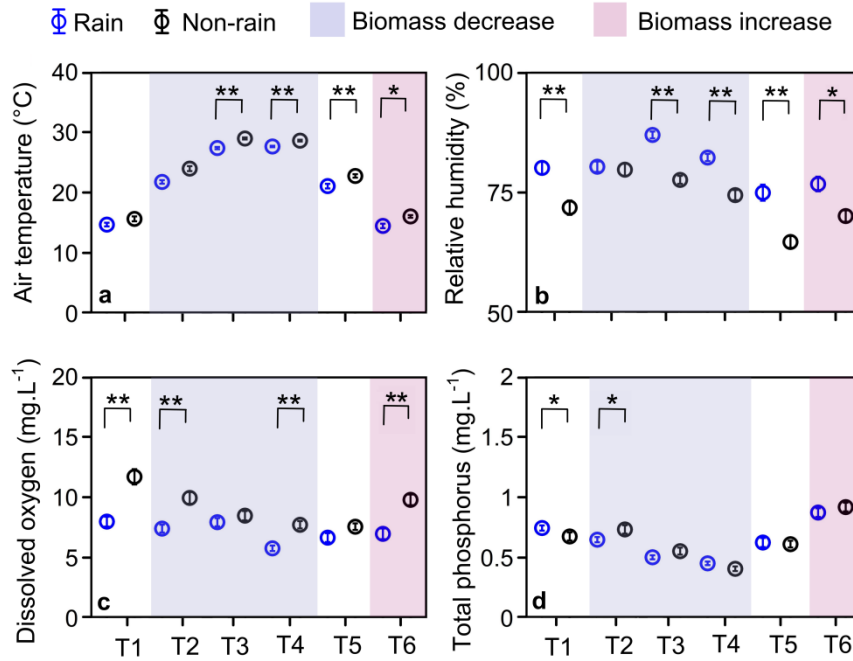


Fig. 7. The seasonal difference (*, $P < 0.05$; **, $P < 0.001$) of the explanatory factors for y_1 between rain and non-rain periods. y_1 is the difference between expected cyanobacterial biomass (CBB) derived from 95% quantile regression for warmer season.

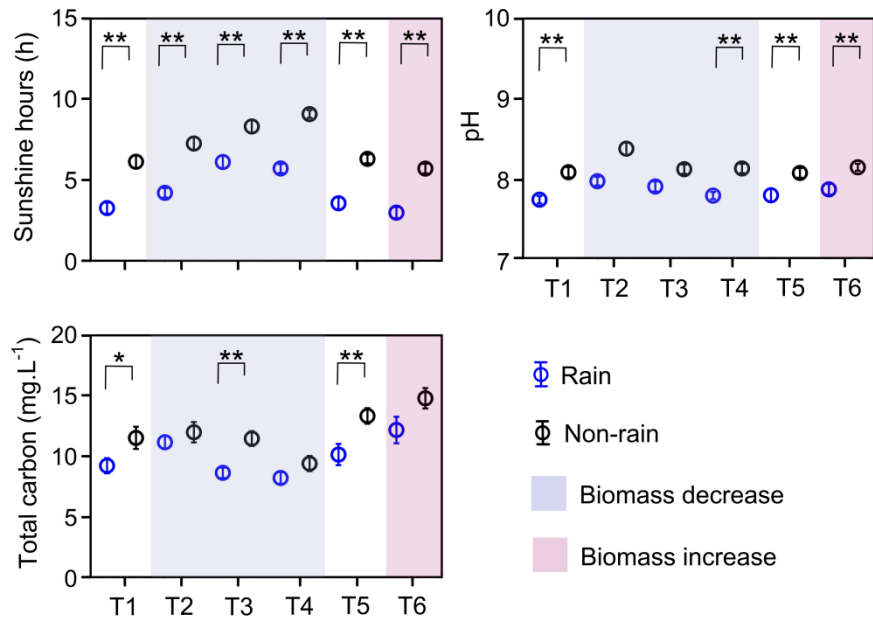


Fig. 8. The seasonal difference (*, $P < 0.05$; **, $P < 0.001$) of the explanatory factors for y_2 between rain and non-rain periods. y_2 is the difference between expected cyanobacterial biomass (CBB) derived from 95% quantile regression for colder season.

Table 1

Seasonal analysis of cyanobacterial biomass (CBB) and all explanatory factors for samplings with prior rain events (2016-2020).

Variables	T1	T2	T3	T4	T5	T6
	N = 101	N = 121	N = 100	N = 129	N = 58	N = 72
CBB ($\mu\text{g.L}^{-1}$)	1.54 \pm 4.94	4.81 \pm 14.29	9.69 \pm 18.31	6.53 \pm 15.47	8.29 \pm 14.57	2.32 \pm 4.93
Pcpt (mm)	16.07 \pm 22.14	18.02 \pm 22.66	19.49 \pm 24.96	28.60 \pm 42.63	13.90 \pm 23.63	14.60 \pm 24.91
AT ($^{\circ}\text{C}$)	14.69 \pm 3.07	21.75 \pm 3.29	27.35 \pm 1.85	27.62 \pm 1.31	21.05 \pm 2.96	14.46 \pm 3.17
RH (%)	80.08 \pm 12.73	80.33 \pm 11.98	86.91 \pm 7.11	82.23 \pm 8.57	74.95 \pm 12.64	76.75 \pm 12.86
SH (h)	3.27 \pm 3.45	4.21 \pm 3.15	6.11 \pm 3.96	5.72 \pm 3.10	3.57 \pm 2.75	2.99 \pm 2.89
pH	7.76 \pm 0.47	7.99 \pm 0.56	7.92 \pm 0.62	7.81 \pm 0.50	7.81 \pm 0.62	7.89 \pm 0.56
DO (mg.L^{-1})	7.98 \pm 4.11	7.40 \pm 3.60	7.93 \pm 3.39	5.77 \pm 2.62	6.66 \pm 2.82	6.97 \pm 3.74
TC (mg.L^{-1})	25.17 \pm 9.59	27.09 \pm 9.91	23.64 \pm 6.42	21.17 \pm 9.53	27.01 \pm 11.59	31.73 \pm 15.28
TP (mg.L^{-1})	0.75 \pm 0.22	0.65 \pm 0.24	0.50 \pm 0.17	0.45 \pm 0.17	0.63 \pm 0.29	0.87 \pm 0.31

T1 (February and March), T2 (April and May), T3 (June and July), T4 (August and September), T5 (October and November) and T6 (December and January); CBB, cyanobacterial biomass; Pcpt, the total precipitation 3-days prior to sampling; AT, air temperature; RH, relative humidity; DO, dissolved oxygen; TC, total carbon; TP, total phosphorus.

Table 2

The R^2 -adjusted and P value of significant explanatory variables in the generalized additive model for CBB for the T2 to T4 warm period (April to September).

y1	Log(Pcpt)	Log(AT)	Log(RH)	Log(DO)	Log(TP)	All
R^2 -adjusted	0.06	0.12	0.05	0.16	0.02	0.31
P	0.014	<0.001	0.007	<0.001	<0.001	<0.001

y1 is the difference between expected cyanobacterial biomass (CBB) derived from 95% quantile regression in CBB decreased periods (warm seasons) with precipitation greater than 0 in the 3 days before sampling. Pcpt, the total precipitation during the first three days of sampling; AT, daily air temperature; RH, daily relative humidity; DO, dissolved oxygen; TP, total phosphorus; All; whole y1 equation.

Table 3

The R²-adjusted and P values of significant explanatory variables in the generalized additive model of CBB for the T6 cold period (December & January).

y2	Log(SH)	Log(pH)	Log(TC)	All
R ² -adjusted	0.008	0.009	0.18	0.43
<i>P</i>	<0.001	0.003	0.003	<0.001

y2 is the difference between expected cyanobacterial biomass (CBB) derived from 95% quantile regression in CBB increased periods (cold seasons) with precipitation greater than 0 in the 3 days before sampling. SH, daily sunshine hours; TC, total carbon; All, whole y2 equation.

Journal: Water Research

Supplementary information of the article

Short-term responses of cyanobacteria to the environment after rainfall events in a subtropical urban reservoir

Anqi Luo^{1,2}, Huihuang Chen^{1,2}, Xiaofei Gao^{1,3}, Laurence Carvalho⁴, Jun Yang^{1,*}

¹ *Aquatic EcoHealth Group, Fujian Key Laboratory of Watershed Ecology, Key Laboratory of Urban Environment and Health, Institute of Urban Environment, Chinese Academy of Sciences, Xiamen 361021, China*

² *University of Chinese Academy of Sciences, Beijing 100049, China*

³ *College of Fisheries, Henan Normal University, Xinxiang 453007, China*

⁴ *Centre for Ecology & Hydrology, Penicuik, UK*

***Corresponding author:**

Jun Yang, E-mail address: jyang@iue.ac.cn

This supplementary information contains:

- 5 Pages
- 1 Figure
- 3 Tables

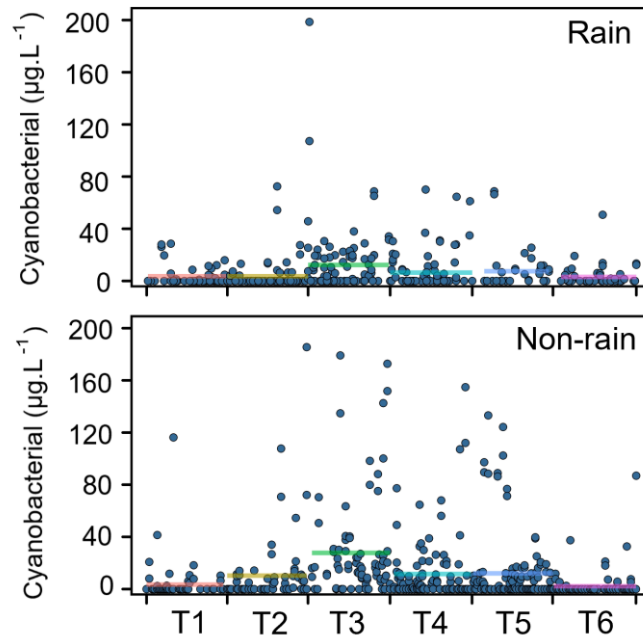


Fig. S1. Scatter plot of cyanobacteria biomass in periods of rain or non-rain from T1 to T6 (2016-2020). The colored horizontal lines represent the average value of cyanobacterial biomass in each Tgroup (T1 to T6).

Table S1

Factor load matrix of warmer season (T2, T3 and T4). The bold black number represent the factors with the highest absolute value score of each Factor (Factor1-Factor6).

Variables	Factor1	Factor2	Factor3	Factor4	Factor5	Factor6
Water temperature (°C)	-0.08	0.929	0.147	-0.093	-0.001	-0.024
pH	-0.05	0.103	0.827	-0.124	0.074	-0.177
DO (mg.L ⁻¹)	0.004	0.085	0.858	0.005	0.029	-0.049
Turbidity	0.038	0.048	-0.069	-0.123	-0.267	0.602
EC (µS/cm)	0.858	-0.131	0.167	0.243	-0.031	0.023
Salinity	0.85	-0.129	0.169	0.247	-0.036	0.027
ORP (mV)	-0.115	-0.165	-0.159	-0.002	-0.739	0.041
TC (mg.L ⁻¹)	0.689	0.071	-0.003	0.204	0.512	-0.210
TOC (mg.L ⁻¹)	0.601	0.050	0.044	0.209	0.607	-0.194
TN (mg.L ⁻¹)	0.688	-0.212	-0.061	-0.035	0.459	-0.247
NH ₄ -N (mg.L ⁻¹)	0.672	-0.076	-0.242	-0.039	-0.014	-0.438
NO ₃ -N (mg.L ⁻¹)	-0.644	-0.306	-0.137	0.027	0.226	0.395
NO ₂ -N (mg.L ⁻¹)	-0.164	-0.104	0.099	-0.147	0.555	-0.402
TP (mg.L ⁻¹)	0.871	-0.148	0.015	-0.059	0.144	0.071
PO ₄ -P (mg.L ⁻¹)	0.745	-0.167	-0.220	-0.096	0.019	0.155
Large evaporation (mm)	0.329	0.425	0.168	0.709	-0.032	-0.122
Relative humidity (%)	-0.231	0.058	-0.219	-0.795	-0.102	0.220
Sunshine hours (h)	-0.022	0.569	0.518	0.399	0.123	-0.023
Air temperature (°C)	0.246	0.939	-0.009	-0.004	0.020	0.046
Surface temperature (°C)	-0.154	0.925	0.204	0.068	0.078	-0.066
Wind speed (m/s)	-0.147	-0.111	-0.293	0.660	-0.072	0.007
AT _{max} -AT _{min} (°C)	0.091	0.270	0.557	0.326	0.230	-0.061
Precipitation (mm)	-0.163	-0.119	-0.176	-0.138	-0.066	0.674

DO, dissolved oxygen; EC, electrical conductivity; ORP , oxidation-reduction potentia; TC, total carbon; TOC, total oxygen carbon; TN, total nitrogen; NH₄-N, ammonium nitrogen; NO₃-N, nitrate nitrogen ; NO₂-N, nitrite nitrogen; TP, total phosphorus; PO₄-P, phosphate phosphorus; AT_{max}-AT_{min}, value of the maximum temperature minus minimum temperature; Precipitation, The total rainfall for the first three days of the sampling date.

Table S2

Factor load matrix of colder season (T6). The bold black number represent the factors with the highest absolute value score of each Factor (Factor1-Factor8).

DO, dissolved oxygen; EC, electrical conductivity; ORP , oxidation-reduction potentia; TC, total carbon; TOC, total oxygen carbon; TN, total nitrogen; NH₄-N, ammonium nitrogen; NO₃-N, nitrate nitrogen ; NO₂-N, nitrite nitrogen; TP, total phosphorus; PO₄-P, phosphate phosphorus; AT_{max}-AT_{min}, value of the maximum temperature minus minimum temperature; Precipitation, The total rainfall for the first three days of the sampling date.

Variables	Factor1	Factor2	Factor3	Factor4	Factor5	Factor6	Factor7	Factor8
Water temperature (°C)	0.067	0.084	0.903	0.050	0.121	-0.050	0.112	0.115
pH	-0.078	0.231	0.091	-0.249	-0.094	-0.504	-0.199	0.614
DO (mg.L ⁻¹)	0.430	0.420	-0.092	-0.164	0.138	-0.167	-0.162	0.497
Turbidity	0.299	-0.127	0.116	-0.702	-0.219	0.438	-0.089	-0.050
EC (μS/cm)	0.376	0.053	0.025	0.041	0.880	-0.163	0.135	-0.084
Salinity	0.367	0.054	0.022	0.036	0.886	-0.161	0.129	-0.083
ORP (mV)	-0.094	-0.151	-0.139	-0.116	-0.051	-0.132	0.052	-0.813
TC (mg.L ⁻¹)	0.883	-0.077	0.061	-0.058	0.166	-0.115	0.251	0.126
TOC (mg.L ⁻¹)	0.882	0.001	0.051	0.038	0.255	-0.035	0.177	0.238
TN (mg.L ⁻¹)	0.459	0.026	-0.013	0.738	-0.120	0.202	-0.015	0.145
NH ₄ -N (mg.L ⁻¹)	0.338	-0.003	0.045	0.405	0.254	-0.253	-0.027	-0.526
NO ₃ -N (mg.L ⁻¹)	-0.137	0.096	-0.171	-0.011	-0.350	0.846	-0.099	0.147
NO ₂ -N (mg.L ⁻¹)	0.240	0.100	0.189	0.214	-0.178	0.075	0.113	0.727
TP (mg.L ⁻¹)	0.882	0.115	-0.021	0.043	0.275	-0.092	-0.155	-0.071
PO ₄ -P (mg.L ⁻¹)	0.690	0.044	-0.136	0.363	0.239	0.034	-0.295	-0.374
Large evaporation (mm)	0.133	0.687	-0.074	-0.200	0.203	-0.075	0.532	0.093
Relative humidity (%)	-0.105	-0.813	0.196	0.046	-0.215	0.055	-0.360	-0.063
Sunshine hours (h)	-0.053	0.912	0.096	0.012	-0.062	-0.026	-0.027	0.176
Air temperature (°C)	-0.054	-0.137	0.931	-0.026	-0.071	-0.105	-0.164	0.075
Surface temperature (°C)	0.016	0.005	0.935	-0.096	-0.014	-0.129	-0.153	0.088
Wind speed (m/s)	0.057	0.044	-0.159	0.065	0.148	-0.173	0.881	-0.074
AT _{max} -AT _{min} (°C)	-0.019	0.803	0.037	0.254	-0.084	-0.019	-0.202	0.087
Precipitation (mm)	-0.088	-0.141	-0.143	-0.073	-0.036	0.895	-0.153	0.045

Table S3

The GAM fitting between each selected factor with cyanobacteria, from y1 (warmer season) or y2 (colder season) respectively. The bold black number represent the factors with the $P < 0.05$ of each Factor (Factor1-Factor8).

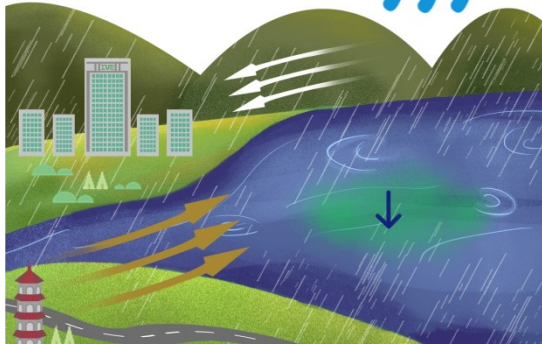
TP, total phosphorus; DO, dissolved oxygen; ORP , oxidation-reduction potentia; Precipitation, The total rainfall for the first three days of the sampling date; TC, total carbon; TN, total nitrogen.

Variables	y1		y2	
	Explained	<i>P</i>	Explained	<i>P</i>
Air temperature (°C)	13.20 %	< 0.001		
Relative humidity (%)	6.29 %	0.002		
TP (mg.L ⁻¹)	2.77 %	0.050		
DO (mg.L ⁻¹)	16.5 %	< 0.001		
ORP (mV)	2.73 %	0.020		
Precipitation (mm)	6.52 %	< 0.001	4.05 %	0.342
Sunshine hours (h)			11.2 %	0.090
Surface temperature (°C)			4.15 %	0.392
Wind speed (m/s)			20.1%	0.050
Salinity			15.4 %	0.102
pH			9.89 %	0.007
TC (mg.L ⁻¹)			21.9 %	0.004
TN (mg.L ⁻¹)			6.57 %	0.297

GRAPHICAL ABSTRACT

Rain

Wind, large runoff, flushing and dilution reduce cyanobacteria



Non-rain

Small runoff and sunlight with more cyanobacteria



Short-term responses of cyanobacteria to the environment after rainfall events in a subtropical urban reservoir

Keywords: climate change; harmful algal bloom; hydrodynamics regimes; flushing; dilution; subtropical reservoir

Highlight:

1. The short-term (3 days) effects of rainfall events on cyanobacterial biomass and hydrodynamic conditions were analyzed using the long-term (5 years), high-frequency (half a week) sampling data from Xinglinwan Reservoir.
2. Rainfall events reduced cyanobacterial biomass within 3 days by changing the magnitude and frequency of rainfall (and sunshine) in both warm season and cold seasons.
3. Cyanobacterial biomass decreased in warmer season in response to precipitation, air temperature, relative humidity, dissolved oxygen and total phosphorus compared with decreased in the colder season in response to sunshine hours, pH, total carbon.
4. Hydrodynamic factors that rainfall, wind speed and sunshine hours are the driving factors for the short-term reduction of cyanobacterial biomass.
5. The explanatory factors for cyanobacterial biomass in warmer season and colder season were examined using factors analysis combined with generalized additive models. And used the least-square path analysis to explore the relationship between cyanobacterial biomass and hydrodynamic regimes.

Hypothesis:

1. The short-term response of cyanobacterial biomass varies seasonally in relation to the magnitude and frequency of rainfall (and sunshine) events.

2. Short-term rainfall events indirectly affected cyanobacteria biomass by changing different environmental factors in warmer season and colder season.

3. Hydrodynamic regimes are the main factors that affects the relationship between cyanobacterial biomass and rainfall in the short term.